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1 Original article

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3 Rapid local adaptation in both sexual and asexual invasive populations of monkeyflowers

- 4 (Mimulus spp.)
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- 13 **Running tittle:** Adaptation in sexual and asexual Mimulus

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19 Abstract

20 • Background and Aims

Traditionally, local adaptation has been seen as the outcome of a long evolutionary history,
particularly in sexual lineages. In contrast, phenotypic plasticity has been thought to be most
important during the initial stages of population establishment and in asexual species. We
evaluated the roles of adaptive evolution and phenotypic plasticity in the invasive success of two
closely related species of invasive monkeyflowers (*Mimulus*) in the United Kingdom (UK) that
have contrasting reproductive strategies: *M. guttatus* combines sexual (seeds) and asexual (clonal
growth) reproduction while *M. × robertsii* is entirely asexual.

28 • Methods

29 We compared the clonality (number of stolons), floral and vegetative phenotype, and phenotypic plasticity of native (M. guttatus) and invasive (M. guttatus and M. \times robertsii) populations grown 30 31 in controlled environment chambers under the environmental conditions at each latitudinal 32 extreme of the UK. The goal was to discern the roles of temperature and photoperiod on the expression of phenotypic traits. Next, we tested the existence of local adaptation in the two 33 species within the invasive range with a reciprocal transplant experiment at two field sites in the 34 latitudinal extremes of the UK, and analysed which phenotypic traits underlie potential local 35 fitness advantage in each species. 36

37 • Key Results

Populations of *M. guttatus* in the UK showed local adaptation through sexual function (fruit
production), while *M. x robertsii* showed local adaptation via asexual function (stolon
production). Phenotypic selection analyses revealed that different traits are associated with
fitness in each species. Invasive and native populations of *M. guttatus* had similar phenotypic

42	plasticity and clonality. M . × robertsii presents greater plasticity and clonality than native M .
43	guttatus, but most populations have restricted clonality under the warm conditions of the south of
44	UK.
45	• Conclusions
46	Our study provides experimental evidence of local adaptation in a strictly asexual invasive
47	species with high clonality and phenotypic plasticity. This indicates that even asexual taxa can
48	rapidly (< 200 years) adapt to novel environmental conditions in which alternative strategies may
49	not ensure the persistence of populations.
50	
51	Keywords: Asexual, introduced species, local adaptation, <i>Mimulus guttatus</i> , <i>M</i> . × <i>robertsii</i> , <i>M</i> .

52 *luteus*, phenotypic plasticity, reciprocal transplants.

1 Introduction

2 Populations of broadly distributed species adapt to local conditions through genetic

3 differentiation (Williams, 1966; Kawecki and Ebert, 2004; Hereford, 2009) and phenotypic

4 plasticity (Bradshaw, 1965; Donohue, 2013). These two mechanisms are universal, interacting,

5 and non-mutually exclusive (Price *et al.*, 2003; de Jong, 2005; West-Eberhard, 2005; Kelly,

6 2019). Yet, the traditional view was that local adaptation has a greater importance in sexual

7 populations with a long evolutionary history (i.e. those with a greater number of recombination

8 events behind; Weissmann, 1889; Crow and Kimura, 1965; Maynard Smith, 1968; Burt, 2000;

9 Rushworth et al., 2020). In contrast, clonal propagation has been considered to reduce the

10 opportunities for local adaptation (Schon *et al.*, 1998; Rouzine *et al.*, 2003; Schiffels *et al.*, 2011)

11 despite this mechanism can theoretically occur through selection on genes or genotypes

12 (Vrijenhoek, 1979; Lushai *et al.*, 2003). Given the expected reduction in genotypic diversity,

13 phenotypic plasticity has been attributed a more important role in asexual lineages (Lynch, 1984;

14 Van Kleunen and Fischer, 2001; Oplaat and Verhoeven, 2015; Fazlioglu and Bonser, 2016; Geng

et al., 2016) and during the initial stages of population establishment (Davidson *et al.*, 2011;

16 Liao et al., 2016).

Introduced species often evolve to cope with novel biotic and abiotic conditions in nonnative ranges (Bossdorf *et al.*, 2005; Vandepitte *et al.*, 2014; Oduor *et al.*, 2016; Mitchell and
Whitney, 2018; Liu *et al.*, 2020), and thus constitute an excellent model system to study adaptive
evolution occurring over short periods of time (Thompson, 1998; Colautti and Lau, 2015). In
addition, phenotypic plasticity seems to make a major contribution to the establishment and
spread of introduced species in novel environments (Ghalambor *et al.*, 2007; Riis *et al.*, 2010;
Ebeling *et al.*, 2011; Pahl *et al.*, 2013; Liao *et al.*, 2016; Liu *et al.*, 2016). Interestingly, clonal

1 propagation is an advantageous trait for plant invasions, and numerous invasive plant species combine both sexual and asexual modes of reproduction or are mostly asexual (Pyšek, 1997; 2 Silvertown, 2008; Roiloa, 2019). However, although an increasing number of studies have 3 shown evolution at a contemporary scale in invasive plants with sexual (Lucek et al., 2004; 4 Maron et al., 2004; Leger and Rice, 2007; Novy et al., 2013; Li et al., 2015; Bhattarai et al., 5 6 2017; Marchini et al., 2018) or mixed reproductive systems (Michel et al., 2004; Lambertini et al., 2010), field tests of local adaptation and phenotypic plasticity are rare for obligately asexual 7 flowering plants (Lovell et al., 2014; Rushworth et al. 2020). 8

In this study, we investigate the evolutionary strategies of two invasive Minulus 9 10 (Phrymaceae) species that differ in their ability to reproduce sexually: *Mimulus guttatus* DC. (which combines sexual and asexual reproduction) and M. × robertsii Silverside (strictly 11 asexual). We evaluate the roles of adaptive evolution and phenotypic plasticity in the invasive 12 13 success of *Mimulus* at two nested levels: (i) between native and introduced populations, and (ii) among introduced populations. In a first experiment, we assess phenotypic differences and 14 compared the clonality and plasticity of ancestral native (M. guttatus) and invasive (M. guttatus 15 and M. × robertsii) populations under the environmental conditions at each latitudinal extreme of 16 the UK, discerning the roles of temperature and photoperiod in a full-crossed design 17 implemented in controlled environment chambers. Our hypothesis here is that as $M. \times$ 18 19 *robertsii* should display levels of clonality and plasticity equal or higher than the sexual taxa (native *M. guttatus* and invasive *M. guttatus*). In a second experiment, we test the existence of 20 21 local adaptation of the two species within the invasive range with a reciprocal transplant 22 experiment at two field sites in the latitudinal ends of UK and analyse which phenotypic traits underlie the local fitness advantage in each species. We predict that if sexual reproduction boosts 23

1 adaptation, *M. guttatus* would be more likely to be locally adapted than $M. \times robertsii$. To

2 explore the possible mechanisms driving local adaptation, we also carry out phenotypic selection

3 analyses to identify the phenotypic traits related to fitness in each species.

4 Materials and Methods

5 *Study system*

Mimulus guttatus (2n = 2x = 28) is an herbaceous, annual or perennial, plant native to Western 6 7 North America (Grant, 1924; Wu et al., 2007; Lowry and Willis, 2010). M. guttatus was 8 introduced in the United Kingdom (UK) for ornamental purposes 200 years ago (Roberts, 1964; 9 Parker, 1975; Puzey and Vallejo-Marín, 2014) and perennial forms, which combine reproduction via seeds (sexual) and stolons (asexual), became naturalised in wetlands, riverbanks and wet 10 ditches across the entire country (Preston et al., 2002), as in other areas in Europe and New 11 12 Zealand (Howell and Sawyer, 2006; Truscott et al., 2006; Da Re et al., 2020). The second taxon, *Mimulus* \times *robertsii*, is a triploid (2n = 3x = 44-46) originated in the UK, product of an unknown 13 14 number of hybridisation events between introduced populations of the diploid *M. guttatus* and 15 the closely related South American tetraploid *M. luteus* L. (2n = 4x = 60-62). *M. luteus* was 16 introduced in the UK soon after *M. guttatus* but is currently rare (Vallejo-Marín and Lye, 2013). The hybrid $M \times robertsii$, which is perennial and sexually sterile (Parker, 1975; Meeus *et al.*, 17 2020) but capable of extensive clonal reproduction via stolons, has become well established 18 19 across UK, though it is far less abundant than *M. guttatus* in the south range of the country 20 (Preston et al., 2002; Stace, 2010; Vallejo-Marín and Lye, 2013; Da Re et al. 2020). M. guttatus and M. × robertsii are very similar in their morphology, phenology and habitat in the UK. Both 21 22 species bear high genetic diversity and low genetic structure (Vallejo-Marín and Lye, 2013;

Pantoja *et al.*, 2017), suggesting that metapopulation dynamics with high gene flow are
 important in the spatial structuring of the introduced range.

3 Experiment 1: controlled environment chambers

4 Plant material

For *M. guttatus*, we used seeds from five native populations from North America and from five 5 6 introduced populations in the UK [Supplementary Information - Table S1]. We follow Lowry et 7 al., (2019) and use the classic taxonomical definition of M. guttatus DC. (Grant, 1924), rather than the recent nomenclature proposed by Nesom (2014). All seeds were field-collected, except 8 9 seeds from accessions in the Alaskan range, the putative ancestral range of UK populations (Puzev and Vallejo-Marín, 2014; Vallejo-Marin et al., 2020). Three Alaskan accessions were 10 11 retrieved from herbarium specimens preserved at University of Alaska Museum Herbarium (accessions V153408, V127607, V142998). As each accession represents a single sampled 12 13 individual and locality, these three accessions were pooled into a single Alaskan "population" 14 (ALA). From each population, we selected three to five maternal seed families (seeds collected from the same maternal parent). In total, we had 43 families from 10 populations. For the 15 sexually sterile hybrid M. \times robertsii, we collected in the field vegetative fragments (clones) 16 from five UK populations [Supplementary Information - Table S1]. In each population, we 17 18 sampled 1-5 ramets (limited by population size) separated at least 1m to reduce the probability of sampling the same genet multiple times (15 ramets total from five populations). All maternal 19 families in both species were randomly collected with regards of their clonality and phenotypic 20 traits. Native and invasive populations of *M. guttatus* cover a wide latitudinal range of their 21 22 distribution, while M. \times robertsii populations proceeded from the centre and north of its narrower 23 range.

1 *Experimental treatments*

2 We used the Controlled Environment Facility at the University of Stirling to create environmental conditions that resembled the UK *Mimulus* growing season (Fig. 1a). To model 3 the conditions, we used two opposite localities that encompass the latitudinal range of *Mimulus* 4 in the UK: Newport, in the Isle of Wight (50.70° N, 1.29° W), and Baltasound, in the Shetland 5 6 Isles (60.76° N, 0.86° W). For each of these localities and for every two-week period between April and September (the UK *Mimulus* growing season), we calculated the photoperiod with the 7 package geosphere (Hijmans, 2014) on R (R Core Team, 2019) and obtained maximum and 8 9 minimum temperatures from the WorldClim database (Hijmans et al., 2005). Photoperiod and temperature temporal series were combined in a full-crossed design to create four experimental 10 treatments that allowed us to disentangle the effects of temperature and photoperiod on plant 11 performance: a short day, warm temperature treatment (SW; the natural conditions in Newport), 12 a long day, cold temperature treatment (LC; the natural conditions in Baltasound), a short day, 13 cool temperature treatment (SC) and a long day, warm temperature one (LW) [Supplementary 14 Information - Fig. S1 and Table S2]. The different growth conditions in each temporal series 15 were substituted every 10 days to allow completing the experiment in 120 days. Each 16 17 experimental treatment was implemented in one Snijder Scientific (Tilburg, Netherlands) MC1750E controlled environmental chamber. 18

19 *Plant growth*

We planted seeds from each of 43 maternal families of *M. guttatus* into four 0.5 L pots (172 pots in total) filled with modular seed growing medium (Sinclair, Lincoln, Lincolnshire, UK), and placed them in the dark at 4°C for one week. For *M.* × *robertsii*, we planted individually eight cuttings from each of 15 maternal families in 0.5 L pots (120 pots in total) filled with All-

1 Purpose growing medium (Sinclair, Lincoln, Lincolnshire, UK). All cuttings had a similar size, two small leaves and ~ 2 cm. of roots. We moved one pot per maternal family of *M. guttatus* and 2 two pots per maternal family of $M_{\star} \times robertsii$ to each chamber on 1st May 2014. We noted the 3 day of first germination for each M. guttatus pot and, four weeks after first germination, we 4 selected and thinned the two biggest seedlings to one per pot filled with All-Purpose growing 5 6 medium in order to get two replicates per maternal family in each chamber. The maximum difference in transplant time among pots was one week within each chamber and two weeks 7 across the entire experiment. Pots were randomly repositioned within each chamber every other 8 9 day throughout the experiment.

10 Measurements and statistical analyses

11 Most individuals survived until the end of the experiment, and all measurements were taken at 12 this moment except otherwise specified. To investigate the phenotypic differences in clonality among the three population types (classified by their origin and species, i.e., native M. guttatus, 13 invasive M. guttatus, and $M. \times robertsii$), we recorded the total number of stolons produced by 14 individuals in the four environmental chambers. To compare their phenotypes, we recorded days 15 16 to flower since germination or planting of the clonal fragment, corolla width of the first flower (measured with a digital calliper to the nearest 0.1mm in the second day after anthesis), whether 17 18 plants flowered or not, the number of branches, floral stems and flowers, plant height (from the 19 soil surface to the highest meristem, measured to the nearest cm), and length and diameter of the first internode. Finally, the entire individuals (above- and belowground) were harvested, washed 20 out gently in water and dried at 60° C in individual paper bags for estimating final total dry 21 22 biomass. Despite being sterile, we consider M. x robertsii flowering as indicative of individual performance. In order to avoid over-parameterization in subsequent analyses, we averaged the 23

two values from each family (cuttings in *M. × robertsii*, siblings in *M. guttatus*), for each trait
under each of four treatments (except for germination time in *M. guttatus*, which had a single
data point per family).

Preliminary analyses showed low correlation between most phenotypic traits 4 [Supplementary Information - Fig. S2] and thus each variable was analysed separately. We used 5 6 Generalized Linear Mixed Models (GLMMs) to analyse the variation in clonal reproduction and in phenotypic traits as a function of the population type (native *M. guttatus*, introduced *M*. 7 8 guttatus and M. × robertsii), treatment photoperiod (Short vs. Long), treatment temperature (Warm vs. Cold), and their two- and three-way interaction terms. Analogous GLMMs were also 9 10 carried out for each population type separately. In all models, population was included as a random effect. We used a Poisson error distribution for number of stolons, branches, floral 11 stems, and flowers, a binomial model for flowering, and a Gaussian model for germination time, 12 13 flowering time, corolla width, plant height, internode length, internode diameter, and dry mass. The survival of plants was above 96% and thus this variable was not modelled. The significance 14 of the fixed effects and their interactions were assessed by type-III Wald $\gamma 2$ tests on the 15 corresponding GLMMs. Where the interactions were not significant, we removed them and 16 tested also the effect of the main effects alone with type-II Wald χ^2 tests. To account for multiple 17 tests, we applied a Bonferroni correction, dividing the significance alpha level by the number of 18 variables analysed (corrected P-value = 0.004). Where population type or any interaction of 19 fixed factors were significant, we performed post hoc contrasts based on estimated marginal 20 21 means (EMMs) of the corresponding model. These procedures were repeated for all GLMMs in 22 this study. All analyses were performed in R 3.4.0 (R Core Team, 2019) with packages lme4 (Bates et al., 2015), car (Fox et al., 2012) and emmeans (Lenth et al., 2018). 23

1 To investigate differences in phenotypic plasticity among population types, we estimated the Relative Distances Plasticity Index (RDPI; Valladares et al., 2006) for each trait measured in 2 the chambers, for each family. RDPI were first estimated from trait distances between the two 3 temperature (RDPI_{*l*}) and the two photoperiod (RDPI_{*p*}) treatments separately, pooling data from 4 two chambers in each treatment. Trait distances were calculated as the absolute value of the 5 6 difference of trait values of the same family (the average of the two individual replicates) at each of two treatments, divided by the maximum of the two trait values. RDPI were also estimated for 7 each family across the four environmental chambers $(RDPI_{tp})$ as the average of the six trait 8 9 distances between each pair of chambers. We analysed the variation in phenotypic plasticity with GLMMs modelling RDPI_t, RDPI_p, and RDPI_{tp} estimates as a function of the population type. In 10 addition, we run multivariate analyses of variance (MANOVAs) with RDPI_t, RDPI_p, or RDPI_{tp} 11 estimates for all phenotypic traits as response variables and population type as independent 12 variable. RDPI estimates for germination day were excluded for multivariate analyses because 13 14 the lack of data for M. × robertsii. Finally, to test for differences in plasticity in response to temperature and photoperiod, we pooled $RDPI_t$ and $RDPI_p$ estimates and used a GLMM to 15 model RDPI values as a function of the RDPI type, population type and their interaction. All 16 17 RDPI GLMMs used a Gaussian distribution of errors.

18 Experiment 2: Reciprocal transplants

19 *Population survey and plant material*

20 We used the distribution database of the Botanical Society of Britain and Ireland (BSBI;

- 21 http://bsbidb.org.uk/) and personal records to design a survey of the northerb and the southern
- extremes of the distribution of *Mimulus guttatus* and M. × *robertsii* in the UK in summer 2014.
- 23 We focused on BSBI records from the year 2000 with a precision of at least 100 m^2 . In total, we

1 visited 60 localities between 50.1132° and 51.1489° N for the south of the country and 57.4963° 2 and 60.8087° N for the north and found 39 populations. Because we were interested in identifying potential ecotypes adapted to the latitudinal extremes of the UK, we prioritized 3 sampling fewer individuals in a higher number of populations instead of large numbers of 4 5 individuals in fewer populations. This strategy has shown great statistical power (Blanquart et 6 al., 2013) and was suited to our study system as many populations of *Mimulus* were small and likely contained few genets. To avoid sampling clones more than once, collected plants were at 7 8 least 1m apart from each other. Cuttings were transported and planted at the greenhouse of the 9 University of Stirling within one week after collection. In total, we sampled 155 cuttings from 36 populations for this study (Fig. 2) [Supplementary Information - Table S3]. M. guttatus and $M. \times$ 10 robertsii are morphologically very similar and sometimes difficult to distinguish (Vallejo-Marín 11 and Lye, 2013). To verify the species identity of each sampled individual we determined their 12 relative genome size with flow cytometry (see methods in Simón-Porcar et al., 2017). To allow 13 14 comparing M. \times robertsii with both parental species, the only available population of M. luteus in the UK for which we had seeds was included in this experiment (Fig. 2) [Supplementary 15 Information - Table S3]. For *M. luteus*, field-collected seeds from 25 different maternal 16 17 individuals were planted and, once germinated, one seedling was transplanted and grown until 18 adult.

We kept individual plants in 9-cm diameter pots filled with All-Purpose growing medium until next summer season. To buffer maternal resources effects, we transplanted a similar size fragment from each individual and randomized its position within the greenhouse at least four times between summer seasons. We cloned each individual four times to obtain replicates by April 2015, one month before setting up the experiment. All cuttings had similar architecture and

size, and they were weighted prior to planting to evaluate possible maternal resources effects on
subsequent measures of fitness. Clones were allowed to establish and develop roots and
belowground biomass, similarly to how they naturally persist between growing seasons, but we
restricted aerial biomass to the initial status by pruning elongating branches until the start of the
experiment.

6 *Experimental design*

7 Two replicates per individual, for a total of 360 plants, were transplanted into each of two common gardens [Supplementary Information Table S3]. We established one common garden in 8 9 the southernmost extreme of the UK at Ventnor Botanic Garden (Ventnor, Isle of Wight, England; 50.5890°, -1.2285°; IOW hereafter; Fig. 1b and 2) on May 14th 2015, and a second 10 common garden in the northernmost extreme of the UK at Da Gairdins i Sand (Sand, Shetland, 11 Scotland; 60.2112°, -1.3761°; SHE hereafter; Fig. 1c and 2) on May 18th 2015. The two common 12 13 gardens were set up to be identical, consisting of a 100m² square pond built up with a PVC pond liner (Aquatex, LBS Horticultural, UK), filled with 1cm of gravel to imitate natural conditions 14 and provide an appropriate environment for root growth. Individual clones were planted in 10L 15 pots filled with 7L of all-purpose commercial growing medium (LBS Horticultural, UK), which 16 were placed in the pond in a regular grid with individuals from different species and origins 17 completely randomized. Pots were 25 cm apart and pot walls precluded stolons to get out the pot, 18 avoiding mingle or competition. The ponds were permanently flooded at a level of 10 cm so that 19 plants were always moist as in natural habitats. The experiment was terminated at the end of the 20 growing season, after senescence of the aerial parts of all individuals on August 24th and 30th in 21 22 IOW and SHE, respectively.

1 Measurements and statistical analyses

2 To explore local adaptation, we assessed the fitness of individuals at each site recording their survival and reproductive success (number of stolons, and fruits in *M. guttatus*) at the end of the 3 experiment. We also explored phenotypic differentiation and the traits contributing to local 4 fitness differences within each species through phenotypic selection. For this aim we recorded 5 plant height and stomata density (in the 6th week of the experiment, when plants seemed to have 6 achieved their maximum vigour); whether plants had flowered or not, plant cover, total dry 7 biomass, and total number of branches, flowering stems, and flowers produced (at the end of the 8 9 experiment); and days to flower, first flowering node, and corolla width of the first flower (at flowering of each individual). Stomata density, a trait involved in the hydric balance of plants 10 and thus potential indicator of physiological variations (Raven 2002), was calculated under a 11 50X microscope from stomata imprints taken with transparent nail paint, adhesive tape and 12 microscope slides from the beam of three new unshaded leaves per individual. Plant cover was 13 14 measured over scaled overhead view photographs of each individual that were analysed with the software ImageJ (Abramoff et al., 2004). To estimate the dry biomass, the entire individuals 15 (above- and belowground) were harvested, washed out and dried at 60°C in individual paper 16 17 bags.

To ascertain the existence of local adaptation in *M. guttatus* and *M. × robertsii*, we analysed the variation in the sexual and asexual fitness measures (number of fruits and stolons) of each species with GLMMs, including Site, Origin, and their interaction as fixed factors in the models for each variable. The models used a Poisson distribution and included initial cutting weight as covariate and population and individual nested within population as random factors. We consider a pattern of fitness advantage at home sites jointly with a significant effect of the

interaction Site x Origin as evidence of local adaptation. The survival of plants was nearly 100%
 (see Results) and thus this variable was not modelled.

To explore the phenotypic differentiation between possible latitudinal ecotypes and 3 compare the natural environmental effects on the growth of plants with the effects found in the 4 5 environmental chambers experiment, we carried similar GLMMs for each species and 6 phenotypic trait measured. A Poisson model was used to analyse the number of branches, floral stems, and flowers, and a Gaussian model was used for the remaining variables. We consider a 7 significant effect of Origin as evidence of genetic differentiation, and a significant effect of Site 8 as evidence of strong environmental effects (i.e. plasticity) on the development and growth of 9 10 plants.

11 To investigate the phenotypic traits contributing to local fitness, we carried out phenotypic selection analyses by regressing the sexual and asexual fitness measures (number of 12 13 fruits and stolons produced) on standardized phenotypic traits separately for each species and 14 site. Only selection gradients were estimated to determine the magnitude and sign of directional and stabilizing selection on each trait, excluding indirect selection on correlated traits (Lande and 15 Arnold, 1983). Separately for each species and site, we calculated the relative fitness (individual 16 fitness divided by mean fitness) and standardized trait values (with a mean of 0 and a variance of 17 18 1). To improve the normality of the residuals in the regression models, the relative numbers of fruits and stolons were root squared. Since preliminary analyses had showed correlation between 19 various phenotypic traits for each species in this experiment [Supplementary Information - Fig. 20 21 S3], we calculated the variance inflation factors (VIF) in each model and excluded those traits 22 with VIF > 5 (i.e. number of branches and number of floral stems in models for SHE). Quadratic

regression coefficients were doubled to estimate the stabilizing/disruptive selection differentials
 (Stinchcombe *et al.*, 2008).

3 To investigate the causes of the low occurrence of *M. luteus* in the UK and compare the patterns found in the hybrid $M \times robertsii$ with both parental species, we assessed the sexual and 4 5 asexual fitness and the phenotypic patterns of the single population included in our experiment. 6 We recognise that the study of a single population does not allow robust inferences on the 7 species patterns but given the great scarcity of *M. luteus* populations in UK, we still consider this approach worthy and valuable for species comparisons. The production of fruits and stolons, and 8 9 each phenotypic trait measured, were analysed as a function of experimental site with GLMMs 10 including initial weight as covariate and individual as random factor. Then, we compared the fitness of *M. luteus* with the other two *Mimulus* species and tested the phenotypic similarity of 11 $M. \times robertsii$ and M. luteus with GLMMs including species, experimental site, and their 12 13 interaction as fixed factors, initial weight as covariate, and population and individual nested within population as random factors. Because of *M. luteus* had a single population in the north, 14 the southern populations of M. × robertsii and M. guttatus, and the variable "population origin" 15 were excluded from these analyses. Finally, we carried out phenotypic selection analyses on M. 16 *luteus* as explained above. 17

18

19 **Results**

20 Experiment 1: controlled environment chambers

21 *Clonality*

The population types differed in clonality (χ² = 17.974; P < 0.001), with M. × robertsii
 producing the most stolons, significantly more than native M. guttatus (Fig. 3). Overall, clonal
 reproduction was not affected by photoperiod but it was affected by temperature (χ² = 8.670; P = 0.003). The significant interaction of population type and temperature (χ² = 32.035; P < 0.001)

5 reflected that warm treatments increased clonality in both *M. guttatus* groups, but decreased

6 clonality in M. × robertsii (Fig. 3) [Supplementary Information - Tables S4 and S5].

7 *Phenotypes*

Overall, M. × robertsii plants were shorter, thinner, and produced fewer branches, floral stems 8 and flowers than both M. guttatus groups. Invasive M. guttatus produced fewer floral stems and 9 flowers than native *M. guttatus* ($\chi^2 > 13.928$; *P* < 0.001) [Supplementary Information - Fig. S4 10 and Tables S4 and S5]. Warm treatments strongly accelerated germination and flowering in all 11 12 population types, increased flower production, most significantly in native M. guttatus, and decreased corolla width, most significantly in invasive *M. guttatus*. Warm treatments also 13 14 increased plant height in all groups, more sharply in M. guttatus than in M. × robertsii, increased internode length and decreased internode diameter and dry mass, most significantly in invasive 15 *M. guttatus*, and increased the number of branches in both *M. guttatus* groups ($\chi^2 > 9.094$; *P* < 16 0.002) [Supplementary Information - Fig. S4 and Tables S4 and S5]. Short days delayed 17 flowering in both *M. guttatus* groups, and strongly decreased the probability of flowering in 18 invasive M. guttatus and M. \times robertsii, the production of stems in M. guttatus, and flower 19 production in all groups, more markedly in *M. guttatus* than in *M.* × *robertsii*. Short days also 20 reduced plant height, most significantly in invasive M. guttatus and $M. \times robertsii$, decreased 21 internode length in both *M. guttatus* groups, and decreased internode diameter and dry mass, 22 most significantly in invasive *M. guttatus* ($\chi^2 > 8.911$; *P* < 0.002) [Supplementary Information -23

1 Fig. S4 and Tables S4 and S5]. The interaction of temperature and photoperiod had an effect on

2 the production of flowers in *M. guttatus*, with SC and LW treatments having the lowest and

3 greatest flower production, respectively ($\chi^2 > 14.412$; P < 0.001). The three-way interaction of

4 factors was always non-significant [Supplementary Information - Fig. S4 and Tables S4 and S5].

5 Phenotypic plasticity

6 The overall values for RDPI_t, RDPI_p and RDPI_{tp} were 0.307 \pm 0.031, 0.27 \pm 0.034 and 0.32 \pm 7 0.028 (mean \pm sd), respectively. RDPI_t estimates did not differ among groups for any trait after Bonferroni correction ($\chi^2 < 8.678$; P > 0.013). The RDPI_p estimates for flowering day, number 8 of flowers, floral stems, branches and plant height varied significantly among population types 9 $(\chi^2 > 11.991; P < 0.002)$. In most cases the *post hoc* tests indicated significantly greater plasticity 10 in M. × robertsii than in the other groups [Supplementary Information - Table S6]. RDPI_{tp} 11 estimates for dry mass were also significantly greater in M. × robertsii than in the other groups 12 $(\chi^2 = 13.655; P = 0.001)$ [Supplementary Information - Table S6]. MANOVAs found significant 13 14 differences among population types for $RDPI_{tv}$, $RDPI_t$, and $RDPI_v$ estimates (Pillai's trace = 0.642-0.903; F > 2.261; P < 0.01; Table 1). M. × robertsii showed the greatest RDPI values, 15 although the post hoc analyses showed only significantly differences in RDPI_p between M. × 16 robertsii and native *M. guttatus* (Table 1). In the GLMM pooling RDPI_t and RDPI_p estimates, all 17 fixed factors (RDPI type, population type and their interaction) were significant ($\chi^2 > 6.716$; P <18 0.02). *M. guttatus* had higher RDPI_t than RDPI_p estimates and the opposite was found in M. × 19 robertsii (differences were significant only within native M. guttatus). RDPI_p estimates of M. × 20 robertsii were significantly higher than RDPI_t estimates of native *M. guttatus*. All RDPI 21 estimates for the production of stolons were similar for all population types ($\chi^2 < 5.034$; P >22 0.081) [Supplementary Information - Table S6]. 23

1 Experiment 2: reciprocal transplants

2 Local adaptation

The survival and flowering of plants was respectively above 98% and 96% across the 3 experiment. The fruit set of *M. guttatus* populations was significantly dependent on the 4 experimental site and the interaction of experimental site and population origin ($\chi^2 > 50.669$; P <5 0.001; Table 2). This species produced less fruits in SHE than in IOW, with a significant higher 6 decrease for southern populations (Fig. 4). The production of stolons was not significantly 7 dependent on any modelled factor in *M. guttatus* (Table 2), but it was also dependent on the 8 experimental site and the interaction of experimental site and population origin in M. × robertsii 9 $(\chi^2 > 14.81; P < 0.001;$ Table 2). Overall, the production of stolons in *M*. × robertsii was higher 10 in SHE than in IOW, and this was based on a high increase in northern populations. On the 11 contrary, southern populations showed a slightly lower production of stolons in SHE than in 12 13 IOW (Fig. 4).

14 Phenotypic differentiation

Across most traits and for both species, plants were similar regardless of their latitudinal origin. 15 In M. guttatus, northern individuals produced flowers with bigger corollas than southern 16 individuals ($\chi^2 = 6.566$; P = 0.01) [Supplementary Information - Table S7 and Fig. S5]. 17 Experimental site had a strong effect on the development of plants. M. guttatus individuals 18 flowered later, produced less flowers, had lower stomata density and grew less according to plant 19 cover and final dry mass in SHE than in IOW ($\chi^2 > 24.524$; P < 0.001). $M. \times robertsii$ flowered 20 later, produced fewer floral stems and flowers, had lower dry mass, and produced more stolons, 21 in SHE than in IOW ($\chi^2 > 8.4$; P < 0.01) [Supplementary Information - Table S7 and Fig. S5]. 22

The interaction site x origin was significant for the number of branches, stems and flowers in *M*.
 guttatus, with negative estimates for south plants in SHE. In *M*. × *robertsii*, site x origin was
 significant for the number of flowers, with positive estimates for south plants in SHE (χ² > 9.933;
 P < 0.01) [Supplementary Information - Table S7 and Fig. S5].

5 Phenotypic selection

6 The selection gradients differed between species, fitness traits and sites, suggesting diverse 7 mechanisms of local adaptation in each species. In M. guttatus, fruit set (sexual fitness) in IOW showed significant positive linear selection and stabilizing selection on flowering day and dry 8 mass, and significant negative linear selection and disruptive selection in height (t > 2.453; P <9 0.015). In SHE there was significant positive linear selection and stabilizing selection in corolla 10 11 width (t > 2.09; P < 0.037; Fig. 5) [Supplementary Information - Table S8]. For M. guttatus 12 stolons (asexual fitness) we found only significant negative linear selection and disruptive selection in dry mass in SHE (t > 2.02; P < 0.044; Fig. 5). For M. x robertsii stolons we found 13 14 positive linear selection in dry mass in IOW, and significant positive linear selection and stabilizing selection in corolla width in SHE (t > 2.073; P < 0.042; Fig. 5) [Supplementary 15 Information - Table S8]. 16

17 Mimulus luteus

18 *M. luteus* produced significantly more fruits in IOW than in SHE ($\chi 2 > 96.962$; P < 0.001) and a 19 similar number of stolons in both sites ($\chi 2 = 3.329$; P = 0.068). There were not differences 20 between *M. luteus* and north *M. guttatus* in the sexual or asexual fitness overall ($\chi 2 < 4.336$; P > 21 0.1), but *M. luteus* produced relatively more fruits than north *M. guttatus* in SHE ($\chi 2 = 11.018$; P 22 < 0.001). The production of stolons was higher in *M. luteus* than in north *M. × robertsii* overall,

1	but it was lower in SHE ($\chi 2 > 9.078$; P < 0.003). The models of phenotypic traits showed that <i>M</i> .
2	luteus flowered later, produced more branches and floral stems, and had lower stomata density
3	and dry mass in SHE than in IOW ($\chi 2 > 8.234$; P < 0.01). The phenotypic traits of north <i>M</i> . ×
4	robertsii and <i>M. luteus</i> did not differ significantly, but north $M. \times robertsii$ produced relatively
5	less branches, floral stems and flowers than <i>M. luteus</i> in SHE (negative coefficients for north <i>M</i> .
6	\times robertsii in SHE; $\chi 2 > 9.596;$ P $< 0.01)$ [Supplementary Information - Table S9]. The
7	phenotypic selection analyses through fruit set in M. luteus showed significant positive linear
8	selection in dry mass, stabilizing selection in the number of branches and dry mass, and
9	disruptive selection in stomata density in IOW ($t > 2.169$; $P < 0.04$) [Supplementary Information
10	- Table S8 and Fig. S6]. The models regressing the number of stolons indicated positive linear
11	and stabilizing selection on stomata density in IOW ($t > 2.232$; $P < 0.035$) [Supplementary
12	Information - Table S8].

13 Discussion

14 Clonality, phenotypic and plasticity changes in invasive Mimulus

15 The reproductive systems of native M. guttatus, invasive M. guttatus, and invasive $M. \times robertsii$ showed a transition from a relatively higher investment in sexual organs (i.e. floral stems and 16 17 flowers) to higher clonality (i.e. stolons). Native and invasive M. guttatus were similar in other phenotypic traits, suggesting that the reproductive system has been under selection in the UK, 18 and thus supporting the important role of clonality in plant invasions (Pyšek, 1997; Song et al., 19 2013; Wang et al., 2017; Bock et al., 2018; Wang et al., 2019). Remarkably, annual forms of M. 20 21 guttatus without clonal propagation do not seem to have established in the introduced range of this species. Consistent with our results, van Kleunen and Fischer (2008) found greater clonality 22

in Scottish than in native populations of *M. guttatus*, which they related with the latitude of
populations, and suggested signatures of differentiation after the species introduction at the
phenotypic level. In contrast, we did not find differences in flowering time at this level as
suggested by genomic analyses of selective sweeps in invasive *Mimulus* populations (Puzey and
Vallejo-Marin, 2014).

Clonality has been associated with persistence at higher latitudes in Mimulus (Van 6 7 Kleunen and Fisher, 2008) and other taxa (e.g. Dorken and Eckert, 2001). Our experiment in the 8 controlled environment chambers allowed us to disentangle the particular drivers of this association revealing that, interestingly, warm temperatures increased clonality in both M. 9 10 guttatus groups, but decreased clonality in $M. \times robertsii$. Given the dependence of $M. \times$ *robertsii* on clonality for the long-term persistence of populations, we hypothesize that limited 11 ability to clone in warmer environments underlies the lower abundance of M. × robertsii in the 12 13 south of the UK (Hargreaves et al., 2014). Consistently, a previous study associated higher 14 thermal tolerance with wider distributions in *Mimulus* (Sheth and Angert, 2014). The mechanism by which some populations of M. × robertsii can persist in the south of the UK (Da Re *et al.*, 15 2020) given the reduced clonality we observe when northern populations are translocated, 16 remains to be established. 17

18 The flowering and growth of all *Mimulus* population types were similarly affected by the 19 temperatures and photoperiods associated to the latitudinal range of UK. In contrast, the 20 germination of seeds was only accelerated under warm temperature treatments consistent with 21 previous *Mimulus* work (Vickery, 1983). Warm treatments had their strongest effect on 22 accelerating the flowering phenology of individuals, while long day treatments had their 23 strongest effect on increasing the production of sexual organs. Warm temperatures also increased

1 the vertical growth of plants, but not thickness nor biomass, while long photoperiods increased 2 plant growth through all measured traits. Given the natural association of growth-promoting and growth-hindering conditions of temperature and photoperiod across latitudinal gradients, fine 3 local microclimatic variations superimposed to large scale environmental patterns might play an 4 5 important role in the performance of natural populations. In our reciprocal transplants, 6 individuals grew bigger, flowered earlier and produced more flowers and fruits in IOW than in SHE, suggesting that, overall, the positive effects of high temperatures in the south site 7 outperformed those of long photoperiods in the north site. 8

Phenotypic plasticity has been considered a distinctive trait of invasive species (Davidson 9 10 et al., 2011) which could be also under positive selection in introduced populations (Bossdorf et al., 2005; Richards et al., 2006; but see Godoy et al., 2011). In our experiment, average RDPI 11 estimates ranged 0.27-0.32 and did not differ between native and introduced populations of M. 12 13 guttatus. This suggests a role for phenotypic plasticity as a pre-adaptation in invasive M. guttatus (Vickery, 1974). Native and invasive populations of *M. guttatus* showed greater phenotypic 14 plasticity in response to temperature than to photoperiod. Given that photoperiod cycles are more 15 constant than temperature at the local scale, this result is consistent with the hypothesis that 16 phenotypic plasticity evolves in response to environmental variation (Via and Lande, 1985). 17 18 Consisting with the classic view that clonal species rely more on phenotypic plasticity than 19 sexual species to overcome environmental variation, our analyses indicated greater phenotypic plasticity in M. × robertsii than in M. guttatus (Lynch, 1984; Geng et al., 2007). The fact that 20 21 some phenotypic traits analysed may be related to individual performance may raise doubt about 22 whether higher plasticity in M. x robertsii is a result of the lower performance of this species under certain conditions. Remarkably, the production of stolons, the clearest performance trait, 23

showed similar plasticity in the two taxa, suggesting that this cannot explain either their different
 reaction norms in the reciprocal transplants.

3 Local adaptation in introduced Mimulus

4 In our reciprocal transplants experiment, we found robust patterns of local adaptation in 5 introduced sexual populations of *M*. guttatus and asexual populations of M. × robertsii. As far as 6 we are aware, our study is the first assessing rapid local adaptation in a strictly asexual plant 7 species. Although there are reports of local adaptation in natural populations of other asexual 8 multicellular organisms (e.g. Via, 1991; Ayre, 1995; Doroszuk et al., 2006), and in partly clonal plant populations (e.g. Lenssen et al., 2004), studies comparing sexual and asexual lineages are 9 10 scarce and mostly based in microorganisms under laboratory conditions (e.g. Colegrave, 2002; 11 McDonald et al., 2016; but see also Mariette et al., 2016). As remarkable exceptions in plants, 12 recent studies have compared sexual and asexual lineages of *Boechera* (Lovell et al., 2014; 13 Rushworth et al., 2020). In contrast to our, these studies found signs of local adaptation only in sexual lineages. Although we could not distinguish obvious ecotypes at each latitude for M. 14 15 guttatus nor for M. × robertsii, and all populations were capable to survive over one season in the two extremes of the country, both species showed significant home site advantages in their 16 respective sexual and asexual reproductive success. Parallel patterns were found in some related 17 18 traits in *M. guttatus* (branches, stems, flowers) and *M. × robertsii* (flowers, with the opposite trend). Reproductive traits can reveal local adaptation patterns more readily than survival 19 20 (Baughman et al., 2018), and their effects on the persistence of species are likely to act in the 21 longer term but unequivocally. Nevertheless, the low occurrence of M. \times robertsii in the south of the UK suggests that local adaptation may be more difficult to achieve in this taxon than in M. 22 23 guttatus. Only a few populations of M. \times robertsii seem to have overcome the challenges present

1	in the south through local genotypic adaptation, which may have been facilitated by restricted
2	dispersal opportunities (Ayre, 1995) in combination with more stressful environmental
3	conditions (Ram and Hadany, 2002).

4 Our study suggest that asexual reproduction does not necessarily constrain evolution at a contemporary time scale, and this is congruent with genomic studies in other asexual plant 5 6 lineages (Ferreira de Carvalho et al., 2016; Lovell et al., 2017). However, it has been also 7 suggested that an increased heterozygosity level of hybrid polyploids in comparison with their 8 diploid ancestors could boost their ability to adapt to different environments (Levin, 2002; Abbott et al., 2013; Vallejo-Marín and Hiscock, 2016; Meier et al, 2017). M. × robertsii differs 9 10 from both parents in showing local adaptation through as exual fitness, which might propel $M \times$ robertsii into an independent evolutionary trajectory from its parents. Further studies are 11 required to see if rapid local adaptation can be found also in non-hybrid asexual plants. In 12 13 contrast, the fitness and phenotypic patterns of *M. luteus* were similar to those of the two other species, discarding climatic constrains on the performance of this species as explanation for its 14 low occurrence in UK (cf. Da Re et al., 2020). Other environmental or ecological factors not 15 included in our experiment, such as soil tolerance or competition with M. × robertsii and M. 16 guttatus (Da Re et al., 2020), might be responsible of limiting the current distribution of M. 17 18 *luteus* in the UK.

The phenotypic selection analyses showed that few and different traits were related to the sexual and/or asexual fitness of *M. guttatus* and *M.* × *robertsii* at each site in our reciprocal transplants experiment. This suggest that different mechanisms may have driven the local adaptation in each species. In *M. guttatus*, large-flowered individuals had greater sexual fitness in SHE, while shorter, heavier and late-flowering individuals had greater sexual fitness in IOW.

1 Remarkably, flowering time is considered a principal trait under selection during species range 2 expansions (Barrett et al., 2008), and in local adaptation and speciation in native Mimulus (e.g. Hall and Willis, 2006; Friedman and Willis, 2013). In M. x robertsii, heavier individuals had 3 greater asexual fitness in IOW, and large-flowered individuals had greater asexual fitness in 4 SHE. The later result contrasts with the common finding of trade-offs between sexual and 5 6 asexual allocation in sexual *Mimulus* species (Sutherland and Vickery, 1988), and might be an indirect consequence of resources acquisition determined by individual quality. The selection 7 gradients estimated for *M. luteus* were also highly different to those of *M. guttatus* and *M.* x 8 9 robertsii. The production of fruits was positively associated with dry mass in IOW and with flower production in SHE, while the production of stolons was positively associated with 10 stomata density in IOW. Overall, our results present partial support for a previous study which, 11 comparing native and invasive populations of *M. guttatus*, found that introduced populations 12 showed adaptative differentiation though selection on various traits, including large vegetative 13 size and large floral displays and flower size (Pantoja et al., 2018). 14 The traits underlying the local adaptation of M. \times robertsii in UK are yet to be fully 15 identified, and thus populations of this species are an ideal target for further research on the

identified, and thus populations of this species are an ideal target for further research on the
mechanisms mediating rapid evolution in asexual species (see also Rushworth *et al.*, 2020).
Selection on clonal taxa could occur through genotypic selection in genetically diverse founding
populations (clonal selection), or, perhaps, through other mechanisms including epigenetic
modification (Wilschut *et al.*, 2016). Although further comparisons between sexual and asexual
taxa in other suitable natural systems are needed for inferences on the evolutionary rates and
mechanisms of asexual taxa across plant lineages, our study provides a starting point for
understanding the early evolutionary trajectory of invasive asexual plant populations.

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5

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16 Literature cited

- Abbott R, Albach D, Ansell S, Arntzen JW, Baird SJE, Bierne N, *et al.* 2013. Hybridization and
 speciation. *Journal of Evolutionary Biology* 26: 229–246.
- 19 Abramoff MD, Magelhaes PJ, Ram SJ. 2004. Image processing with ImageJ. *Biophotonics*

20 *International* 11: 36–42.

1	Ayre DJ. 1995. Localized adaptation of sea anemone clones: evidence from transplantation over
2	two spatial scales. Journal of Animal Ecology 64: 186–196.
3	Barrett SCH, Colautti RI, Eckert CG. 2008. Plant reproductive systems and evolution during
4	biological invasion. Molecular Ecology 17: 373–383.
5	Bates D, Maechler M, Bolker B, Walker S. 2015. Fitting linear mixed-effects models using lme4.
6	Journal of Statistical Software 67: 1–48.
7	Baughman OW, Agneray AC, Forister ML, Kilkenny FF, Espeland EK, Fiegener R, et al. 2019.
8	Strong patterns of intraspecific variation and local adaptation in Great Basin plants
9	revealed through a review of 75 years of experiments. Ecology and Evolution 9: 6259-
10	6275.
11	Bhattarai GP, Meyerson LA, Anderson J, Cummings D, Allen WJ, Cronin JT. 2017.
12	Biogeography of a plant invasion: genetic variation and plasticity in latitudinal clines for
13	traits related to herbivory. Ecological Monographs 87: 57–75.
14	Blanquart F, Kaltz O, Nuismer SL, Gandon S. 2013. A practical guide to measuring local
15	adaptation. Ecology Letters 16: 1195–1205.
16	Bossdorf O, Auge H, Lafuma L, Rogers WE, Siemann E, Prati D. 2005. Phenotypic and genetic
17	differentiation between native and introduced plant populations. Oecologia 144: 1–11.
18	Bradshaw AD. 1965. Evolutionary significance of phenotypic plasticity in plants. Advances in
19	Genetics 13: 115–155.
20	Burt A. 2000. Sex, recombination, and the efficacy of selection—was Weismann right?
21	Evolution 54: 337–351.
22	Colautti RI, Lau JA. 2015. Contemporary evolution during invasion: evidence for differentiation,
23	natural selection, and local adaptation. Molecular Ecology 24: 1999–2017.

1	Colegrave N. 2002. Sex releases the speed limit on evolution. Nature 420: 664.
2	Crow JF, Kimura M. 1965. Evolution in sexual and asexual populations. The American
3	Naturalist 99: 439–450.
4	Da Re D, Olivares AP, Smith W, Vallejo-Marin M. 2020. Global analysis of ecological niche
5	conservation and niche shift in exotic populations of monkeyflowers (Mimulus guttatus,
6	<i>M. luteus</i>) and their hybrid (M . × robertsii). Plant Ecology and Diversity.
7	https://doi.org/10.1080/17550874.2020.1750721.
8	Davidson AM, Jennions M, Nicotra AB. 2011. Do invasive species show higher phenotypic
9	plasticity than native species and, if so, is it adaptive? A meta-analysis. Ecology Letters
10	14: 419–431.
11	de Carvalho JF, de Jager V, van Gurp TP, Wagemaker NC, Verhoeven KJ. 2016. Recent and
12	dynamic transposable elements contribute to genomic divergence under asexuality. BMC
13	Genomics 17: 884.
14	de Jong G. 2005. Evolution of phenotypic plasticity: patterns of plasticity and the emergence of
15	ecotypes. New Phytologist 166: 101–118.
16	Dorken ME, Eckert CG. 2001. Severely reduced sexual reproduction in northern populations of a
17	clonal plant, Decodon verticillatus (Lythraceae). Journal of Ecology 89: 339-350.
18	Doroszuk A, Wojewodzic MW, Kammenga JE. 2006. Rapid adaptive divergence of life-history
19	traits in response to abiotic stress within a natural population of a parthenogenetic
20	nematode. Proceedings of the Royal Society B: Biological Sciences 273: 2611–2618.
21	Ebeling SK, Stöcklin J, Hensen I, Auge H. 2011. Multiple common garden experiments suggest
22	lack of local adaptation in an invasive ornamental plant. Journal of Plant Ecology 4:
23	209–220.

1	Fazlioglu F, Bonser SP. 2016. Phenotypic plasticity and specialization in clonal versus non-
2	clonal plants: A data synthesis. Acta Oecologica 77: 193-200.
3	Fox J, Weisberg S, Adler D, Bates D, Baud-Bovy G, Ellison S, et al. 2012. Package 'car'.
4	Vienna: R Foundation for Statistical Computing.
5	Friedman J, Willis JH. 2013. Major QTL s for critical photoperiod and vernalization underlie
6	extensive variation in flowering in the Mimulus guttatus species complex. New
7	Phytologist 199: 571–583.
8	Geng Y, Pan XY, Xu CY, Zhang WJ, Li B, Chen JK, et al. 2007. Phenotypic plasticity rather
9	than locally adapted ecotypes allows the invasive alligator weed to colonize a wide range
10	of habitats. Biological Invasions 9: 245–256.
11	Geng Y, van Klinken RD, Sosa A, Li B, Chen J, Xu CY. 2016. The relative importance of
12	genetic diversity and phenotypic plasticity in determining invasion success of a clonal
13	weed in the USA and China. Frontiers in Plant Science 7: 213.
14	Ghalambor CK, McKay JK, Carroll SP, Reznick DN. 2007. Adaptive versus non-adaptive
15	phenotypic plasticity and the potential for contemporary adaptation in new environments.
16	Functional Ecology 21: 394–407.
17	Godoy O, Valladares F, Castro-Díez P. 2011. Multispecies comparison reveals that invasive and
18	native plants differ in their traits but not in their plasticity. Functional Ecology 25: 1248-
19	1259.
20	Grant AL. 1924. A monograph of the genus Mimulus. Annals of the Missouri Botanical Garden
21	11: 99–388.
22	Hall MC, Willis JH. 2006. Divergent selection on flowering time contributes to local adaptation
23	in Mimulus guttatus populations. Evolution 60: 2466–2477.

1	Hargreaves AL, Samis KE, Eckert CG. 2014. Are species' range limits simply niche limits writ
2	large? A review of transplant experiments beyond the range. The American Naturalist
3	183: 157–173.
4	Hereford J. 2009. A quantitative survey of local adaptation and fitness trade-offs. The American
5	Naturalist 173: 579–588.
6	Hijmans RJ, Cameron SE, Parra JL, Jones PG, Jarvis A. 2005. Very high resolution interpolated
7	climate surfaces for global land areas. International Journal of Climatology 25: 1965-
8	1978.
9	Hijmans RJ, Williams E, Vennes C. 2014. Geosphere: spherical trigonometry. R package version
10	1.3-11.
11	Howell CJ, Sawyer JWD. 2006. New Zealand naturalised vascular plant checklist. Wellington:
12	New Zealand Plant Conservation Network.
13	Kawecki TJ, Ebert D. 2004. Conceptual issues in local adaptation. Ecology Letters 7: 1225-
14	1241.
15	Kelly M. 2019. Adaptation to climate change through genetic accommodation and assimilation
16	of plastic phenotypes. Philosophical Transactions of the Royal Society B 374: 20180176.
17	Lambertini C, Riis T, Olesen B, Clayton JS, Sorrell BK, Brix H. 2010. Genetic diversity in three
18	invasive clonal aquatic species in New Zealand. BMC Genetics 11:52.
19	Lande R, Arnold SJ. 1983. The measurement of selection on correlated characters. <i>Evolution</i> 37:
20	1210–1226.
21	Leger EA, Rice KJ. 2007. Assessing the speed and predictability of local adaptation in invasive
22	California poppies (Eschscholzia californica). Journal of Evolutionary Biology 20: 1090-
23	1103.

1	Lenssen JP, Van Kleunen M, Fischer M, De Kroon H. 2004. Local adaptation of the clonal plant
2	Ranunculus reptans to flooding along a small-scale gradient. Journal of Ecology 92:
3	696–706.
4	Lenth R, Lenth MR. 2018. Package 'Ismeans'. The American Statistician 34: 216–221.
5	Levin D. 2002. The Role of Chromosomal Change in Plant Evolution. Oxford University Press.
6	Li XM, She DY, Zhang DY, Liao WJ. 2015. Life history trait differentiation and local adaptation
7	in invasive populations of Ambrosia artemisiifolia in China. Oecologia 177: 669–677.
8	Liao H, D'Antonio CM, Chen B, Huang Q, Peng S. 2016. How much do phenotypic plasticity
9	and local genetic variation contribute to phenotypic divergences along environmental
10	gradients in widespread invasive plants? A meta-analysis. Oikos 125: 905-917.
11	Liu Y, Zhang L, Xu X, Niu H. 2016. Understanding the wide geographic range of a clonal
12	perennial grass: plasticity versus local adaptation. Annals of Botany Plants 8: plv141.
13	Liu W, Zhang Y, Chen X, Maung-Douglass K, Strong DR, Pennings SC. 2020. Contrasting plant
14	adaptation strategies to latitude in the native and invasive range of Spartina alterniflora.
15	New Phytologist 226: 623–634.
16	Lovell JT, Grogan K, Sharbel TF, McKay JK. 2014. Mating system and environmental variation
17	drive patterns of adaptation in Boechera spatifolia (Brassicaceae). Molecular Ecology 23:
18	4486–4497.
19	Lovell JT, Williamson RJ, Wright SI, McKay JK, Sharbel TF. 2017. Mutation accumulation in
20	an asexual relative of Arabidopsis. PLoS Genetics 13: e1006550.
21	Lowry DB, Willis JH. 2010. A widespread chromosomal inversion polymorphism contributes to
22	a major life-history transition, local adaptation, and reproductive isolation. PLoS Biology
23	8: e1000500.

1	Lowry DB, Sobel JM, Angert AL, Ashman T-L, Baker RL, Blackman BK, et al. 2019. The case
2	for the continued use of the genus name Mimulus for all monkeyflowers. Taxon 68: 617-
3	623.
4	Lucek K, Sivasundar A, Seehausen O. 2014. Disentangling the role of phenotypic plasticity and
5	genetic divergence in contemporary ecotype formation during a biological invasion.
6	Evolution 68: 2619–2632.
7	Lushai G, Loxdale HD, Allen JA. 2003. The dynamic clonal genome and its adaptive potential.
8	Biological Journal of the Linnean Society 79: 193–208.
9	Lynch M. 1984. Destabilizing hybridization, general-purpose genotypes and geographic
10	parthenogenesis. The Quarterly Review of Biology 59: 257-290.
11	Marchini GL, Maraist CA, Cruzan MB. 2018. Trait divergence, not plasticity, determines the
12	success of a newly invasive plant. Annals of Botany 123: 667-679.
13	Mariette N, Androdias A, Mabon R, Corbiere R, Marquer B, Montarry J, et al. 2016. Local
14	adaptation to temperature in populations and clonal lineages of the Irish potato famine
15	pathogen Phytophthora infestans. Ecology and Evolution 6: 6320-6331.
16	Maron JL, Vilà M, Bommarco R, Elmendorf S, Beardsley P. 2004. Rapid evolution of an
17	invasive plant. Ecological Monographs 74: 261–280.
18	Maynard Smith J. 1968. Evolution in sexual and asexual populations. The American Naturalist
19	102: 469–473.
20	McDonald MJ, Rice DP, Desai MM. 2016. Sex speeds adaptation by altering the dynamics of
21	molecular evolution. Nature 531: 233–236.

1	Meeus S, Šemberová K, De Storme N, Geelen D, Vallejo-Marín, M. 2020. Effect of whole-
2	genome duplication on the evolutionary rescue of sterile hybrid monkeyflowers. Plant
3	Communications. https://doi.org/10.1016/j.xplc.2020.100093.
4	Meier JI, Marques DA, Mwaiko S, Wagner CE, Excoffier L, Seehausen O. 2017. Ancient
5	hybridization fuels rapid cichlid fish adaptive radiations. Nature Communications 8: 1–
6	11.
7	Michel A, Arias RS, Scheffler BE, Duke SO, Netherland M, Dayan FE. 2004. Somatic mutation-
8	mediated evolution of herbicide resistance in the nonindigenous invasive plant hydrilla
9	(Hydrilla verticillata). Molecular Ecology 13: 3229–3237.
10	Mitchell N, Whitney KD. 2018. Can plants evolve to meet a changing climate? The potential of
11	field experimental evolution studies. American Journal of Botany 105: 1613–1616.
12	Nesom GL. 2014. Updated classification and hypothetical phylogeny of <i>Erythranthe</i> sect.
13	Simiola (Phrymaceae). Phytoneuron 81: 1–6.
14	Novy A, Flory SL, Hartman JM. 2013. Evidence for rapid evolution of phenology in an invasive
15	grass. Journal of Evolutionary Biology 26: 443-450.
16	Oduor AM, Leimu R, van Kleunen M. 2016. Invasive plant species are locally adapted just as
17	frequently and at least as strongly as native plant species. Journal of Ecology 104: 957-
18	968.
19	Oplaat C, Verhoeven KJ. 2015. Range expansion in asexual dandelions: selection for general-
20	purpose genotypes? Journal of Ecology 103: 261-268.
21	Pahl AT, Kollmann J, Mayer A, Haider S. 2013. No evidence for local adaptation in an invasive
22	alien plant: field and greenhouse experiments tracing a colonization sequence. Annals of
23	Botany 112: 1921–1930.

1	Pantoja PO, Simón-Porcar VI, Puzey JR, Vallejo-Marín M. 2017. Genetic variation and clonal
2	diversity in introduced populations of Mimulus guttatus assessed by genotyping at 62
3	single nucleotide polymorphism loci. Plant Ecology & Diversity 10: 5–15.
4	Pantoja PO, Paine CT, Vallejo-Marín M. 2018. Natural selection and outbreeding depression
5	suggest adaptive differentiation in the invasive range of a clonal plant. Proceedings of the
6	Royal Society B: Biological Sciences 285: 20181091.
7	Parker PF. 1975. Mimulus in Great Britain-a cytotaxonomic note. New Phytologist 74: 155–160.
8	Preston CD, Pearman D, Dines TD. 2002. New atlas of the British & Irish flora. Oxford
9	University Press.
10	Price TD, Qvarnström A, Irwin DE. 2003. The role of phenotypic plasticity in driving genetic
11	evolution. Proceedings of the Royal Society of London. Series B: Biological Sciences
12	270: 1433–1440.
13	Puzey JR, Vallejo-Marín M. 2014. Genomics of invasion: diversity and selection in introduced
14	populations of monkeyflowers (Mimulus guttatus). Molecular Ecology 23: 4472–4485.
15	Pyšek P. 1997. Clonality and plant invasions: can a trait make a difference? In: de Kroon H, van
16	Groenendael J, eds. The ecology and evolution of clonal plants. Leiden: Backhuys
17	Publishers, 405–427.
18	R Core Team. 2019. R: A language and environment for statistical computing. Vienna: R
19	Foundation for Statistical Computing. https://www.R-project.org/.
20	Ram Y, Hadany L. 2012. The evolution of stress-induced hypermutation in asexual populations.
21	Evolution 66: 2315–2328.

Raven JA. 2002. Selection pressure on stomatal evolution. *New Phytologist* 153: 371–386.

1	Richards CL, Bossdorf O, Muth NZ, Gurevitch J, Pigliucci M. 2006. Jack of all trades, master of
2	some? On the role of phenotypic plasticity in plant invasions. <i>Ecology Letters</i> 9: 981–
3	993.
4	Riis T, Lambertini C, Olesen B, Clayton JS, Brix H, Sorrell BK. 2010. Invasion strategies in
5	clonal aquatic plants: are phenotypic differences caused by phenotypic plasticity or local
6	adaptation? Annals of Botany 106: 813-822.
7	Roberts RH. 1964. Mimulus hybrids in Britain. Watsonia 6: 70-75.
8	Roiloa SR. 2019. Clonal traits and plant invasiveness: the case of Carpobrotus NE Br.
9	(Aizoaceae). Perspectives in Plant Ecology, Evolution and Systematics 125479.
10	Rouzine IM, Wakeley J, Coffin JM. 2003. The solitary wave of asexual evolution. Proceedings
11	of the National Academy of Sciences 100: 587–592.
12	Rushworth CA, Brandvain Y, Mitchell-Olds T. 2020. Identifying the fitness consequences of sex
13	in complex natural environments. Evolution Letters. https://doi.org/10.1002/evl3.194.
14	Schiffels S, Szöllősi GJ, Mustonen V, Lässig M. 2011. Emergent neutrality in adaptive asexual
15	evolution. Genetics 189: 1361–1375.
16	Sheth SN, Angert AL. 2014. The evolution of environmental tolerance and range size: a
17	comparison of geographically restricted and widespread Minulus. Evolution 68: 2917-
18	2931.
19	Silvertown J. 2008. The evolutionary maintenance of sexual reproduction: evidence from the
20	ecological distribution of asexual reproduction in clonal plants. International Journal of
21	Plant Sciences 169: 157–168.

1	Simón-Porcar VI, Silva JL, Meeus S, Higgins JD, Vallejo-Marín M. 2017. Recent
2	autopolyploidization in a naturalized population of Mimulus guttatus (Phrymaceae).
3	Botanical Journal of the Linnean Society 185: 189–207.
4	Song YB, Yu FH, Keser LH, Dawson W, Fischer M, Dong M, et al. 2013. United we stand,
5	divided we fall: a meta-analysis of experiments on clonal integration and its relationship
6	to invasiveness. Oecologia 171: 317-327.
7	Stace CA. 2010. New Flora of the British Isles. Cambridge: Cambridge University Press.
8	Sutherland S, Vickery RK. 1988. Trade-offs between sexual and asexual reproduction in the
9	genus Mimulus. Oecologia 76: 330–335.
10	Thompson JN. 1998. Rapid evolution as an ecological process. Trends In Ecology and Evolution
11	13: 329–332.
12	Truscott AM, Soulsby C, Palmer SCF, Newell L, Hulme PE. 2006. The dispersal characteristics
13	of the invasive plant Mimulus guttatus and the ecological significance of increased
14	occurrence of high-flow events. Journal of Ecology 94: 1080-1091.
15	Valladares F, Sánchez-Gómez D, Zavala MA. 2006. Quantitative estimation of phenotypic
16	plasticity: bridging the gap between the evolutionary concept and its ecological
17	applications. Journal of Ecology 94: 1103–1116.
18	Vallejo-Marin M, Lye GC. 2013. Hybridisation and genetic diversity in introduced Mimulus
19	(Phrymaceae). Heredity 110: 111.
20	Vallejo-Marín M, Hiscock SJ. 2016. Hybridization and hybrid speciation under global change.
21	New Phytologist 211: 1170–1187.
22	Song YB, Yu FH, Keser LH, Dawson W, Fischer M, Dong M, et al. 2013.

1	Vallejo-Marín M, Friedman J, Twyford AD, Lepais O, Ickert-Bond SM, Streisfeld MA, et al.
2	2020. Population genomic and historical analysis reveals a global invasion by bridgehead
3	processes in Mimulus guttatus. bioRxiv.
4	van Kleunen M, Fischer M. 2001. Adaptive evolution of plastic foraging responses in a clonal
5	plant. Ecology 82: 3309–3319.
6	van Kleunen M, Fischer M. 2008. Adaptive rather than non-adaptive evolution of Mimulus
7	guttatus in its invasive range. Basic and Applied Ecology 9: 213-223.
8	Vandepitte K, De Meyer T, Helsen K, Van Acker K, Roldán-Ruiz I, Mergeay J, et al. 2014.
9	Rapid genetic adaptation precedes the spread of an exotic plant species. Molecular
10	<i>Ecology</i> 23: 2157–2164.
11	Via S. 1991. The genetic structure of host plant adaptation in a spatial patchwork – demographic
12	variability among reciprocally transplanted pea aphid clones. Evolution 45: 827-852.
13	Via S, Lande R. 1985. Genotype-environment interaction and the evolution of phenotypic
14	plasticity. Evolution 39: 505–522.
15	Vickery RK Jr. 1983. Plasticity and polymorphism in seed germination of Mimulus guttatus
16	(Scrophulariaceae). The Great Basin Naturalist 43: 470-474.
17	Vickery RK Jr. 1974. Crossing barriers in the yellow monkeyflowers in the genus Mimulus
18	(Scrophulariaceae). Genetics Lectures 3: 33-82.
19	Vrijenhoek RC. 1979. Factors affecting clonal diversity and coexistence. American Zoologist 19:
20	787–797.
21	Wang YJ, Müller-Schärer H, van Kleunen M, Cai AM, Zhang P, Yan R, et al. 2017. Invasive
22	alien plants benefit more from clonal integration in heterogeneous environments than
23	natives. New Phytologist 216: 1072–1078.

1	Wang YJ, Chen D, Yan R, Yu FH, van Kleunen M. 2019. Invasive alien clonal plants are
2	competitively superior over co-occurring native clonal plants. Perspectives in Plant
3	Ecology, Evolution and Systematics 40: 125484.
4	Weissmann A. 1889. The significance of sexual reproduction in the theory of natural selection.
5	In: Poulton EB, Schonland S, Shipley AE, eds. Essays upon heredity and kindred
6	biological problems. Oxford: Clarendon, 251–332.
7	West-Eberhard MJ. 2005. Developmental plasticity and the origin of species differences.
8	Proceedings of the National Academy of Sciences 102: 6543–6549.
9	Williams GC. 1966. Adaptation and Natural Selection. Princeton: Princeton University Press.
10	Wilschut RA, Oplaat C, Snoek LB, Kirschner J, Verhoeven KJ. 2016. Natural epigenetic
11	variation contributes to heritable flowering divergence in a widespread asexual dandelion
12	lineage. Molecular Ecology 25: 1759–1768.
13	Wu CA, Lowry DB, Cooley AM, Wright KM, Lee YW, Willis JH. 2008. Mimulus is an
14	emerging model system for the integration of ecological and genomic studies. Heredity
15	100: 220.

16 Figure	e legends.
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18	Figure 1. Experimental set ups to test for phenotypic plasticity and local adaptation of <i>Mimulus</i>
19	in the UK. Seedlings growing in a controlled environmental chamber at the University of Stirling
20	in 2014 (a). Common gardens set in the Isle of Wight (b) and Shetland (c) in summer 2015.
21	
22	Figure 2. Map of populations (circles) and experimental sites (triangles) used in the reciprocal

transplants. 23

24

25 Figure 3. Clonality of M. × robertsii (ROB), introduced M. guttatus (UK) and native Mimulus guttatus (US) populations grown in four different controlled environmental chambers with 26 contrasting photoperiods (L: long; S: short) and temperatures (C: cold; W: warm) in a crossed 27 design. Mean values and standard errors of the variables measured are indicated by dots and 28 29 error bars, respectively.

30

31	Figure 4. Fitness of <i>M. guttatus</i> and <i>M.</i> \times robertsii individuals included in the reciprocal
32	transplant experiment between different latitudes in the UK. Mean values and standard errors of
33	the number of fruits and stolons are indicated by dots and error bars, respectively.

34

Figure 5. Estimates and 95% confidence intervals for the phenotypic selection coefficients on
each trait included in the selection gradient analyses of *M. guttatus* and *M.* × robertsii in the
reciprocal transplants experiments at each site. Significant values are in blue. The fitness
measures used were fruits and stolons, respectively.

39 Tables

- 40 **Table 1.** Comparison of phenotypic plasticity among population types. Results of the MANOVAs and *post hoc* tests analysing RDPI phenotypic
- 41 plasticity indexes for all traits measured in the controlled environmental chambers as a function of the population type. Mean \pm s.d. RDPI values
- 42 across traits for each population type are provided jointly with the results of the post hoc tests. Different letters indicate significant differences. *
- 43 P<0.05; ** P<0.01; *** P<0.001.

44

	MA	NOVA					Mean \pm sd; <i>post hoc</i>				
	Df	df residuals	Pillai's trace	approx F	num df	den df	Р	<i>M. guttatus</i> native	<i>M. guttatus</i> invasive	<i>M</i> . × robertsii	
RDPIt	2	50	0.642	2.261	18	86	0.006 **		0.291 ± 0.042 a	0.318 ± 0.071 a	
RDPIp	2	47	0.830	3.154	18	80	< 0.001 ***	0.218 ± 0.041 a	$0.247 \pm 0.056 \text{ ab}$	$0.354 \pm 0.078 \ b$	
RDPItp	2	32	0.903	2.287	18	50	0.011 *	0.299 ± 0.038 a	0.317 ± 0.046 a	0.346 ± 0.065 a	

46 **Table 2.** Results of the GLMMs modelling the effects of experimental site (S), population origin

47 (O) and their interaction in the sexual and asexual fitness traits recorded in a reciprocal transplant

Trait	Fixed factor	<i>M. guttatus</i> Estimate (SE)	χ ²	<i>M.</i> × <i>robertsii</i> Estimate (SE)	χ ²
Fruits	Intercept	2.776 (0.266)	108.875 ***		
	\mathbf{W}_0	0 (0.001)	0.003		
	O (South)	0.317 (0.353)	0.805		
	Site (SHE)	-0.599 (0.038)	248.638 ***		
	Site:O (SHE:South)	-0.357 (0.05)	50.669 ***		
Stolons	Intercept	1.718 (0.14)	150.518 ***	1.409 (0.127)	122.591 ***
	\mathbf{W}_0	0.001 (0.001)	0.235	0 (0)	0.016
	O (South)	0.132 (0.119)	1.235	0.314 (0.233)	1.809
	Site (SHE)	0.07 (0.061)	1.295	0.467 (0.072)	41.635 ***
	Site:O (SHE:South)	-0.033 (0.073)	0.212	-0.522 (0.136)	14.81 ***

48 experiment with introduced *Mimulus guttatus* and *M*. × *robertsii* populations.

49

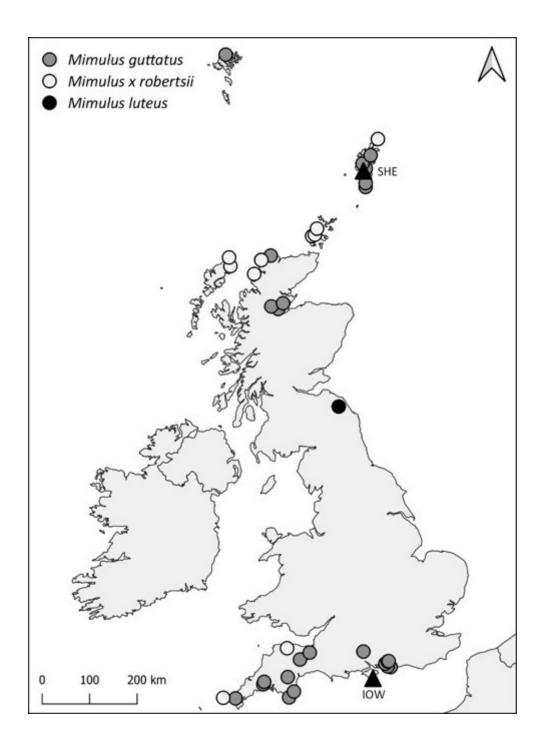
50 Table note: Model Estimates and Standard Errors for each fixed factor and interaction are

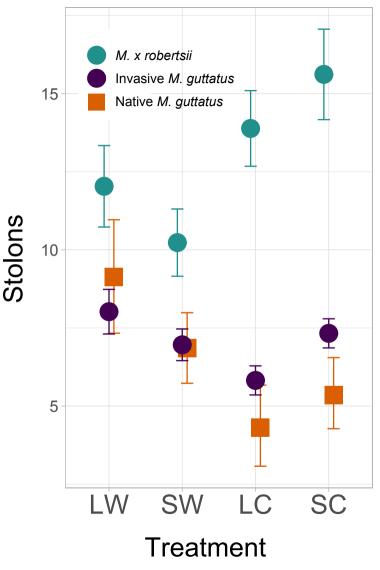
provided jointly with results of the type-III Wald χ^2 tests. χ^2 values and indications of their

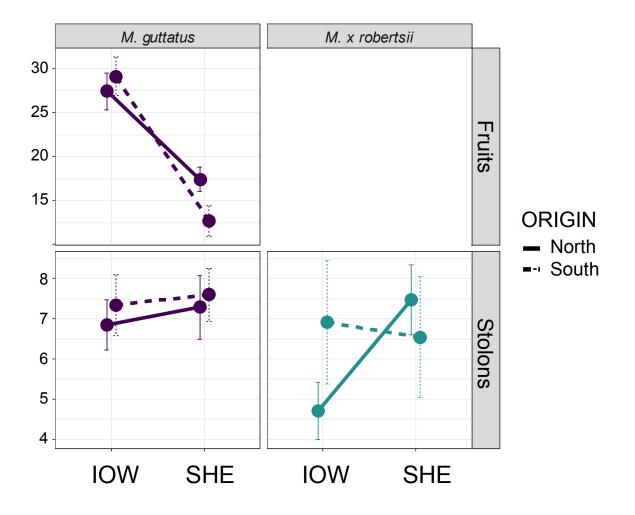
52 associated *P*-values are provided. * P<0.05; ** P<0.01; *** P<0.001. Significant effects after

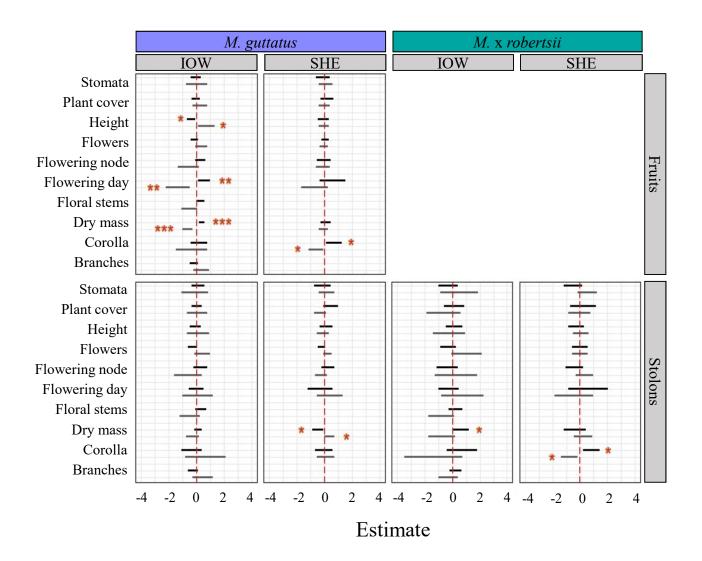
53 Bonferroni correction of *P*-values are indicated in bold. χ^2 degrees of freedom=1.











1 Supplementary material

- 2 **Table S1.** Source populations for the controlled environment chambers experiment. N indicates number of families per population.
- 3 Two individuals per family (seedlings per maternal plant in M. guttatus; cuttings per individual in M. × robertsii), were used in the
- 4 experiment. G = M. guttatus; R = M. x robertsii.

Population	Herbarium Accession	Species	Range	Locality	Lat.	Long.	m a.s.l.	N (Families)
BOD		G	Introduced		59.9042	-1.3027	44	5
BRA		G	Introduced		52.7681	-1.2979	12	5
DBL		G	Introduced Introduced	Dunblane, Stirlingshire Houghton Lodge,	56.1886	-3.9661	64	5
HOU		G		Hampshire	51.0970	-1.5084	33	5
TOM		G	Introduced	_	57.2550	-3.3678	318	5
CPB		G	Native		53.1710	-131.785	12	4
DAV		G	Native		37.0250	-122.2175	6	3
ALA		G	Native					
	V153408				59.793	-141.085		1
	V127607				62.70	-150.32		1
	V142998				59.05	-155.85		1
ORO		G	Native		35.2733	-120.8891	11	4
WTB		G	Native		38.4053	-123.0961	35	4
ALS		R	Introduced		54.8149	-2.4292	299	4
GON		R	Introduced		55.4668	-3.7377	285	4
GOO		R	Introduced		57.1620	-3.1863	357	3
NEN		R	Introduced		54.8061	-2.3764	355	1
WAN		R	Introduced		55.3973	-3.7804	405	3

- 6 **Table S2.** Temperature and photoperiod conditions of the environmental treatments implemented in the chambers experiment.
- 7 Treatment codes indicate photoperiod (L, Long; S, Short) and temperature (W, Warm; C, Cold). All chambers were set to a relative

SEGMENT		SW	SW		LC		SC		LW	
(FORTNIGTH)		duration	temp C							
1	DAY	12 h 54 m	13.2	13 h 12 m	7.3	12 h 54 m	7.3	13 h 12 m	13.2	
(1-15 April)	NIGTH	11 h 6 m	5.3	10 h 48 m	2.6	11 h 6 m	2.6	10 h 48 m	5.3	
2	DAY	13 h 48 m	14.8	14 h 36 m	8.6	13 h 48 m	8.6	14 h 36 m	14.8	
(16-30 April)	NIGTH	10 h 12 m	6.6	9 h 24 m	4	10 h 12 m	4	9 h 24 m	6.6	
3	DAY	14 h 42 m	16.4	16 h 0 m	9.8	14 h 42 m	9.8	16 h 0 m	16.4	
(1-15 May)	NIGTH	9 h 18 m	7.9	8 h 0 m	5.3	9 h 18 m	5.3	8 h 0 m	7.9	
4	DAY	15 h 30 m	18.1	17 h 18 m	11.2	15 h 30 m	11.2	17 h 18 m	18.1	
(16-31 May)	NIGTH	8 h 30 m	9.5	6 h 42 m	6.7	8 h 30 m	6.7	6 h 42 m	9.5	
5	DAY	16 h 12 m	19.8	18 h 30 m	12.6	16 h 12 m	12.6	18 h 30 m	19.8	
(1-15 June)	NIGTH	7 h 48 m	11	5 h 30 m	8.1	7 h 48 m	8.1	5 h 30 m	11	
6	DAY	16 h 30 m	20.8	19 h 6 m	13.1	16 h 30 m	13.1	19 h 6 m	20.8	
(16-30 June)	NIGTH	7 h 30 m	11.8	4 h 54 m	8.8	7 h 30 m	8.8	4 h 54 m	11.8	
7	DAY	16 h 24 m	21.7	19 h 0 m	13.6	16 h 24 m	13.6	19 h 0 m	21.7	
(1-15 July)	NIGTH	7 h 36 m	12.6	5 h 0 m	9.5	7 h 36 m	9.5	5 h 0 m	12.6	
8	DAY	16 h 6 m	21.6	18 h 18 m	13.6	16 h 6 m	13.6	18 h 18 m	21.6	
(16-31 July)	NIGTH	7 h 54 m	12.6	5 h 42 m	9.5	7 h 54 m	9.5	5 h 42 m	12.6	
9	DAY	15 h 24 m	21.5	17 h 6 m	13.6	15 h 24 m	13.6	17 h 6 m	21.5	
(1-15 August)	NIGTH	8 h 36 m	12.5	6 h 54 m	9.4	8 h 36 m	9.4	6 h 54 m	12.5	
10	DAY	14 h 36 m	20.1	15 h 48 m	12.4	14 h 36 m	12.4	15 h 48 m	20.1	
(16-31 August)	NIGTH	9 h 24 m	11.4	8 h 12 m	8.4	9 h 24 m	8.4	8 h 12 m	11.4	
11	DAY	13 h 36 m	18.7	14 h 18 m	11.2	13 h 36 m	11.2	14 h 18 m	18.7	
(1-15 September)	NIGTH	10 h 24 m	10.2	9 h 42 m	7.3	10 h 24 m	7.3	9 h 42 m	10.2	
12	DAY	12 h 42 m	16.7	12 h 54 m	9.9	12 h 42 m	9.9	12 h 54 m	16.7	
(16-30 September)	NIGTH	11 h 18 m	8.7	11 h 6 m	6.1	11 h 18 m	6.1	11 h 6 m	8.7	

8 humidity of 70% and a luminosity of 400 μ mol·m⁻²·s⁻¹.

10 Table S3. Source populations for the reciprocal transplants experiment. Species codes stand

11 for *M. guttatus* (G) and *M.* \times *robertsii* (R). The origin of populations is classified as sampled

Population	Species	Origin	Lat.	Long.	m a.s.l.	N (individuals)
CRO	G	South	50.16293	-5.29331	129	2
EAS	G	South	50.216213	-3.713068	129	2
DAR	G	South	50.329354	-3.574899	69	3
MOO	G	South	50.45142	-4.486005	54	3
TCO	G	South	50.49812	-4.465601	227	1
SOU	G	South	50.6016	-3.767733	259	4
BOG	G	South	50.797265	-0.698253	6	11
HUN	G	South	50.810705	-0.788876	7	3
FUN	G	South	50.862581	-0.855275	20	4
DEA	G	South	50.904513	-0.779717	50	4
SIN	G	South	50.911711	-0.753315	60	4
UPL	G	South	50.938462	-3.412896	127	4
TOU	G	South	51.074453	-3.123822	124	4
HOU	G	South	51.09699	-1.5084	33	3
MOR	R	South	50.163817	-5.656375	29	5
DRI	R	South	51.148021	-3.808392	384	6
MAR	G	North	57.572341	-4.427486	3	5
GAR	G	North	57.615064	-4.673473	75	4
DAL	G	North	57.682614	-4.265258	6	4
BLA	G	North	58.48755	-5.10636	44	4
BKN	G	North	58.5759	-4.76774	8	4
BOD	G	North	59.90418	-1.30274	55	11
NIN	G	North	59.97777	-1.30036	87	5
WEI	G	North	60.254393	-1.289859	6	4
MUK	G	North	60.34808	-1.41373	8	4
HAM	G	North	60.5034	-1.09931	4	4
NORG	G	North	60.808743	-0.807753	13	5
GJO	G	North	62.32533	-6.94162	3	2
CLS	R	North	58.215153	-5.33411	46	7
GIO	R	North	58.33754	-6.20187	38	2
POL	R	North	58.483497	-5.099521	20	11
EOR	R	North	58.49937	-6.26996	8	3
STR	R	North	58.9692	-3.28341	9	3
TOR	R	North	58.9957	-3.18338	26	2
EVI	R	North	59.11226	-3.10809	37	1
NOR	R	North	60.808743	-0.807753	13	7
COL	L	North	55.6550	-2.2401	9	25

12 in the south or north of the British Isles.

Table S4. Summary of results of the GLMMs modelling the variation in all traits measured in the controlled environmental chambers as a function of

15 temperature, photoperiod, population type (where appropriate) and their interactions.

			All populations		Native <i>M. guttati</i>	us	Invasive <i>M. gutta</i>	tus	M. × robertsii	
Trait	Fixed factor		Estimate (SE)	γ^2	Estimate (SE)	γ^2	Estimate (SE)	γ^2	Estimate (SE)	γ^2
Germination				λ.		λ.		λ		λ
day	Intercept		13.475 (0.498)		13.2 (0.353)		13.41 (0.65)			
	Т	(Warm)	-6.087 (0.25)	590.496 ***	-6.2 (0.408)	231.463 ***	-6.02 (0.314)	367.683 ***		
	Р	(Short)	0.237 (0.25)	0.899	0.133 (0.408)	0.107	0.3 (0.314)	0.913		
	0	(US)	-0.366 (0.707)	0.268						
Flower day	Intercept		91.183 (2.248)		87.645 (1.589)		90.637 (1.906)		81.767 (4.25)	
	Т	(Warm)	-22.014 (1.279)	296.237 ***	-24.022 (1.504)	255.145 ***	-22.294 (1.111)	402.322 ***	-17.892 (4.976)	12.93 ***
	Р	(Short)	8.52 (1.297)	43.156 ***	9.949 (1.504)	43.763 ***	10.018 (1.114)	80.876 ***	0.443 (5.221)	0.007
	0	(US)	-3.838 (2.986)	8.298 *						
		(rob)	-9.103 (3.16)							
Flowered	Intercept		1.031 (0.052)		1.001 (0.032)		1.00 (0.044)		0.972 (0.099)	
	Т	(Warm)	-0.015 (0.028)	0.289	-0.013 (0.03)	0.194	-0.02 (0.033)	0.356	-0.01 (0.079)	0.016
	Р	(Short)	-0.167 (0.028)	33.538 ***	-0.042 (0.03)	2.055	-0.1 (0.033)	8.911 **	-0.423 (0.079)	28.559 ***
	0	(US)	0.032 (0.07)	10.339 **						
		(rob)	-0.186 (0.071)							
Corolla	Intercept		38.972 (1.42)		39.274 (1.258)		38.888 (2.016)		35.9 (1.09)	
	Т	(Warm)	-2.637 (0.496)	28.25 ***	-2.693 (1.05)	6.571	-3.298 (0.589)	31.361 ***	-1.19 (1.015)	1.375
	Р	(Short)	-1.254 (0.502)	6.233 *	-2.684 (1.05)	6.528	-0.403 (0.59)	0.467	-0.899 (1.061)	0.717
	0	(US)	-0.479 (1.963)	1.43	· /					
		(rob)	-2.3 (2.01)							

Flowers	Intercept		3.539 (0.118)		3.675 (0.13)		3.54 (0.079)		2.037 (0.178)	
Flowers	Т	(Warm)	0.128 (0.046)	7.713 **	0.315 (0.049)	41.564 ***	0.129 (0.046)	7.712	0.199 (0.102)	3.797
	P	(Warin) (Short)	-0.87 (0.063)	186.062 ***	-0.496 (0.061)	67.063 ***	-0.87 (0.064)	185.731 ***	-0.654 (0.117)	31.037 ***
	Т Т:Р	(Warm:Short)	0.313 (0.082)	14.412 ***	0.348 (0.077)	20.612 ***	0.313 (0.083)	14.343 ***	-0.004 (0.117)	51.057
	0	(US)	0.136 (0.168)	81.676 ***	0.548 (0.077)	20.012	0.313 (0.083)	14.343		
	0	(US) (rob)	-1.462 (0.189)	01.070						
	O : T		0.186 (0.067)	7.883 *						
	0:1	(US:Warm) (rob:Warm)	0.188 (0.087) 0.027 (0.127)	1.003						
	0 · P	· /		10 205 ***						
	O : P	(US:Short)	0.375 (0.087)	18.305 ***						
		(rob:Short)	0.141 (0.184)	0.617						
	O : T : P	(US:W:S)	0.033 (0.112)	0.01/						
		(rob:W:S)	-0.154 (0.245)							
Stems	Intercept		1.436 (0.095)		1.786 (0.115)		1.62 (0.092)		0.782 (0.179)	
Stems	T	(Warm)	-0.089 (0.072)	1.532	0.072 (0.104)	0.474	-0.287 (0.114)	6.323	-0.061 (0.211)	0.084
	P	(Warm) (Short)	-0.571 (0.076)	56.545 ***	-0.373 (0.106)	12.357 ***	-0.847 (0.125)	46.294 ***	-0.476 (0.238)	4.002
	0	(US)	0.506 (0.118)	57.637 ***	-0.575 (0.100)	12.337	-0.047 (0.123)	40.274	-0.470 (0.250)	4.002
	U	(US) (rob)	-0.612 (0.151)	37.037						
		(100)	-0.012 (0.131)							
Height	Intercept		29.5 (2.383)		33.459 (2.017)		31.445 (2.866)		20.969 (2.09)	
	Т	(Warm)	23.48 (2.132)	121.317 ***	15.683 (1.888)	68.976 ***	19.59 (1.605)	148.986 ***	7.217 (1.775)	16.535 ***
	Р	(Short)	-2.86 (2.132)	1.8	-4.905 (1.888)	6.747	-6.75 (1.605)	17.688 ***	-10.531 (1.782)	34.93 ***
	T : P	(Warm:Short)	-7.78 (3.015)	6.66 **						
	0	(US)	3.38 (3.501)	13.928 ***						
		(rob)	-10.076 (3.614)							
	O : T	(US:Warm)	-6.839 (3.325)	14.855 ***						
		(rob:Warm)	-13.247 (3.481)							
	O : P	(US:Short)	-1.088 (3.325)	1.847						
		(rob:Short)	-4.677 (3.486)							
	O : T : P	(US:W:S)	5.865 (4.704)	1.575						
	-	(rob:W:S)	1.747 (4.923)							
		· · · · · · · · · · · · · · · · · · ·	· · · /							

Branches	Intercept T P O	(Warm) (Short) (US) (rob)	0.46 (0.196) 0.322 (0.107) 0.128 (0.105) -0.275 (0.262) -1.521 (0.313)	9.094 ** 1.474 24.271 ***	0.081 (0.32) 0.339 (0.182) 0.203 (0.18)	3.487 1.271	0.414 (0.138) 0.373 (0.142) 0.195 (0.14)	6.932 1.936	-0.429 (0.323) -0.154 (0.392) -0.799 (0.425)	0.154 3.534
Internode length	Intercept T P O	(Warm) (Short) (US) (rob)	14.128 (0.971) 2.197 (0.727) -3.833 (0.727) -1.503 (1.224) -2.07 (1.269)	9.134 ** 27.761 *** 2.988	13.671 (1.021) 0.498 (1.19) -4.081 (1.19)	0.175 11.753 ***	14.492 (1.004) 3.322 (0.912) -5.686 (0.912)	13.273 *** 38.884 ***	10.283 (1.787) 2.315 (1.773) -0.402 (1.778)	1.704 0.051
Internode diameter	Intercept T P O	(Warm) (Short) (US) (rob)	7.234 (0.212) -0.638 (0.145) -0.654 (0.145) -0.562 (0.272) -2.934 (0.281)	19.311 *** 20.268 *** 116.882 ***	6.808 (0.33) -0.786 (0.294) -0.798 (0.294)	7.1424 7.357	7.382 (0.203) -0.853 (0.221) -0.733 (0.221)	14.869 *** 10.98 ***	3.908 (0.26) -0.103 (0.214) -0.354 (0.215)	0.232 2.712
Drymass	Intercept T P O	(Warm) (Short) (US) (rob)	5.199 (0.528) -0.545 (0.173) -0.606 (0.173) -0.596 (0.732) -0.137 (0.739)	9.891 ** 12.206 *** 0.72	4.28 (0.299) -0.122 (0.206) -0.373 (0.206)	0.349 3.259	5.36 (0.214) -0.7 (0.163) -0.774 (0.163)	18.521 *** 22.67 ***	5.185 (0.959) -0.787 (0.549) -0.609 (0.55)	2.054 1.226
Stolons	Intercept T P T : P O	(Warm) (Short) (Warm:Short) (US)	1.744 (0.209) 0.319 (0.108) 0.231 (0.11) -0.369 (0.151) -0.468 (0.308)	8.67 ** 4.369 * 5.972 * 17.974 ***	1.26 (0.332) 0.872 (0.138) 0.332 (0.154) -0.756 (0.196)	39.749 *** 4.644 14.909 ***	1.75 (0.112) 0.32 (0.109) 0.231 (0.111) -0.37 (0.151)	8.6635 4.3664 5.9679	2.616 (0.126) -0.234 (0.073) -0.005 (0.073)	10.222 0.005

	(rob)	0.801 (0.294)	
O : T	(US:Warm)	0.549 (0.175)	32.035 ***
	(rob:Warm)	-0.418 (0.148)	
O : P	(US:Short)	0.096 (0.189)	1.452
	(rob:Short)	-0.111 (0.147)	
O : T : P	(US:W:S)	-0.379 (0.247)	
	(rob:W:S)	0.092 (0.21)	3.881

- **Table note:** Fixed effect abreviations: Origin (O); Temperature (T); Photoperiod (P). χ2 values and their associated *P*-values from type-III Wald χ2
- 20 tests (or type-II Wald χ 2 tests, when all interaction terms were unsignificant and removed) are provided. * P<0.05; ** P<0.01; *** P<0.001.
- 21 Significant effects after Bonferroni correction of *P*-values are indicated in bold.

Table S5. Mean value, standard deviation and sample size for each trait and group defined by
a significant fixed effect in the GLMMs modelling the phenotypic data for all population

26 types in the controlled environmental chambers.
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Trait	Fixed effect	Group	mean ± sd (N)
Germination day	Т	Cold	13.45 ± 2.349 (80)
		Warm	7.363 ± 1.343 (80)
Flower day	Т	Cold	91.461 ± 11.811 (103)
		Warm	69.622 ± 11.311 (107)
	Р	Long	76.662 ± 14.635 (114)
		Short	84.693 ± 16.305 (96)
Flowered	Р	Long	0.987 ± 0.08 (115)
		Short	0.819 ± 0.351 (116)
Corolla	Т	Cold	38 ± 5.289 (106)
		Warm	35.275 ± 4.61 (106)
Flowers	0	M. guttatus invasive	28.332 ± 16.368 (98) b
		M. guttatus native	43.333 ± 22.34 (69) c
		M. x robertsii	7.929 ± 6.009 (49) a
	Р	Long	33.622 ± 23.02 (115)
		Short	22.658 ± 17.523 (101)
	O:P	M. guttatus invasive : Long	37.33 ± 16.892 (50) de
		<i>M. guttatus</i> invasive : Short	18.958 ± 8.923 (48) c
		<i>M. guttatus</i> native : Long	$49 \pm 23.446 (35) e$
		M. guttatus native : Short	37.5 ± 19.821 (34) d
		<i>M. x robertsii</i> : Long	9.5 ± 6.745 (30) b
		<i>M. x robertsii</i> : Short	5.447 ± 3.515 (19) a
	T:P	Cold : Long	30.034 ± 18.115 (58) c
		Cold : Short	16.48 ± 11.116 (50) a
		Warm : Long	37.272 ± 26.793 (57) d
		Warm : Short	$28.716 \pm 20.43 \ (51) \ b$
Stems	0	M. guttatus invasive	3.179 ± 2.179 (98) b
		M. guttatus native	5.239 ± 2.182 (69) c
		M. x robertsii	1.816 ± 1.121 (49) a
Height	0	M. guttatus invasive	$37.865 \pm 14.325 \ (100) \ b$
		M. guttatus native	39.186 ± 11.725 (70) b
		M. x robertsii	19.475 ± 9.861 (60) a
	Т	Cold	$25.904 \pm 9.904 \ (115)$

		Warm	41.035 ± 15.436 (115)
	Р	Long Short	4.396 ± 2.486 (115) 2.54 ± 1.767 (101)
	O:T	M. guttatus invasive : Cold M. guttatus invasive : Warm M. guttatus native : Cold M. guttatus native : Warm M. x robertsii : Cold M. x robertsii : Warm	$28.07 \pm 7.465 (50) \text{ b}$ $47.66 \pm 12.769 (50) \text{ c}$ $31.414 \pm 7.832 (35) \text{ b}$ $46.957 \pm 9.66 (35) \text{ c}$ $15.867 \pm 8.398 (30) \text{ a}$ $23.083 \pm 10.017 (30) \text{ b}$
Branches	0	<i>M. guttatus</i> invasive <i>M. guttatus</i> native <i>M. x robertsii</i>	$2.01 \pm 1.409 (100) b$ $1.779 \pm 1.634 (70) b$ $0.458 \pm 0.825 (60) a$
	Т	Cold Warm	$\begin{array}{c} 1.27 \pm 1.372 \ (115) \\ 1.8 \pm 1.585 \ (115) \end{array}$
Internode	Т	Cold Warm	$\begin{array}{c} 11.259 \pm 6.199 \ (115) \\ 13.423 \pm 5.949 \ (115) \end{array}$
	Р	Long Short	$14.233 \pm 6.136 (115) \\ 10.45 \pm 5.595 (115)$
Diameter	Τ	Cold Warm	$5.976 \pm 1.791 (115) \\ 5.332 \pm 1.651 (115)$
	Р	Long Short	$5.982 \pm 1.741 (115) \\ 5.325 \pm 1.701 (115)$
	0	<i>M. guttatus</i> invasive <i>M. guttatus</i> native <i>M. x robertsii</i>	$6.588 \pm 1.256 (100) b$ $6.068 \pm 1.443 (70) b$ $3.613 \pm 0.931 (60) a$
Drymass	Т	Cold Warm	$\begin{array}{l} 4.704 \pm 1.619 \ (115) \\ 4.142 \pm 1.797 \ (116) \end{array}$
	Р	Long Short	4.737 ± 1.701 (115) 4.11 ± 1.708 (116)
Stolons	0	M. guttatus invasive M. guttatus native M. x robertsii	7.025 ± 3.018 (100) ab 6.407 ± 7.506 (70) a 12.558 ± 5.899 (60) b
	Т	Cold Warm	$\begin{array}{c} 7.996 \pm 6.203 \; (115) \\ 8.565 \pm 5.878 \; (115) \end{array}$
	O: T	<i>M. guttatus</i> invasive : Cold <i>M. guttatus</i> invasive : Warm	6.56 ± 2.697 (50) abc 7.49 ± 3.269 (50) abc

	M. guttatus native : Cold	4.871 ± 6.543 (35) a
	M. guttatus native: Warm	7.943 ± 8.165 (35) bc
	M. x robertsii : Cold	14.033 ± 5.975 (30) c
	<i>M. x robertsii</i> : Warm	11.083 ± 5.531 (30) b
7		

- 28 **Table note:** Fixed effect abreviations: Origin (O); Temperature (T); Photoperiod (P). When
- 29 more than two groups are defined by the significant fixed term, significantly different groups
- 30 according to *post hoc* tests are indicated with different letters.

31 Table S6. Results of the GLMMs analysing the RDPI indexes of phenotypic plasticity as a

32 function of the population type for all traits measured in the controlled environmental

33 chambers.

(a) RDPIt	χ^2	df	Р	M. guttatus native	M. guttatus invasive	M. x robertsii
Branches	8.679	2	0.013 *	$0.636 \pm 0.304 \ (18)$	$0.469 \pm 0.297 \ (25)$	0.821 ± 0.244 (13)
Flower day	1.015	2	0.602	$0.26 \pm 0.042 \; (18)$	$0.23 \pm 0.055 \ (25)$	0.239 ± 0.185 (13)
Germination day	2.106	1	0.147	$0.466 \pm 0.035 \; (15)$	$0.441 \pm 0.064 \ (25)$	
Stems	5.552	2	0.062	0.227 ± 0.171 (18)	$0.395 \pm 0.222 \ (25)$	$0.214 \pm 0.202 \; (15)$
Corolla width	0.15	2	0.928	$0.093 \pm 0.056 \; (18)$	$0.091 \pm 0.054 \ (25)$	$0.089 \pm 0.054 \; (13)$
Height	5.064	2	0.08	0.365 ± 0.117 (18)	$0.404 \pm 0.098 \ (25)$	0.307 ± 0.202 (16)
Diameter	0.872	2	0.647	0.169 ± 0.143 (18)	$0.139 \pm 0.104 \ (25)$	0.165 ± 0.114 (16)
Flowers	5.02	2	0.081	0.428 ± 0.156 (18)	0.296 ± 0.19 (25)	0.431 ± 0.24 (15)
Internode	3.256	2	0.196	0.32 ± 0.109 (18)	0.273 ± 0.138 (25)	0.353 ± 0.183 (16)
Drymass	3.468	2	0.177	0.159 ± 0.118 (18)	0.175 ± 0.132 (25)	0.247 ± 0.201 (16)
Stolons	4.321	2	0.115	0.426 ± 0.198 (18)	0.273 ± 0.178 (25)	0.317 ± 0.242 (16)
(b) RDPIp	Chisq	df	Р	M. guttatus snative	M. guttatus invasive	M. x robertsii
Branches	18.272	2	< 0.001 ***	0.46 ± 0.355 (18) a	0.341 ± 0.188 (25) a	0.797 ± 0.366 (13) b
Flower day	23.334	2	< 0.001 ***	0.128 ± 0.054 (18) a	0.109 ± 0.061 (25) a	0.237 ± 0.123 (10) b
Germination day	0.007	1	0.933	$0.055\pm 0.054\;(15)$	$0.056 \pm 0.045 \ (25)$	
Stems	16.526	2	< 0.001 ***	0.335 ± 0.174 (18) a	0.575 ± 0.204 (25) b	0.445 ± 0.219 (11) at
Corolla width	2.77	2	0.25	$0.085 \pm 0.064 \; (18)$	$0.068 \pm 0.047~(25)$	0.055 ± 0.035 (11)
Height	15.837	2	< 0.001 ***	0.169 ± 0.135 (18) a	0.199 ± 0.159 (25) a	0.426 ± 0.199 (16) b
Diameter	0.214	2	0.899	0.158 ± 0.116 (18)	$0.145 \pm 0.101 \; (25)$	0.141 ± 0.112 (16)
Flowers	6.201	2	0.045 *	0.315 ± 0.18 (18)	0.471 ± 0.199 (25)	0.484 ± 0.216 (11)
Internode	0.405	2	0.817	0.31 ± 0.137 (18)	0.343 ± 0.176 (25)	0.317 ± 0.22 (16)
Drymass	7.342	2	0.025 *	0.162 ± 0.102 (18)	0.164 ± 0.09 (25)	0.283 ± 0.215 (16)
Stolons	1.342	2	0.511	0.303 ± 0.19 (18)	0.289 ± 0.139 (25)	0.243 ± 0.16 (16)
(c) RDPItp	Chisq	df	Р	M. guttatus native	M. guttatus invasive	M. x robertsii
Branches	1.453	2	0.484	$0.527 \pm 0.209 \ (13)$	$0.543 \pm 0.176 \ (24)$	$0.685\pm 0.085\;(13)$
Flower day	5.123	2	0.077	0.209 ± 0.028 (16)	$0.198 \pm 0.042 \ (21)$	0.247 ± 0.095 (6)
Germination day	0.893	1	0.345	$0.332\pm 0.037\ (15)$	$0.319 \pm 0.044 \ (25)$	
Stems	7.756	2	0.021 *	0.342 ± 0.153 (16)	$0.498 \pm 0.092 \; (23)$	$0.364 \pm 0.211 \ (7)$
Corolla width	2.069	2	0.355	0.105 ± 0.038 (16)	$0.095 \pm 0.029~(22)$	$0.078 \pm 0.056 \ (6)$
Height	4.747	2	0.093	$0.293 \pm 0.099 \ (17)$	$0.34 \pm 0.07 \; (25)$	0.389 ± 0.172 (12)
Diameter	1.203	2	0.548	0.21 ± 0.097 (17)	$0.18 \pm 0.074~(25)$	$0.202 \pm 0.104 \; (12)$
Flowers	0.646	2	0.724	0.417 ± 0.097 (16)	0.431 ± 0.11 (23)	0.456 ± 0.14 (7)
Internode	0.169	2	0.919	0.347 ± 0.102 (17)	0.349 ± 0.104 (25)	0.371 ± 0.184 (12)
Drymass	13.655	2	0.001 **	0.207 ± 0.068 (17) a	0.215 ± 0.091 (25) a	0.319 ± 0.126 (13) b
Stolons	5.034	2	0.081	0.43 ± 0.114 (17)	0.34 ± 0.131 (25)	0.327 ± 0.172 (12)

Table note: Mean \pm s.d. (N) is given for each trait and group, and significant differences in

the *post hoc* tests are labelled with different letters. * P<0.05; ** P<0.01; *** P<0.001.

Table S7. Results of the GLMMs modelling the effects of initial weigth (W₀), population

37 origin, experimental site and their interaction in 10 phenotypic and phenological traits

recorded in a reciprocal transplant experiment with introduced *Mimulus guttatus* and M. ×

39 *robertsii* populations.

		M. guttatus		M. × robertsii	
Trait	Fixed factor	Estimate (SE)	χ^2	Estimate (SE)	χ^2
	T				
Flowering	Intercept	49.206 (2.227)	488.243 ***	37.919 (2.272)	278.455 ***
day	\mathbf{W}_0	-0.053 (0.021)	6.522 *	-0.007 (0.006)	1.202
	Origin (South)	-0.71 (1.691)	0.176	5.622 (4.715)	1.422
	Site (SHE)	24.121 (1.017)	562.793 ***	25.529 (1.354)	355.649 ***
	S:O (SHE:South)	-0.66 (1.271)	0.269	-5.415 (2.739)	3.908 *
Flowering	Intercept	5.882 (0.433)	184.892 ***	4.536 (0.265)	292.421 ***
node	\mathbf{W}_0	0.002 (0.003)	0.219	0 (0.001)	0.212
	Origin (South)	0.031 (0.415)	0.006	0.256 (0.53)	0.234
	Site (SHE)	-0.103 (0.168)	0.376	0.132 (0.172)	0.59
	S:O (SHE:South)	-0.053 (0.208)	0.064	-0.303 (0.344)	0.774
Branches	Intercept	2.394 (0.116)	422.442 ***	2.158 (0.079)	746.99 ***
Diancies	\mathbf{W}_0	-0.002 (0.001)	5.303 *	0.001 (0)	4.544 *
	Origin (South)	0.142 (0.093)	2.307	0 (0.15)	0
	Site (SHE)	0.011 (0.051)	0.044	-0.073 (0.057)	1.682
	S:O (SHE:South)	-0.197 (0.062)	10.14 **	0.207 (0.11)	3.514
Stems	Intercept	2.844 (0.111)	655.131 ***	2.635 (0.081)	1053.154 ***
	\mathbf{W}_0	-0.002 (0.001)	4.087 *	0.001 (0)	12.247 ***
	Origin (South)	0.095 (0.115)	0.679	0.014 (0.157)	0.008
	Site (SHE)	-0.046 (0.04)	1.353	-0.13 (0.045)	8.4 **
	S:O (SHE:South)	-0.153 (0.048)	9.933 **	0.146 (0.087)	2.822
Flowers	Intercept	4.915 (0.105)	2179.201 ***	5.093 (0.076)	4455.005 ***
	\mathbf{W}_0	-0.001 (0)	11.943 ***	0 (0)	42.282 ***
	Origin (South)	0.196 (0.142)	1.909	-0.017 (0.159)	0.012
	Site (SHE)	-0.181 (0.015)	147.487 ***	-0.146 (0.014)	116.622 ***
	S:O (SHE:South)	-0.101 (0.018)	32.664 ***	0.105 (0.027)	15.11 ***
Corolla	Intercept	399.725 (18.206)	482.072 ***	422.987 (15.085)	786.301 ***
CUIVIIA	\mathbf{W}_0	-0.031 (0.165)	0.036	-0.075 (0.042)	3.141
	Origin (South)	-38.025 (14.839)	6.566 * †	-44.584 (30.787)	2.097
	Site (SHE)	-38.025 (14.839) -14.039 (8.091)	3.011	-44.384 (30.787) 13.015 (9.624)	2.097 1.829
	Site (SHE) S:O (SHE:South)	-4.286 (10.07)	0.181	-41.657 (19.368)	1.829 4.626 *
	5.0 (SITE.SOUUI)	-4.200 (10.07)	0.101	-+1.057 (19.308)	4.020

Stomata	Intercept	18.137 (1.03)	310.282 ***	14.336 (1.095)	171.362 ***
	W ₀	-0.034 (0.01)	10.598 **	0.001 (0.002)	0.285
	Origin (South)	-0.459 (0.656)	0.489	-1.779 (2.334)	0.581
	Site (SHE)	-2.555 (0.516)	24.524 ***	-0.195 (0.525)	0.138
	S:O (SHE:South)	-0.171 (0.651)	0.069	0.25 (1.059)	0.056
Height	Intercept	19.816 (2.646)	56.103 ***	22.081 (1.348)	268.363 ***
	W ₀	0.035 (0.024)	2.142	0.02 (0.007)	9.728 **
	Origin (South)	2.904 (2.153)	1.82	0.424 (2.358)	0.032
	Site (SHE)	-3.204 (1.167)	7.544 **	-0.402 (1.411)	0.081
	S:O (SHE:South)	2.763 (1.453)	3.616	2.26 (2.852)	0.628
Cover	Intercept	906.806 (74.199)	149.36 ***	640.669 (52.4)	149.488 ***
	W ₀	-2.091 (0.667)	9.847 **	0.41 (0.172)	5.688 *
	Origin (South)	8.063 (62.149)	0.017	48.324 (104.377)	0.214
	Site (SHE)	-267.054 (32.495)	67.541 ***	-59.204 (38.272)	2.393
	S:O (SHE:South)	-14.122 (40.777)	0.12	-46.233 (77.581)	0.355
Dry mass	Intercept	37.479 (4.929)	57.814 ***	35.644 (3.015)	139.769 ***
	W ₀	0.013 (0.046)	0.075	0.014 (0.009)	2.567
	Origin (South)	-0.511 (3.793)	0.018	1.745 (6.092)	0.082
	Site (SHE)	-15.924 (2.25)	50.089 ***	-9.85 (2.029)	23.579 ***
	S:O (SHE:South)	0.676 (2.804)	0.058	-2.334 (4.106)	0.323

Table note: Models were built on each individual species dataset. Maximal model Estimates and Standard Errors for each fixed factor and interaction are provided jointly with results of the type-III Wald $\chi 2$ tests. $\chi 2$ values and indications of their associated *P*-values are provided. P < 0.05; ** P < 0.01; *** P < 0.001. Significant effects after Bonferroni correction of *P*values are indicated in bold. $\chi 2$ degrees of freedom=1. † Indicates significant main effects after type-II Wald $\chi 2$ tests on GLMMs excluding non-significant interactions (only significant effects after Bonferroni correction are shown). **Table S8.** Results of the regression models of phenotypic selection on 10 traits measured in *Mimulus guttatus*, *M.* × *robertsii* and *M. luteus*

49 populations included in the reciprocal transplants experiment.

(a) M. guttatus		Fru	its			Sto	lons	
	IOW		SHE		IOW		SHE	
	Estimate (SE)	t						
Intercept	0.868 (0.029)	29.955 ***	0.82 (0.039)	21.273 ***	0.859 (0.037)	22.951 ***	0.849 (0.044)	19.496 ***
Height	-0.392 (0.16)	-2.453 *	-0.071 (0.203)	-0.348	-0.084 (0.207)	-0.405	0.164 (0.237)	0.693
Stomata	-0.024 (0.189)	-0.127	-0.11 (0.252)	-0.435	0.104 (0.244)	0.428	-0.129 (0.291)	-0.441
Flowering day	0.57 (0.217)	2.631 **	0.589 (0.482)	1.223	-0.029 (0.277)	-0.106	-0.301 (0.459)	-0.656
Flowering node	0.293 (0.199)	1.471	-0.01 (0.256)	-0.04	0.302 (0.259)	1.168	0.26 (0.244)	1.065
Corolla	0.187 (0.292)	0.64	0.701 (0.28)	2.503 *	-0.326 (0.379)	-0.86	-0.053 (0.327)	-0.162
Plant cover	-0.056 (0.155)	-0.363	0.178 (0.238)	0.748	0.029 (0.199)	0.147	0.445 (0.275)	1.617
Branches	-0.175 (0.143)	-1.223			-0.223 (0.186)	-1.198		
Floral stems	0.27 (0.152)	1.781			0.314 (0.196)	1.603		
Flowers	-0.134 (0.126)	-1.062	0.078 (0.129)	0.602	-0.277 (0.163)	-1.697	-0.203 (0.141)	-1.434
Dry mass	0.412 (0.107)	3.864 ***	0.106 (0.188)	0.563	0.142 (0.137)	1.033	-0.463 (0.219)	-2.109 *
Height ²	0.743 (0.309)	2.407 *	-0.015 (0.189)	-0.08	0.121 (0.401)	0.301	-0.107 (0.223)	-0.477
Stomata^2	0.019 (0.38)	0.051	0.111 (0.256)	0.435	-0.118 (0.491)	-0.241	0.151 (0.294)	0.515
Flowering day [^] 2	-1.328 (0.447)	-2.97 **	-0.704 (0.505)	-1.394	0.106 (0.567)	0.187	0.406 (0.464)	0.875
Flowering node [^] 2	-0.591 (0.394)	-1.501	-0.093 (0.249)	-0.372	-0.613 (0.511)	-1.201	-0.229 (0.238)	-0.963
Corolla^2	-0.333 (0.589)	-0.566	-0.582 (0.277)	-2.098 *	0.651 (0.764)	0.852	0.124 (0.327)	0.38
Plant cover^2	0.248 (0.284)	0.874	0.02 (0.199)	0.1	0.057 (0.366)	0.155	-0.306 (0.23)	-1.331
Branches^2	0.361 (0.287)	1.258			0.443 (0.372)	1.19		
Floral stems ²	-0.497 (0.297)	-1.675			-0.482 (0.384)	-1.256		
Flowers ²	0.342 (0.231)	1.482	-0.019 (0.143)	-0.134	0.438 (0.298)	1.47	0.254 (0.151)	1.683
Dry mass ²	-0.649 (0.192)	-3.376 ***	-0.066 (0.155)	-0.424	-0.265 (0.248)	-1.068	0.373 (0.184)	2.024 *

(b) M. x robertsii	Stolons						
	ΙΟ₩	,	SHE				
	Estimate (SE)	t	Estimate (SE)	t			
Intercept	0.787 (0.058)	13.506 ***	0.837 (0.057)	14.755 ***			
Height	0.164 (0.314)	0.522	-0.238 (0.289)	-0.825			
Stomata	-0.264 (0.362)	-0.73	-0.505 (0.353)	-1.429			
Flowering day	-0.229 (0.383)	-0.598	0.585 (0.728)	0.803			
Flowering node	-0.342 (0.389)	-0.879	-0.406 (0.331)	-1.229			
Corolla	0.725 (0.548)	1.324	0.835 (0.31)	2.698 **			
Plant cover	0.164 (0.381)	0.429	0.217 (0.477)	0.455			
Branches	0.236 (0.219)	1.079					
Floral stems	0.24 (0.265)	0.907					
Flowers	-0.287 (0.303)	-0.949	0.01 (0.296)	0.033			
Dry mass	0.615 (0.297)	2.073 *	-0.39 (0.413)	-0.944			
Height^2	-0.241 (0.595)	-0.405	0.085 (0.295)	0.29			
Stomata^2	0.532 (0.71)	0.75	0.528 (0.353)	1.497			
Flowering day ²	0.74 (0.788)	0.939	-0.438 (0.724)	-0.605			
Flowering node ²	0.293 (0.779)	0.376	0.357 (0.323)	1.104			
Corolla^2	-1.354 (1.09)	-1.242	-0.777 (0.309)	-2.519 *			
Plant cover^2	-0.594 (0.635)	-0.936	-0.057 (0.409)	-0.139			
Branches^2	-0.286 (0.352)	-0.812					
Floral stems ²	-0.801 (0.483)	-1.659					
Flowers [^] 2	1.071 (0.571)	1.877	-0.007 (0.282)	-0.026			
Dry mass^2	-0.758 (0.504)	-1.504	0.219 (0.34)	0.644			

(c) M. luteus	Fruits				Stolons				
	IOW		SHE	SHE		IOW		SHE	
	Estimate (SE)	t	Estimate (SE)	t	Estimate (SE)	t	Estimate (SE)	t	
Intercept	0.941 (0.047)	19.872 ***	0.83 (0.074)	11.272 ***	0.838 (0.055)	15.201 ***	0.817 (0.106)	7.716 ***	
Height	-0.158 (0.486)	-0.325	0.581 (0.545)	1.066	0.987 (0.565)	1.747	-0.12 (0.828)	-0.145	
Stomata	-0.788 (0.39)	-2.023	-0.113 (0.549)	-0.206	1.012 (0.453)	2.232 *	-0.63 (0.842)	-0.748	
Flowering day	-0.11 (0.388)	-0.284	-0.604 (0.955)	-0.632	0.576 (0.451)	1.275	1.831 (1.279)	1.432	
Flowering node	-0.381 (0.333)	-1.147	0.136 (0.415)	0.327	0.336 (0.387)	0.868	-0.344 (0.631)	-0.544	
Corolla	1.34 (0.666)	2.013	0.577 (0.976)	0.591	-1.249 (0.775)	-1.613	-0.525 (1.414)	-0.371	
Plant cover	-0.529 (0.347)	-1.526	-0.118 (0.406)	-0.292	0.08 (0.404)	0.197	0.81 (0.611)	1.326	
Branches	0.524 (0.319)	1.641			-0.422 (0.372)	-1.137			
Floral stems	-0.337 (0.361)	-0.933			-0.286 (0.42)	-0.682			
Flowers	0.04 (0.318)	0.127	0.518 (0.28)	1.851	0.088 (0.37)	0.239	0.342 (0.394)	0.869	
Dry mass	0.86 (0.307)	2.798 **	-0.028 (0.378)	-0.073	-0.492 (0.358)	-1.375	-0.644 (0.57)	-1.131	
Height ²	0.203 (0.936)	0.216	-0.402 (0.521)	-0.771	-1.681 (1.09)	-1.543	0.136 (0.794)	0.172	
Stomata^2	1.592 (0.729)	2.185 *	0.123 (0.544)	0.226	-2.07 (0.848)	-2.442 *	0.567 (0.831)	0.683	
Flowering day [^] 2	0.042 (0.618)	0.067	0.763 (0.977)	0.781	-0.703 (0.72)	-0.978	-1.871 (1.289)	-1.452	
Flowering node [^] 2	0.371 (0.648)	0.573	-0.177 (0.395)	-0.448	-0.923 (0.754)	-1.224	0.512 (0.595)	0.859	
Corolla^2	-2.294 (1.278)	-1.794	-0.571 (0.981)	-0.582	2.142 (1.488)	1.44	0.519 (1.414)	0.367	
Plant cover^2	1.001 (0.648)	1.545	0.314 (0.352)	0.892	0.092 (0.754)	0.122	-0.915 (0.534)	-1.713	
Branches ²	-1.483 (0.684)	-2.169 *			0.744 (0.795)	0.935			
Floral stems^2	1.199 (0.696)	1.723			0.524 (0.81)	0.647			
Flowers^2	0.042 (0.652)	0.065	-0.367 (0.253)	-1.449	-0.264 (0.758)	-0.348	-0.157 (0.373)	-0.42	
Dry mass^2	-1.484 (0.573)	-2.59 *	-0.024 (0.371)	-0.065	0.692 (0.667)	1.038	0.594 (0.57)	1.042	

Table note: * P<0.05; ** P<0.01; *** P<0.001.

Table S9. Results of the GLMMs modelling the effects of experimental site (S), species (Sps) and their interaction in 12 fitness, phenotypic and phenological traits recorded in the only existing introduced population of *M*.*luteus*, and in *M*. *guttatus* and *M*. × *robertsii* populations from the north range of UK, during the reciprocal transplants experiment. The models comprising *M*. *luteus* and north populations of the other species include *M*. *guttatus* for fruits and stolons and *M*. x *robertsii* for stolons and the rest of phenotypic traits.

		<i>M. luteus</i> + north populations		M. luteus	
Trait	Fixed factor	Estimate (SE)	χ^2	Estimate (SE)	χ^2
Fruits	Intercept	2.855 (0.310)	84.708 ***	3.028 (0.244)	153.422 ***
	\mathbf{W}_0	-0.001 (0.000)	1.931	0.001 (0.002)	0.37
	Site (SHE)	-0.632 (0.040)	248.94 ***	-0.432 (0.044)	96.963 ***
	Sps (luteus)	0.310 (0.996)	0.097	× ,	
	Sps:Site (luteus:SHE)	0.188 (0.056)	11.018 ***		
Stolons	Intercept	1.978 (0.161)	151.243	2.169 (0.173)	156.919 ***
(guttatus)	\mathbf{W}_0	-0.002 (0.001)	2.17	-0.006 (0.004)	2.08
	Site (SHE)	-0.01 (0.06)	0.023	0.132 (0.072)	3.33
	Sps (robertsii)	0.041 (0.165)	0.061		
	Sps:Site (robertsii:SHE)	0.156 (0.093)	2.797		
Stolons	Intercept	1.949 (0.109)	316.918 ***		
(robertsii)	\mathbf{W}_0	0 (0)	0.291		
	Site (SHE)	0.155 (0.071)	4.825 *		
	Sps (robertsii)	-0.5 (0.149)	11.333 ***		
	Sps:Site (robertsii:SHE)	0.304 (0.101)	9.078 **		
Flowering	Intercept	41.764 (5.172)	65.216 ***	50.292 (4.257)	139.552 ***
day	\mathbf{W}_0	-0.009 (0.007)	1.691	-0.231 (0.103)	4.986 *
	Site (SHE)	20.916 (1.829)	130.82 ***	19.92 (1.821)	119.599 ***
	Sps (robertsii)	-3.656 (5.632)	0.421		
	Sps:Site (robertsii:SHE)	4.475 (2.362)	3.589		
Flowering	Intercept	5.882 (0.52)	128.041 ***	6.896 (0.789)	76.434 ***
node	\mathbf{W}_0	0 (0.001)	0.081	-0.026 (0.019)	1.858
	Site (SHE)	-0.337 (0.267)	1.593	-0.459 (0.344)	1.783
	Sps (robertsii)	-1.413 (0.585)	5.835 *	. ,	
	Sps:Site (<i>robertsii</i> :SHE)	0.465 (0.346)	1.807		

Branches	Intercept	1.989 (0.095)	442.485 ***	1.735 (0.177)	96.115 ***
	\mathbf{W}_0	0.001 (0)	5.914 *	0.007 (0.004)	3.596
	Site (SHE)	0.202 (0.067)	8.973 **	0.233 (0.07)	11.069 ***
	Sps (robertsii)	0.157 (0.123)	1.642		
	Sps:Site (robertsii:SHE)	-0.272 (0.088)	9.596 **		
Stems	Intercept	2.427 (0.103)	550.925 ***	2.185 (0.165)	175.731 ***
	\mathbf{W}_0	0.001 (0)	17.841 ***	0.007 (0.003)	5.628 *
	Site (SHE)	0.131 (0.053)	5.976 *	0.159 (0.055)	8.251 **
	Sps (robertsii)	0.184 (0.134)	1.878		
	Sps:Site (<i>robertsii</i> :SHE)	-0.254 (0.07)	13.314 ***		
Flowers	Intercept	4.948 (0.083)	3511.655 ***	4.81 (0.118)	1654.504 ***
	\mathbf{W}_0	0 (0)	56.073 ***	0.004 (0.001)	20.025 ***
	Site (SHE)	-0.043 (0.016)	6.993 **	-0.029 (0.016)	3.007
	Sps (robertsii)	0.161 (0.109)	2.194		
	Sps:Site (<i>robertsii</i> :SHE)	-0.101 (0.021)	22.965 ***		
Corolla	Intercept	332.43 (32.097)	107.27 ***	790.771 (95.861)	413.439 ***
	\mathbf{W}_0	-0.071 (0.041)	3.001	-1.474 (2.309)	6.956 **
	Site (SHE)	7.133 (10.831)	0.434	-69.603 (44.065)	0.076
	Sps (robertsii)	89.191 (34.911)	6.527 *		
	Sps:Site (<i>robertsii</i> :SHE)	6.448 (14.033)	0.211		
Stomata	Intercept	16.28 (2.911)	31.252 ***	18.028 (1.263)	203.76 ***
	\mathbf{W}_0	0.001 (0.002)	0.048	-0.046 (0.03)	2.362
	Site (SHE)	-2.036 (0.6)	11.538 ***	-2.254 (0.561)	16.127 ***
	Sps (robertsii)	-1.842 (3.127)	0.347		
	Sps:Site (<i>robertsii</i> :SHE)	1.835 (0.791)	5.378 *		
Height	Intercept	22.48 (1.478)	231.404 ***	19.889 (3.452)	33.197 ***
	\mathbf{W}_0	0.024 (0.006)	14.156 ***	0.092 (0.082)	1.277
	Site (SHE)	1.316 (1.539)	0.731	1.623 (1.541)	1.109
	Sps (robertsii)	-0.809 (1.957)	0.171		
	Sps:Site (<i>robertsii</i> :SHE)	-1.616 (2.028)	0.635		
Cover	Intercept	716.387 (113.072)	40.141 ***	790.771 (95.861)	68.048 ***
	\mathbf{W}_0	0.496 (0.171)	8.422 **	-1.474 (2.309)	0.407
	Site (SHE)	-60.482 (41.944)	2.079	-69.603 (44.065)	2.495
	Sps (robertsii)	-85.94 (124.445)	0.477		
	Sps:Site (<i>robertsii</i> :SHE)	1.896 (55.233)	0.001		
Dry mass	Intercept	37.018 (6.576)	31.692 ***	21.856 (5.162)	17.93 ***
	\mathbf{W}_0	0.021 (0.009)	4.989 *	0.421 (0.126)	11.132 ***
	Site (SHE)	-8.97 (2.304)	15.157 ***	-7.106 (2.476)	8.234 **
	Sps (robertsii)	-1.899 (7.197)	0.07		
	Sps:Site (robertsii:SHE)	-0.73 (3.039)	0.058		

Table note: Maximal model Estimates and Standard Errors for each fixed factor and interaction are provided jointly with results of the type-III Wald $\chi 2$ tests. $\chi 2$ values and indications of their associated *P*-values are provided. * P < 0.05; ** P < 0.01; *** P < 0.001. Significant effects after Bonferroni correction of *P*-values are indicated in bold. $\chi 2$ degrees of freedom=1-2.

Figure S1. Temperature and photoperiod conditions of the environmental models included in the chambers experiment. Squares represent day temperature and circles represent number of light hours per day in each fortnight in each of four chambers. Orange and blue symbols represent natural conditions in Isle of Wight and Shetland, respectively. Treatment codes indicate photoperiod (L, long; S, short) and temperature (W, warm; C, cold).

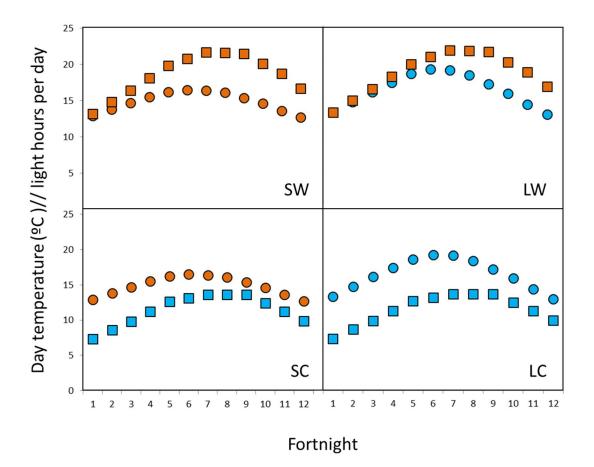


Figure S2. Correlation (A) and Principal Component Analyses (B) of the phenotypic traits measured in the controlled environmental chambers experiments for native *M. guttatus*, invasive *M. guttatus*, and *M. x robertsii* populations, and the three population types together.

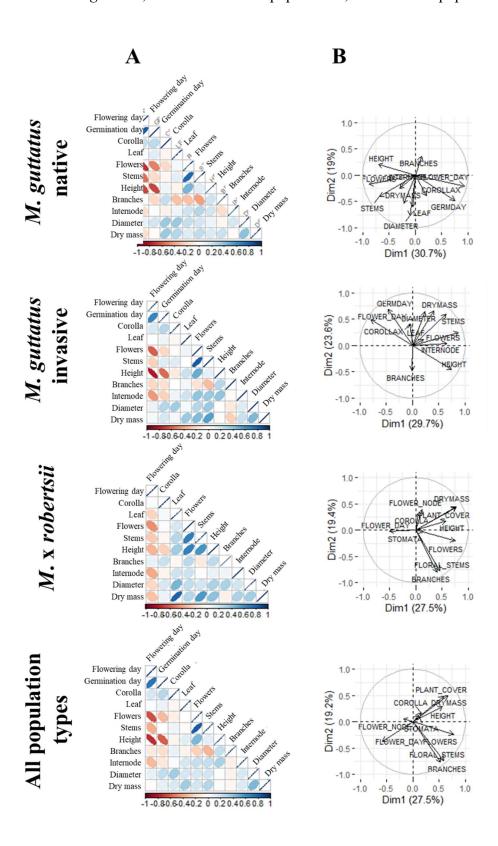


Figure S3. Correlation (A) and Principal Component Analyses (B) of the phenotypic traits measured for phenotypic selection analyses in the reciprocal transplants experiments for *M. guttatus*, *M. x robertsii*, *M. luteus*, and the three species togheter.

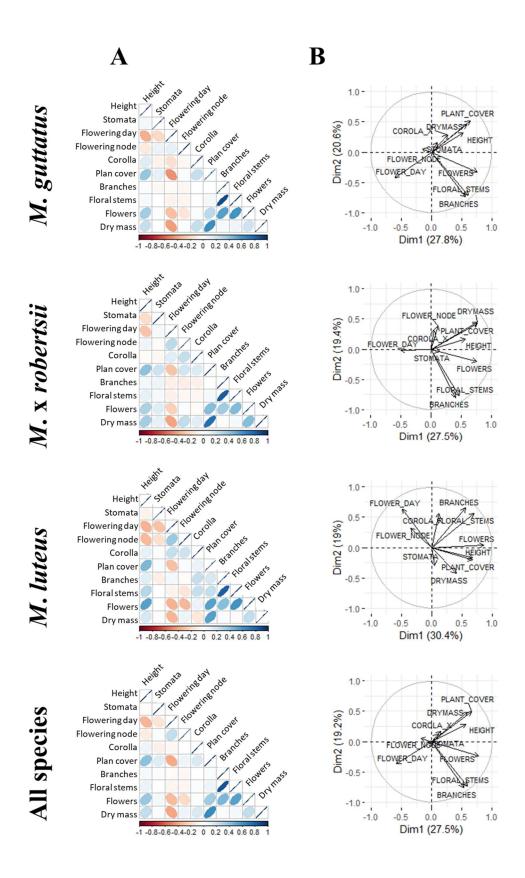


Figure S4. Phenotypic traits of native *Mimulus guttatus* (US), introduced *M. guttatus* (UK), and $M. \times robertsii$ (ROB) grew in four different controlled environmental chambers with contrasting photoperiods (L: long; S: short) and temperatures (W: warm; C: cold) in a crossed design. Mean values and standard errors of the variables measured are indicated by dots and error bars, respectively. Units as follows: Corolla width (mm), node diameter (mm), plant height (cm), internode length (cm), dry mass (g).

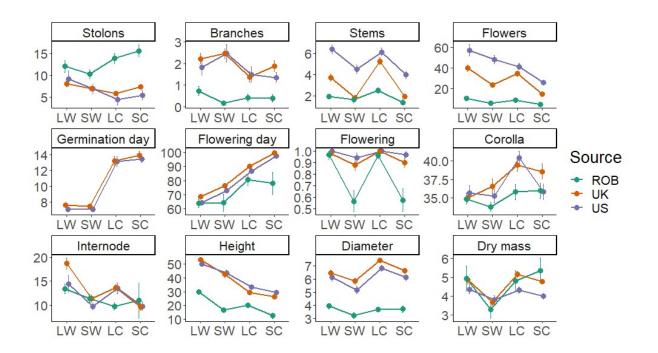


Figure S5. Reaction norms for each variable measured in the reciprocal transplant experiment of populations of *Mimulus guttatus* and *M.* × *robertsii* from different latitudes in the British Isles. Mean values and standard errors of the variables measured are indicated by dots and error bars, respectively. Units as follows: Corolla width (mm), plant height (cm), internode length (cm), dry mass (g), stomata (number per cm²), plant cover (cm²).

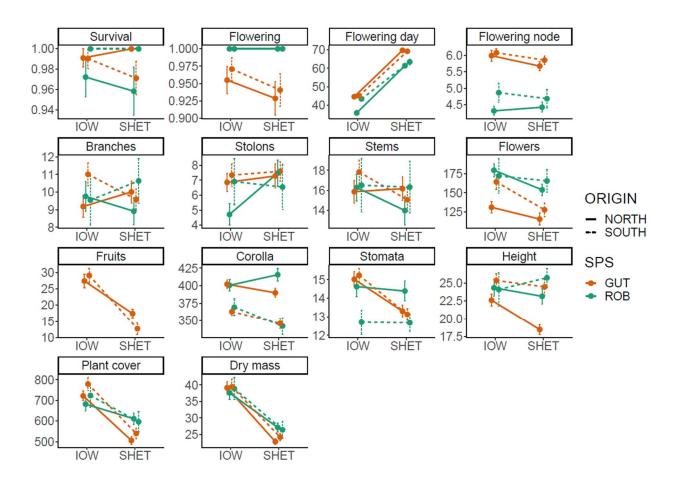


Figure S6. (A) Sexual and asexual fitness (mean number of fruits or stolons \pm s.d.) of the *M*. *luteus* individuals from the single population of this species included in the transplants experiment at different latitudes in the UK. (B) Estimates and 95% confidence intervals for the phenotypic selection coefficients on each trait and site included in the selection gradients.

