

Accepted refereed manuscript of: Quilliam RS, Taylor J & Oliver DM (2019) The disparity between regulatory measurements of *E. coli* in public bathing waters and the public expectation of bathing water quality. *Journal of Environmental Management*, 232, pp. 868-874.

DOI: [10.1016/j.jenvman.2018.11.138](https://doi.org/10.1016/j.jenvman.2018.11.138)

© 2018, Elsevier. Licensed under the Creative Commons Attribution-NonCommercial-NoDerivatives 4.0 International <http://creativecommons.org/licenses/by-nc-nd/4.0/>

1 **The disparity between regulatory measurements of *E. coli* in public bathing**  
2 **waters and the public expectation of bathing water quality**

3

4 Richard S. Quilliam\*, Jessica Taylor and David M. Oliver

5 Biological and Environmental Sciences, Faculty of Natural Sciences, University of

6 Stirling FK9 4LA, UK

7

8 \*corresponding author: Richard Quilliam

9 Email: [richard.quilliam@stir.ac.uk](mailto:richard.quilliam@stir.ac.uk);

10 Tel 0044 1786 467769

11

12

13 Keywords: coastal management; pollution; public health policy; recreation; risk

14 perception; tourism

15 **ABSTRACT**

16 The main objectives of the European Union (EU) Bathing Water Directive (BWD)  
17 2006/7/EC are to safeguard public health and protect designated aquatic  
18 environments from microbial pollution. The BWD is implemented through legislation  
19 by individual EU Member States and uses faecal indicator organisms (FIOs) as  
20 microbial pollution compliance parameters to determine season-end bathing water  
21 classifications (either 'Excellent', 'Good', 'Sufficient' or 'Poor'). These classifications  
22 are based on epidemiological studies that have linked human exposure to FIOs with  
23 the risk of contracting a gastrointestinal illness (GI). However, understanding public  
24 attitudes towards bathing water quality, together with perceptions of relative  
25 exposure risks, is often overlooked and yet critically important for informing  
26 environmental management decisions at the beach and ensuring effective risk  
27 communication. Therefore, this study aimed to determine the effectiveness of current  
28 regulatory strategies for informing beach users about bathing water quality, and to  
29 assess public understanding of the BWD classifications in terms of exposure risk and  
30 public health. Two UK designated bathing waters were selected as case studies, and  
31 questionnaires were deployed to beach-users. The bathing waters had different  
32 classification histories and both had electronic signage in operation for  
33 communicating daily water quality predictions. The majority of respondents did not  
34 recognise the standardised EU bathing water quality classification signs, and were  
35 unaware of information boards or the electronic signs predicting the water quality on  
36 that particular day. In general, respondents perceived the bathing water at their  
37 respective beach to be either 'good' or 'sufficient', which were also the lowest  
38 classifications of water quality they would be willing to accept for bathing. However,  
39 the lowest level of risk of contracting a gastrointestinal illness that respondents would

40 be willing to accept suggested a significant misunderstanding of the BWD  
41 classification system, with the majority (91 %) of respondents finding only a < 1 %  
42 risk level acceptable. The 'Good' classification is much less stringent in terms of  
43 likelihood of GI. This study has shown that the current public understanding of the  
44 BWD classifications in terms of exposure risk and public health is limited, and an  
45 investment in methods for disseminating information to the public is needed in order  
46 to allow beach-users to make more informed decisions about using bathing waters.

## 47 **Introduction**

48 Management of bathing waters in the European Union (EU) is currently  
49 legislated under the EU Bathing Water Directive (BWD) 2006/7/EC. The BWD has  
50 evolved since its introduction in 1976, helping to reduce faecal contamination of UK  
51 bathing waters primarily through the control and treatment of point-source sewage-  
52 related influx to surface waters (Kay et al., 2008). Microbial compliance parameters  
53 are fundamental for determining bathing water quality; the 2006 BWD uses two  
54 faecal indicator organisms (FIOs), *Escherichia coli* and intestinal enterococci (IE), as  
55 measures of microbial pollution in designated bathing waters (Kay et al., 2004).  
56 Sources of FIOs also include diffuse inputs, e.g. agricultural run-off, which are  
57 facilitated by precipitation and flooding events, and can lead to faecal contamination  
58 of both bathing waters and beaches through a variety of hydrological pathways (Kay  
59 et al., 2018; Bradford et al., 2013; Kay et al., 2007). Although FIOs are now  
60 considered poor surrogates for most pathogenic bacteria, viruses and protozoa (Wu  
61 et al., 2011), their presence is still widely accepted as an important indicator of faecal  
62 contamination. Furthermore, epidemiological studies have established that exposure  
63 to FIOs in recreational waters is significantly linked to a decrease in public health  
64 (Leonard et al., 2018; DeFlorio-Barker et al., 2018a), which has led to legislative  
65 pressures on EU member states to maintain and improve the microbial quality of  
66 designated bathing waters (Mansilha et al., 2009; Georgiou & Bateman, 2005).

67 The BWD requires that season-end classifications are produced for each  
68 bathing water, which are subsequently displayed at the start of the following bathing  
69 season. The classification ('Excellent', 'Good', 'Sufficient' or 'Poor') is calculated  
70 based on the previous four years of monitoring data, and uses either the 95<sup>th</sup> or 90<sup>th</sup>  
71 percentile, depending on the classification. There is also a requirement to produce a

72 bathing water profile, which is a freely accessible update on potential catchment  
73 pollutant sources linked to each specific bathing beach (Quilliam et al., 2015).  
74 However, debates on whether the BWD continues to be 'fit for purpose' for protecting  
75 public health (Oliver et al., 2014), together with the impact of increased non-  
76 compliance following the introduction of more stringent microbial water quality  
77 standards in the most recent revision of the BWD (Quilliam et al., 2015; Lušić et al.,  
78 2013), jeopardises the confidence of beach-user stakeholders in the effectiveness of  
79 current bathing water management (Oliver et al., 2016; Hynes et al., 2013; Langford  
80 et al., 2000).

81         Static signs, or information boards, at every designated bathing beach are  
82 also required under the BWD, in order to communicate to the public the classification  
83 of the beach and recent water quality results. In addition, some Member States also  
84 choose to use electronic signage network systems (based on rainfall and/or river  
85 flow data) that provide the public with real-time daily predictions of bathing water  
86 quality (Bedri et al., 2016; McPhail & Stidson, 2009). Public perception of bathing  
87 water quality can be heavily influenced by the media (Pendleton et al., 2001), and  
88 any obvious contamination of bathing water with litter, debris or scum, will often  
89 define a beach-user's perception of the water quality (Smith et al., 1991; House,  
90 1996). The health risks from bathing in recreational waters contaminated with  
91 pollution, however, can also be underestimated by the public (Langford et al., 2000),  
92 although there is some evidence that beach-users can exaggerate the level of health  
93 risk from bathing water (Fleisher & Kay, 2006). Understanding public perception of  
94 bathing water quality can be useful to beach managers for developing future  
95 management strategies (Kelly et al., 2018; Pouso et al., 2018). For example,  
96 particular beach management options such as beach grooming and seaweed

97 removal can increase public perception of the quality of a beach, although evidence  
98 suggests that the removal of seaweed can actually increase FIO levels in sand  
99 (Russell et al., 2014).

100 Beach classifications via the BWD also influence designations such as the  
101 Blue Flag award, the loss of which has the potential to impact tourism and local  
102 coastal economies (Saayman and Saayman, 2017). However, beach certification  
103 schemes are often used as an indicator of 'beach cleanliness' in an attempt to  
104 promote tourism rather than as a tool to promote environmental management (Klein  
105 & Dodds 2017). Their presence also has potential for misinterpretation by the public,  
106 for example beach awards may be perceived to signal zero risk of illness from sea-  
107 bathing. Understanding public attitudes towards bathing water, together with their  
108 perceptions of relative exposure risks, is critical for developing environmental  
109 management decisions and disseminating information to the public (Schernewski et  
110 al., 2018). Therefore, the aim of this study was to determine the effectiveness of  
111 current regulatory strategies for informing beach users about bathing water quality  
112 under the BWD, and to assess public understanding of the BWD classifications in  
113 terms of exposure risk and public health at two designated bathing waters in  
114 Scotland.

115

116 **Methods**

117 *Site description*

118 Two case-study bathing waters were used for this study: Largs (Pencil) Beach and  
119 Ayr (South) Beach, which are located on the West Coast of Scotland within 30 miles  
120 of each other. These two bathing waters were selected due to their different bathing  
121 water quality classifications, and the presence of electronic signage boards that  
122 provide daily water quality predictions for beach-users (McPhail and Stidson, 2009).

123

124 *Largs (Pencil) Beach*

125 Largs Beach is located in North Ayrshire (British National Grid Reference  
126 NS206580), close to the town of Largs, and was first designated as a bathing water  
127 in 2006. It is a sandy, relatively small bathing water of about 300 m in length, with a  
128 shallow bathing zone and two electronic signage boards. The beach is not groomed,  
129 and seaweed can accumulate at high tide marks. It has a small catchment area (1.2  
130 km<sup>2</sup>) with two streams that flow into the sea south of the bathing water, although  
131 there are no direct stream discharge points into the actual bathing water site. Parking  
132 access to the site is restricted, and there are no nearby food outlet areas; picnic  
133 benches for the public are set away from the bathing water and situated near public  
134 toilets. From 2012 to 2014 Largs was classed as 'Guideline' quality and from 2015  
135 onwards it has been classed as 'Excellent', with these classifications representing  
136 the highest standards from the 1976 and 2006 Directives, respectively.

137

138 *Ayr (South) Beach*

139 Ayr Beach is located adjacent to the town centre of Ayr (British National Grid  
140 Reference NS330218), South Ayrshire. Designated as a bathing water in 1987, the

141 beach is sandy and approximately 3.5 km in length. Ayr has a large catchment area:  
142 930 km<sup>2</sup> of mainly agricultural land, which drains into the three rivers within the  
143 catchment. The beach is regularly groomed by tractors removing seaweed and litter.  
144 There are two electronic signage boards, both sited on the promenade. There are  
145 many nearby food outlets, as well as a children's playground and other entertainment  
146 facilities, and public toilets. Parking is available directly along the length of the beach  
147 and access includes mobility-friendly ramps from the promenade down to the beach.  
148 From 2012 to 2014 Ayr was classed as mandatory quality; in 2015, it was classed as  
149 'Sufficient', while in 2016 it was downgraded to 'Poor'.

150

#### 151 *Bathing water predictions*

152 In Scotland, the EU BWD is implemented through The Bathing Water (Scotland)  
153 Regulations 2008, and the environmental management of bathing waters is  
154 regulated by the Scottish Environment Protection Agency (SEPA). Over the course  
155 of the bathing season, SEPA update their online "BeachLine" service with daily water  
156 quality predictions for 31 of the designated bathing waters in Scotland  
157 (<http://apps.sepa.org.uk/bathingwaters/Predictions.aspx>). These predictions are  
158 calculated through a computer-generated algorithm that considers, via telemetry, the  
159 antecedent 24 – 48 h precipitation volume and flows from river gauging stations  
160 against predefined trigger levels, which are used to remotely update the electronic  
161 signage boards located at the beaches. There were 107 days in the 2016 bathing  
162 season; daily water quality predictions were collected from the "BeachLine" website  
163 and recorded as either 'No Predicted Issues', 'Poor', or 'No Forecast' for both  
164 bathing waters.

165



166 *Collation of regulatory water monitoring data*

167 Monitoring of bathing water quality by SEPA occurs on predetermined days (per EU  
168 regulations, and a sampling calendar is released prior to the start of the bathing  
169 water season); results are entered onto their website a few days after the water  
170 sample has been processed. At the end of the bathing season, SEPA make all  
171 monitoring details along with any additional information publicly available. Water  
172 quality data for both Ayr Beach and Largs Beach were retrieved from the SEPA  
173 online database for the 2016 bathing season.

174

175 *Beach-users questionnaire*

176 Public perception of water quality, bathing water management, and associated risks  
177 at both Largs and Ayr were assessed using a face-to-face questionnaire (available  
178 upon request) at both beaches. Weekend days during the bathing season were  
179 chosen for questionnaire deployment due to expected higher numbers of beach  
180 users. Weather conditions were not used as a factor to determine days chosen, with  
181 a similar number of days experiencing 'good' or 'bad' weather at each location. The  
182 questionnaire was designed to assess the participant profile, their knowledge of  
183 beach attributes and bathing water monitoring, their views on bathing water quality  
184 and finally how the water quality would affect their use of the water for bathing. The  
185 questionnaire received ethical approval prior to conducting the study, and all  
186 participants were over 18 years of age.

187

188 *Statistical analyses*

189 All statistical analyses were carried out in Minitab 18.1 (Minitab Inc., PA, USA) and  
190 Differences at the  $p < 0.05$  level (95% confidence interval) were considered

191 statistically significant. The 95<sup>th</sup> and 90<sup>th</sup> percentile results were calculated from the  
192 SEPA water monitoring data for each FIO, using equations set out in the EU Bathing  
193 Water Directive. For the 95<sup>th</sup> percentile the equation  $\text{antilog}(\mu + 1.65\sigma)$  was used,  
194 while for the 90<sup>th</sup> percentile the equation  $\text{antilog}(\mu + 1.282\sigma)$  was used, where  $\mu$   
195 denotes the average concentration of each FIO, and  $\sigma$  denotes the standard  
196 deviation (CEC, 2006).

197

198 For the beach-user questionnaire, Chi-squared tests were used to determine  
199 whether there were significant associations between the location of the survey and  
200 participant responses to questions such as: distance beach-users travelled to the  
201 bathing water, perceived microbial water quality, and lowest classification of water  
202 beach-users considered acceptable for bathing. Mann-Whitney U tests were used to  
203 test for differences between beaches in terms of respondent age, average distance  
204 travelled to the beach, perceived water quality, how the quality of water would affect  
205 bathing decisions, and the level of risk considered acceptable for bathing. Finally,  
206 Spearman's correlation analysis was used to determine relationships between  
207 perceived water quality and the lowest level of water quality found acceptable for  
208 bathing at each bathing water site.

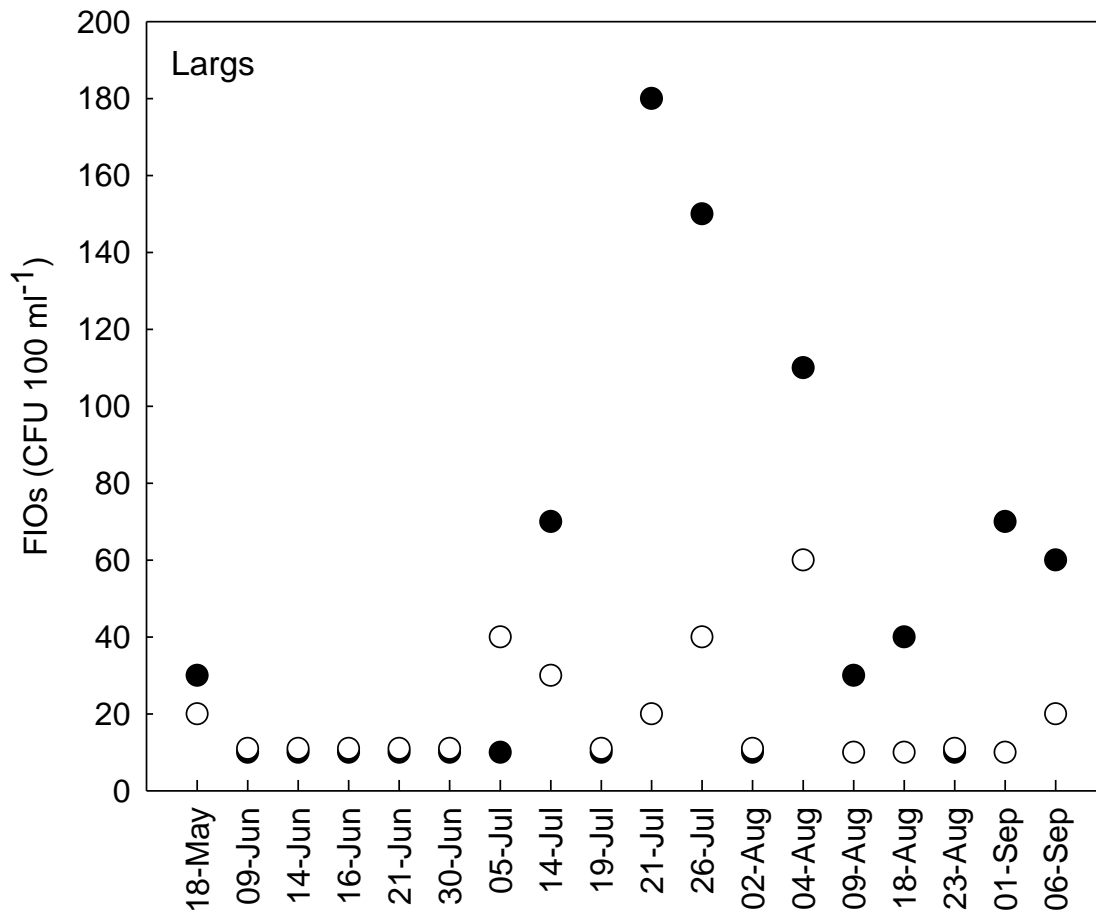
209 **Results**

210 *Bathing water predictions and regulatory water quality monitoring data*

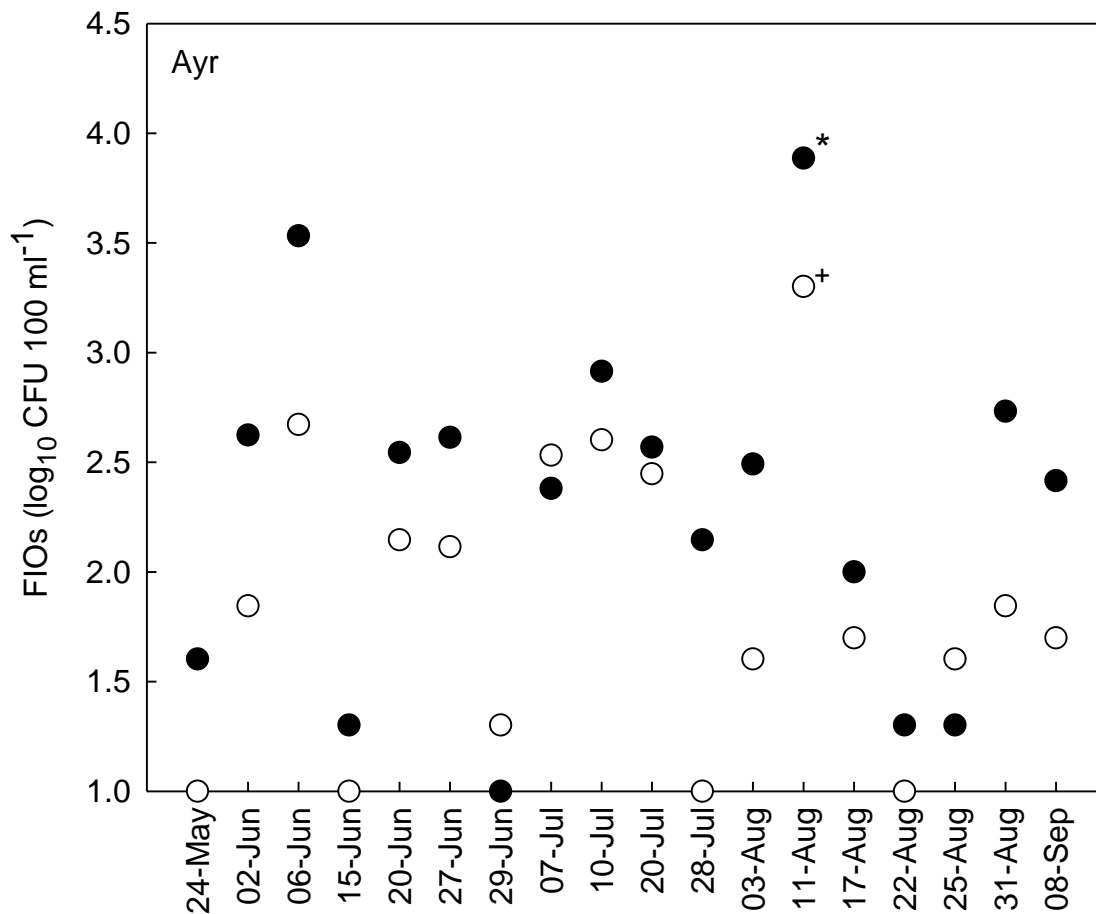
211 Over the course of the 2016 bathing season, 107 daily water quality  
212 predictions were collected from the “BeachLine” website. At Largs there were 17  
213 days classified as ‘Poor’, 88 days with ‘No Predicted Issues’, and two ‘No Forecast’  
214 days; in contrast, the predictions at Ayr Beach consisted of 54 ‘Poor’ days, 50 days  
215 classed as ‘No Predicted Issues’ and three days with ‘No Forecast’. Largs Beach  
216 had five times as many days with ‘No Predicted Issues’ compared to ‘Poor’ predicted  
217 water quality days, while over half of the bathing season at Ayr was predicted as  
218 ‘Poor’ bathing water quality.

219

220 As required by the EU BWD, microbial water quality was monitored at both  
221 Largs and Ayr Beaches 18 times during the 2016 bathing season. Over the course of  
222 the season, concentrations of *E. coli* were higher than IE at both bathing waters ( $P <$   
223 0.05), with generally higher concentrations of both FIOs at Ayr compared to Largs  
224 (Figures 1 and 2).



**Figure 1:** Regulatory microbial water quality monitoring results, *E. coli* (filled symbols) and enterococci (open symbols), at Largs Beach from the complete 2016 bathing water season.



**Figure 2:** Regulatory microbial water quality monitoring results, *E. coli* (filled symbols) and enterococci (open symbols), at Ayr Beach from the complete 2016 bathing water season. Labelled data points for *E. coli* (\*) and enterococci (+) were removed from the final compliance dataset.

225

226

227 At Largs, all 18 samples collected by the environmental regulator were  
 228 included in the algorithm to determine the 2016 season classification; however, at  
 229 Ayr Beach two samples were discounted by SEPA and therefore only 16 samples  
 230 were used to calculate the 2016 classification. Largs Beach, based on the 2016

231 water monitoring results, was classified under the BWD as ‘Excellent’ from both its IE  
 232 sampling results (42 CFU 100 ml<sup>-1</sup> at the 95<sup>th</sup> percentile, with an allowable amount of  
 233 100 CFU 100 ml<sup>-1</sup> for ‘Excellent’ ratings), and its *E. coli* results (156 CFU 100 ml<sup>-1</sup> at  
 234 the 95<sup>th</sup> percentile, with 250 CFU 100 ml<sup>-1</sup> the limit for ‘Excellent’ ratings; Table 1).  
 235

**Table 1:** Regulatory 90<sup>th</sup> and 95<sup>th</sup> percentile results for FIO monitoring. Values in bold denote the final 2017 classification level.

	<i>E. coli</i> (CFU 100 ml <sup>-1</sup> )		Intestinal enterococci (CFU 100 m <sup>-1</sup> )	
	90 <sup>th</sup> percentile	95 <sup>th</sup> percentile	90 <sup>th</sup> percentile	95 <sup>th</sup> percentile
Largs	105	<b>156</b>	34	<b>42</b>
Ayr (full data set)*	1967	3800	631	1223
Ayr (compliance)	<b>1094</b>	1957	<b>346</b>	620

\*The full dataset for Ayr was not used for the final classification due to the removal of two data points

236  
 237 Ayr beach, based on the 2016 compliance water monitoring results, was  
 238 classified as a ‘Poor’ bathing water from its IE concentration (346 CFU 100 ml<sup>-1</sup> at  
 239 the 90<sup>th</sup> percentile, with an allowable amount of 185 CFU 100 ml<sup>-1</sup> for a ‘Sufficient’  
 240 classification), and also its *E. coli* concentration (1094 CFU 100 ml<sup>-1</sup> at the 90<sup>th</sup>  
 241 percentile, with 500 CFU 100 ml<sup>-1</sup> the limit for a ‘Sufficient’ classification; Table 1).  
 242 These classifications were calculated with the compliance data set following the  
 243 removal of the two discounted samples. However, the *E. coli* concentration was not  
 244 significantly different between the compliance and full Ayr datasets.

245  
 246 *Public perception, image recognition and perceived water quality assessments*

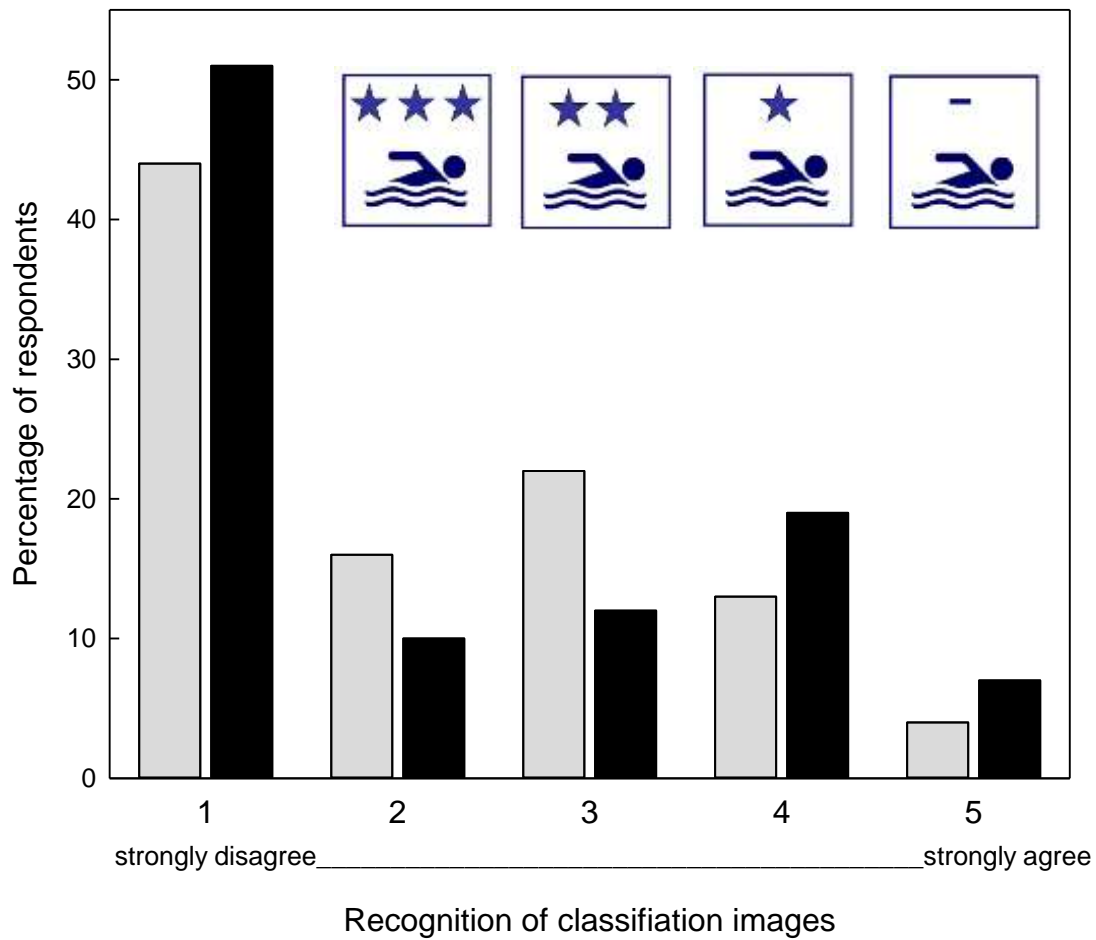
247 In total, there were 303 respondents to the questionnaire, 117 at Largs Beach  
 248 and 186 at Ayr Beach. Overall, 176 of these were women while 126 were men, with

249 similar proportions of male and female respondents at each of the two bathing  
250 waters. There was a significant difference in respondent age profile between the two  
251 bathing waters ( $P < 0.001$ ), with a modal age category for Largs  $> 60$  years, and for  
252 Ayr 30 – 39 years. There was a significant association between the distance that  
253 respondents had travelled to the bathing water and the survey location with people  
254 having travelled greater distances to visit Ayr beach than Largs beach ( $P < 0.001$ ).

255 At Largs, the most common reasons for visiting the beach were relaxation  
256 (42%), picnics (17%) and dog walking (16%), whilst at Ayr, the most common  
257 reasons were relaxation (40%), dog walking (18%) and other (for example, local  
258 attractions and children's play areas; 16%). Over a third of respondents at both  
259 beaches did not know if the bathing water was monitored (38% at Largs; 37% at  
260 Ayr), and of those respondents who did think it was monitored they either considered  
261 it to be the responsibility of the local council to monitor the water (32 % at Largs; 46  
262 % Ayr) or the duty of the environmental regulator (24 % at Largs; 17 % Ayr).  
263 Respondents at both bathing waters were asked whether they recognised the four  
264 EU bathing water quality signs that are used to communicate classifications (Figure  
265 3). A significantly higher proportion of respondents did not recognise the signs (60%  
266 at Largs and 61% at Ayr) compared to respondents who did recognise the signs ( $P <$   
267  $0.05$ ). Just under half (48%) of respondents at Largs knew that a water quality  
268 information board was present at the beach, while the majority of respondents at Ayr  
269 did not know if there was a board present (56%). At both locations, most  
270 respondents (60% at Largs and 72% at Ayr) did not know what the predicted water  
271 quality was on that particular day.

272

273



274 **Figure 3:** Responses to the question, “*Do you recognise these four EU*  
 275 *bathing water signs?*” of respondents at Largs (grey bars) and Ayr (black bars)  
 276 bathing beaches. Where 1 = strongly disagree and 5 = strongly agree. The insert  
 277 shows the four EU bathing water quality signs that are used for informing the public  
 278 about bathing water classifications

279

280

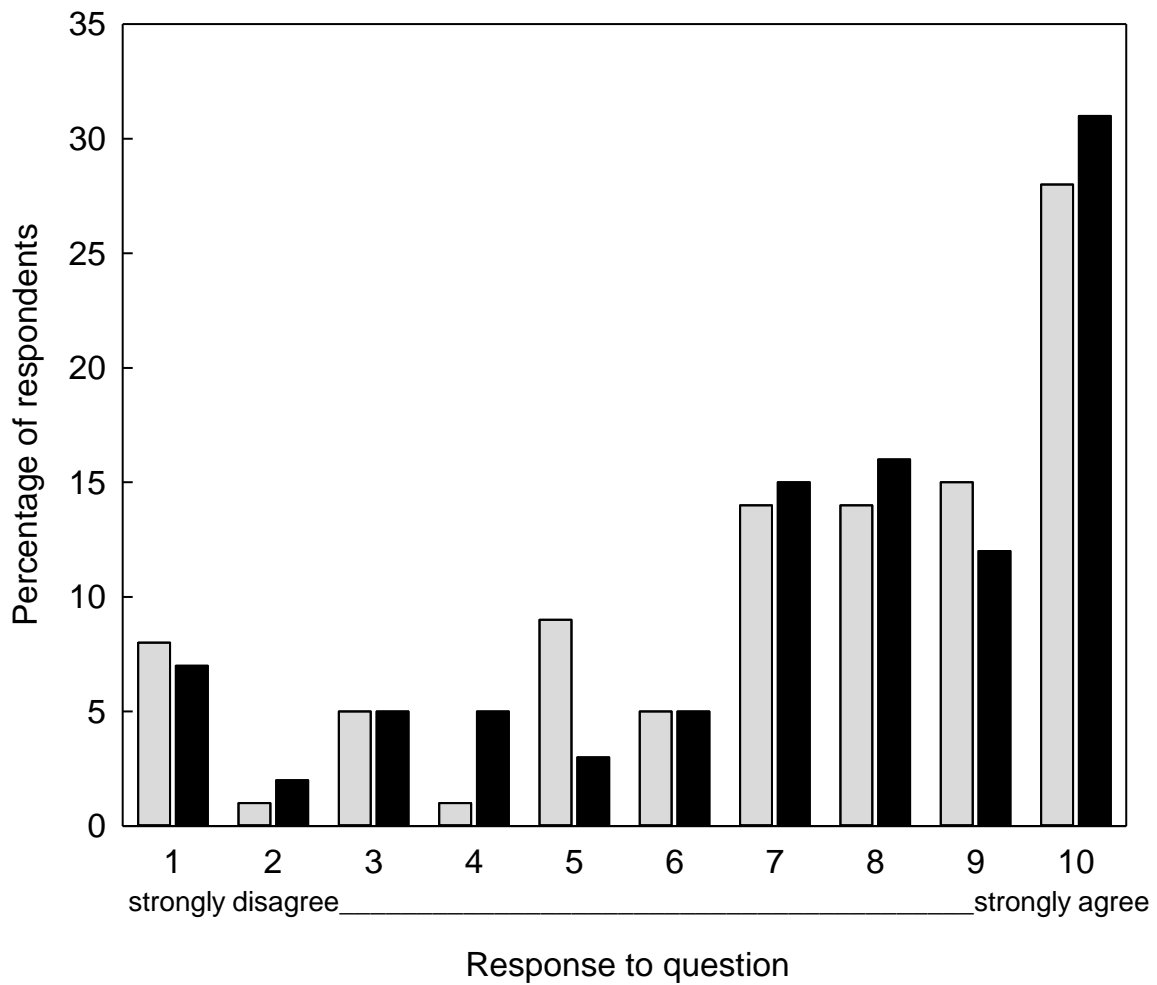


281 Information on perceived microbial water quality was obtained from  
282 respondents at both beaches on the day of each survey (i.e. their classification  
283 based on their perception of the water quality on that day, rather than being derived  
284 over the season), with significant differences between quality assessments at the  
285 two bathing waters ( $P < 0.001$ ). At Largs, 10% of respondents classed the bathing  
286 water as 'Excellent', while only 2% of respondents at Ayr thought the water was  
287 'Excellent' (Figure 4). The highest proportion of respondents at both bathing waters  
288 thought that the water was of 'Good' quality (64% at Largs, and 49% at Ayr). The  
289 majority of respondents at both beaches gave 'clarity of the water' as the reason for  
290 their choice. However, when asked whether the 'quality' of the seawater would affect  
291 their decision to go swimming in the sea, there was no significant difference between  
292 the two bathing waters in how the water quality would affect their decision ( $P > 0.05$ ),  
293 with the majority 'agreeing' or 'strongly agreeing' with this statement (Figure 5).  
294 Respondents were also asked for the lowest level of microbial water quality they  
295 would find acceptable before entering the water (Figure 6). There was a significant  
296 difference between the two bathing waters ( $P < 0.05$ ), with only one person at Largs  
297 willing to swim in poor quality water, compared to 14 people at Ayr. The majority of  
298 people would only accept a water quality classification of at least "Good" before  
299 swimming in the sea at both bathing waters. At Ayr, there were fewer respondents  
300 who perceived the water to be of 'Excellent' or 'Good' quality (52%) compared with  
301 the number of respondents who stated that only 'Excellent' or 'Good' water quality  
302 was acceptable for bathing (70%) (Figures 4 and 6).

303

304

305



**Figure 4:** Responses to the question, “Does the quality of the sea water affect your decision to go swimming in the sea?” by respondents at Largs (grey bars) and Ayr (black bars) bathing beaches. Where 1 = strongly disagree and 10 = strongly agree.

306

307

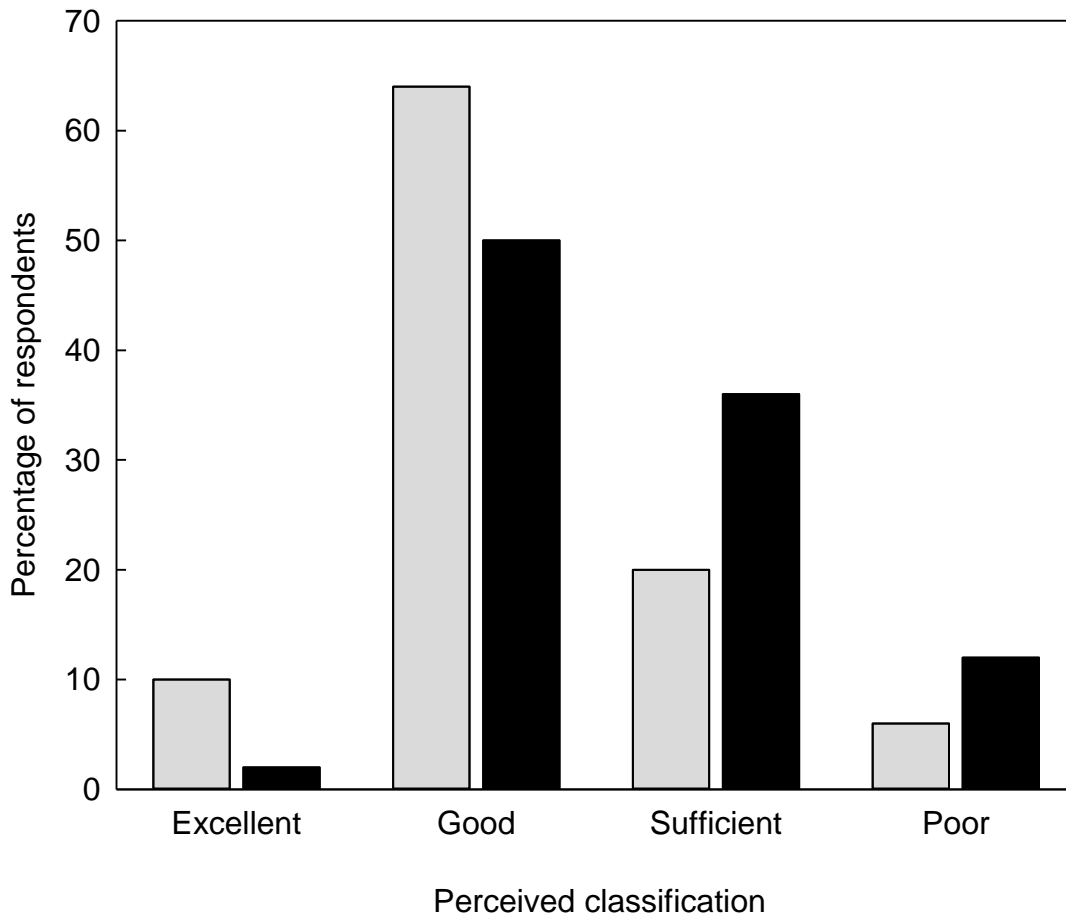
308

309

310

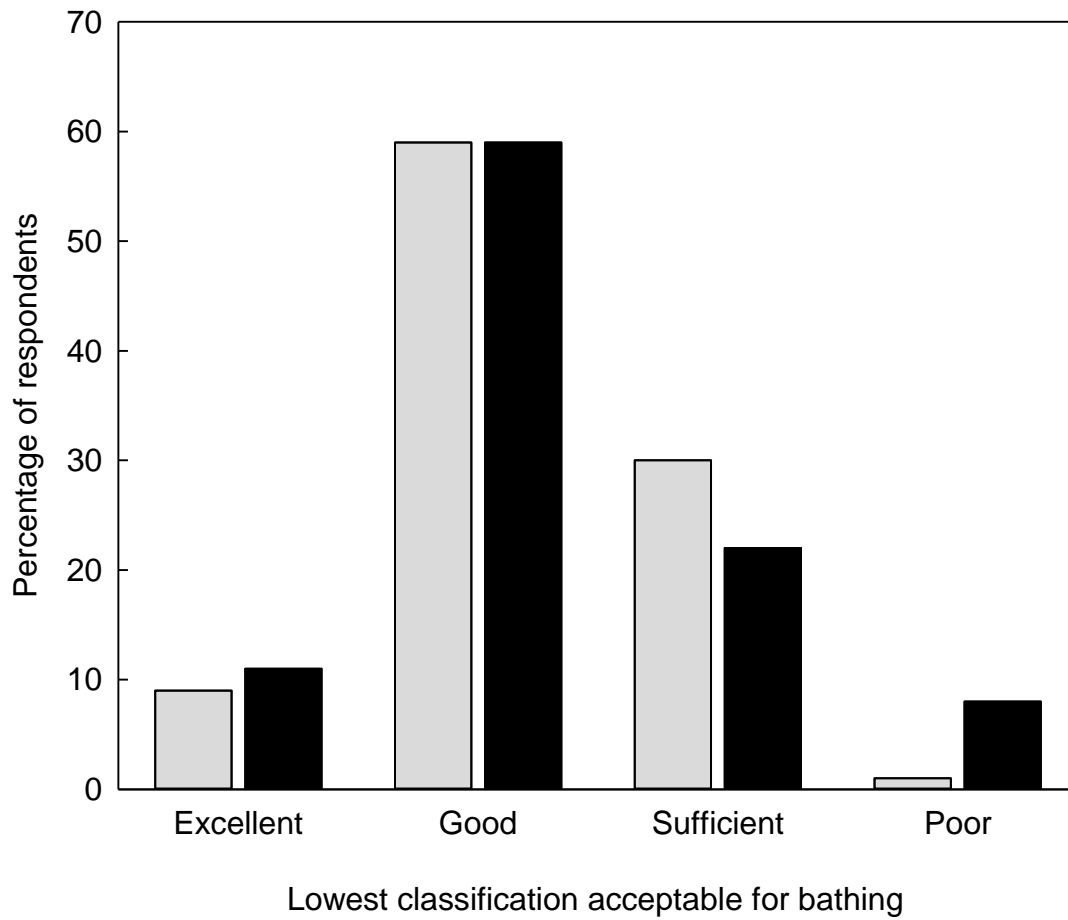
311

312  
313  
314  
315  
316  
317  
318  
319  
320  
321  
322  
323  
324  
325



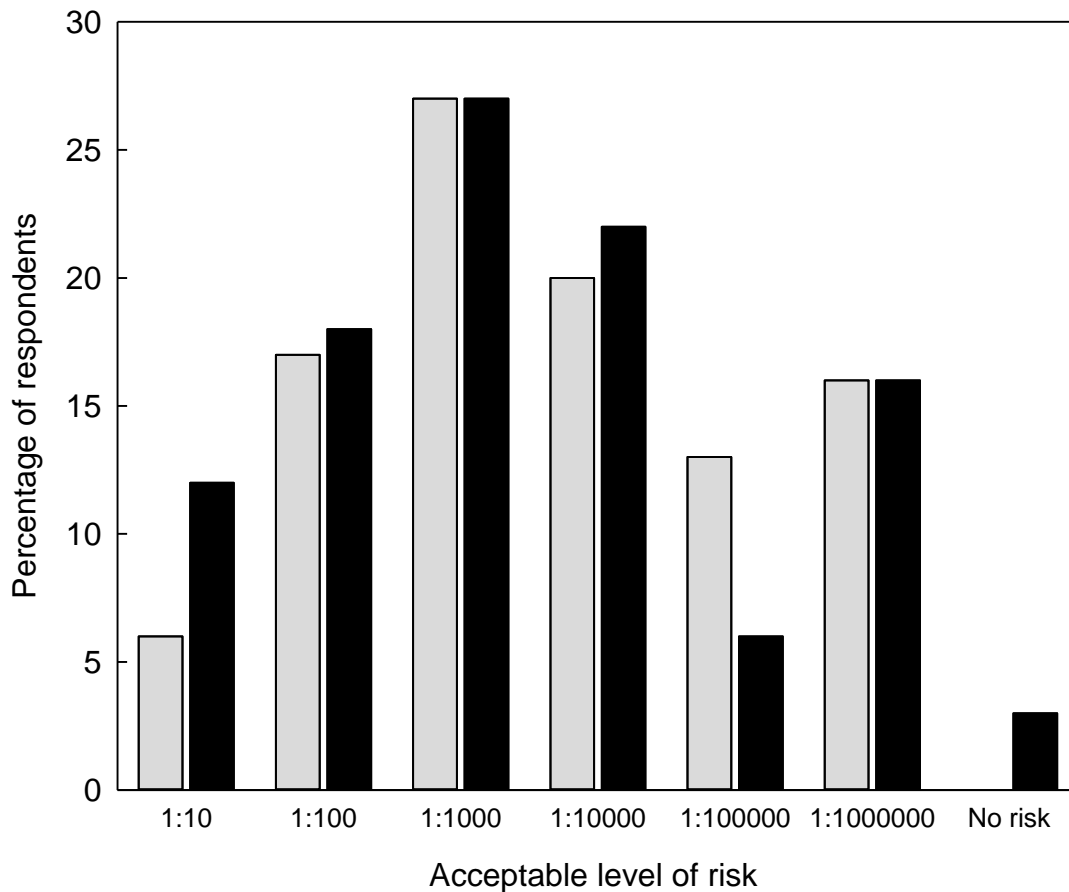
**Figure 5:** Perceived microbial water quality by respondents at Largs (grey bars) and Ayr (black bars) bathing beaches, answering the question, “*What do you think the bathing water quality at this beach is today?*”

326  
327  
328  
329  
330  
331  
332  
333



**Figure 6:** Lowest water quality respondents would find acceptable for bathing at Largs (grey bars) and Ayr (black bars) bathing beaches.

335 Finally, participants were asked what level of risk of becoming ill they would  
336 consider acceptable before entering the bathing water, ranging from a 10% risk  
337 (1:10) to a 0.0001% risk (1:1,000,000). There was no significant difference between  
338 the responses of people at the two bathing waters (Figure 7). At both bathing waters,  
339 the modal group was 1:1,000 (0.1% risk; 27% at both sites) with the next most  
340 frequently cited risk level being 1:10,000 (0.01% risk; 20% for Largs and 21% for  
341 Ayr).



**Figure 7:** The level of risk of illness, e.g. stomach bug or ear infection, that respondents would accept before they considered entering the sea at Largs (grey bars) and Ayr (black bars) bathing beaches.

342

343 **Discussion**

344           The 2006 BWD requires environmental regulators to provide information to  
345 the public in the form of a bathing water profile, freely available updates on potential  
346 pollutant sources, and an information board at each designated bathing water that  
347 clearly displays recent microbial compliance results. In this study, in addition to  
348 information-rich notice boards, both beaches also have electronic signage boards  
349 installed that clearly state the predicted quality of the water for that particular day.  
350 Yet, over half of all respondent at both beaches were unaware of these information  
351 boards on bathing water quality, and over a third of respondents did not know that  
352 the beach was even monitored for water quality. The aim of the BWD is to reduce  
353 public exposure to contaminated bathing waters by providing an evidence-based  
354 classification system linked to an epidemiological risk of getting ill. However, there is  
355 currently a significant lack of data on beach user's knowledge of water quality  
356 classification under the BWD and the public perception of exposure risk at bathing  
357 waters. This study has shown that the current public understanding of the BWD  
358 classifications in terms of exposure risk and public health is limited, irrespective of  
359 the compliance record of a beach, and that there is a need to improve the delivery of  
360 information to beach stakeholders about the quality of bathing water.

361           During the bathing water season, water quality predictions from the  
362 "BeachLine" web service are created daily; however, the predictions are not updated  
363 during the course of the day, and are released each morning and remain accessible  
364 until the evening. Large rain events during the day can deliver FIOs to bathing  
365 waters from upstream in the catchment; yet bathing water advice remains the same  
366 until the following day's prediction is released. Consequently, although most of the  
367 time prediction tools are correct, or at least precautionary, (McPhail and Stidson,

368 2009), 'Beachline' predictions are unable to respond quickly enough to pollution  
369 events.

370 The sampling calendar for collecting compliance data is planned ahead of the  
371 bathing season, with the frequency and number of samples determined by individual  
372 EU member states, which means that pollution events occurring between samples  
373 could go unnoticed by the regulator even though the public continues to use the  
374 bathing water. Short term pollution events can considerably effect the final  
375 classification of a bathing water (Figueras et al., 2015), for example, if regulatory  
376 sampling occurs after a pollution event or heavy rain; therefore, the BWD allows a  
377 maximum of 15 % of samples to be discounted from the final classification  
378 calculation. Being able to remove "Poor" samples to potentially avoid a season's-end  
379 "Poor" classification is arguably more representative of the normal classification  
380 conditions throughout the season, and gives the public more confidence in using the  
381 bathing water in the future.

382 Over 50% of the Beachline predictions at Ayr beach were for "Poor" quality  
383 bathing water, and an electronic sign predicting poor bathing water quality for over  
384 half of the bathing season is likely to have had a negative economic effect on the  
385 local community. With current climate change projections, there are increasing  
386 threats of summer storms driving elevated FIO transfer from land to water. Under  
387 these circumstances, reducing the likelihood of human exposure to contaminated  
388 bathing water through real-time information provision at bathing beaches offers an  
389 extra line of public health defence, if communicated effectively. Bathing waters that  
390 receive runoff from large agricultural catchments, such as Ayr, are particularly  
391 susceptible to high FIO loading after heavy rain, and although the algorithm for  
392 prediction models can be over cautious, they do provide more of a real-time

393 indication of bathing water quality compared to regulatory microbial compliance data,  
394 which by the time it is ready to be communicated to the public is already 2-3 days out  
395 of date (Oliver et al., 2014; Searcy et al., 2018).

396 The majority of respondents in our study did not recognise the four EU  
397 bathing water quality signs that have been developed to give a quick visual guide to  
398 current bathing water quality (European Commission, 2016), which further  
399 demonstrates shortcomings in the public's awareness of the BWD. In the past  
400 decade, regulators in the EU have made bathing water information freely available  
401 on their websites, whilst social media and text-message services are also used to  
402 disseminate current water quality information. However, despite the information  
403 boards and electronic signs, and several sources of online information, most beach-  
404 users in our study were unaware of the predicted quality of the bathing water on the  
405 day that they were at the beach. This issue seems to stem from more than just a lack  
406 of knowledge of who is responsible for monitoring water quality at beaches, but  
407 rather suggests that most beach users make the decision of entering a bathing water  
408 based upon their own perception of the water quality, e.g. by how 'clean' it looks or  
409 smells (Pratap et al., 2013; Jones et al., 2018). Constantly having a prediction of  
410 "Poor" water quality at a beach can also cause user fatigue (Kim and Grant, 2004),  
411 and may lead beach users to dismiss or ignore water quality predictions, leading to a  
412 public health risk as users wilfully ignore warnings that are considered out of date or  
413 overly cautious. This may explain the number of beach-users at Ayr beach who were  
414 willing to go swimming in water classified as 'poor' at the time of our study.

415 Bathing waters are heavily influenced by land use and management within the  
416 catchment, and as part of the EU Water Framework Directive, environmental  
417 regulators target water quality improvements at the catchment scale through their



418 River Basin Management Plan (RBMP). Indicators of faecal contamination in bathing  
419 water, such as *E. coli*, provide a measure of the level of faecal contamination rather  
420 than an indication of the presence of waterborne pathogens (Field and Samadpour,  
421 2007). Nonetheless, the presence of waterborne FIOs have been linked to illness in  
422 bathers (Leonard et al., 2018), which is associated with a significant economic  
423 burden (DeFlorio-Barker et al., 2018b); therefore, exposure to poor quality bathing  
424 waters needs to be managed through evidence-based guidelines and targeted  
425 research leading to improvements in risk communication.

426 The two BWD microbial compliance parameters, *E. coli* and IE, are  
427 enumerated by culture-dependent methods, which is often criticised in terms of the  
428 speed of analysis. Processing FIOs in water samples can take between 18 - 96 h, by  
429 which time the bathing water quality information is out-of-date for the recreational  
430 beach-user. The transfer of existing molecular biological tools, such as qPCR, is  
431 therefore gaining momentum largely because these methods permit a much more  
432 rapid sample processing time (1 - 2 hours). Thus, molecular biological tools offer an  
433 opportunity to provide a more meaningful statement of microbial risk to water users  
434 by providing near-real-time information, and thus enable more informed decision-  
435 making for water-based activities (Dorevitch et al., 2017). However, our study has  
436 suggested that the majority of beach-users don't engage with information on bathing  
437 water quality, and so the value of an increase in the speed of reporting microbial  
438 water quality to a typical beach-user is limited. Furthermore, FIO concentrations in  
439 seawater can vary during the day, e.g. due to fluctuations in UV irradiance, tides and  
440 wind direction; consequently, the time of day water samples are collected, together  
441 with the location they are taken from, can significantly affect concentrations of FIOs  
442 at a bathing water (Jennings et al., 2018; Quilliam et al., 2011), and so even with a

443 more rapid processing time, the information given to the beach-user remains out of  
444 date. Whilst not perfect, culture-based methods, together with molecular approaches  
445 and modelling prediction tools allow beach managers to inform beach users about  
446 the risk of using a bathing water (Bedri et al., 2016; Oliver et al., 2014; Thoe et al.,  
447 2014).. However, our study has clearly demonstrated that resources should now be  
448 focussed on developing improved methods of disseminating information on water  
449 quality to the beach-user.

450         The public perception of bathing water can be influenced by the media or the  
451 historical reputation of a beach (Pendleton et al., 2001). Ayr beach has had a recent  
452 record of poorer water quality than at Largs beach, which may explain the lower  
453 perceived classification that respondents gave about the quality of the water at Ayr,  
454 i.e. beach-users have got accustomed to the water quality being described as poor  
455 at that beach. This also raises questions about how visitors and residents might  
456 respond differently to bathing water quality information (Dodds & Holmes, 2018). The  
457 lowest classification of water quality that respondents would find acceptable for  
458 bathing was very similar between the two beaches, with the majority stating that the  
459 water would need to be at least 'Good' or 'Sufficient' for them to be use it. However,  
460 this is not reflected in the responses to the risk of contracting a gastrointestinal  
461 illness (GI), with the majority of people (91 % of all respondents) only willing to  
462 accept a risk of 1% or less of getting ill (the 'Good' classification represents ~ 5% GI  
463 risk). Over the last decade, the risk of bathers contracting a GI has received  
464 increasing attention, and whilst it is generally accepted that using contaminated  
465 bathing water can increase the risk of GI (Leonard et al., 2018), there is concern that  
466 this correlation is inconsistent due to differences in data collection between  
467 epidemiological studies (King et al., 2014). Although, there remains little evidence for

468 a significant dose response between FIOs and GI in marine water, guidelines  
469 suggest that bathing water with  $< 40$  CFU IE  $100\text{ ml}^{-1}$  presents  $< 1\%$  risk of  
470 contracting a GI disease (this concentration of IE would equate to an 'Excellent'  
471 classification under the BWD) (Kay et al., 2004). This is not consistent with the  
472 lowest classification of water the respondents would be willing to use, with only 11 %  
473 of respondents saying that 'Excellent' was the lowest classification they would  
474 accept. The risk of contracting a GI in bathing water with 41 – 200 CFU IE  $100\text{ ml}^{-1}$   
475 (classification of 'Good') rises to 1 –  $< 4\%$ ; whilst a concentration of  $> 500$  CFU IE  
476  $100\text{ ml}^{-1}$  presents a risk of  $> 10\%$  of GI disease (Kay et al., 2004). Our study has  
477 shown that 8 % of respondents were willing to accept a 10 % risk of contracting a GI,  
478 which would equate to a 'Sufficient' classification under the BWD. The perception of  
479 risk of illness from using bathing water suggests a lack of understanding of how the  
480 classification system of the BWD relates to risk. However, this is not exclusive to the  
481 BWD as there is a similar situation in the US, with a limited awareness of water  
482 quality regulatory guidelines among beach-users, and an unrealistic perception of the  
483 risk of contracting a GI (Pratap et al., 2013).

484

## 485 **Conclusion**

486 Recreational bathing waters are an integral part of the environment, and  
487 provide significant health and economic benefits. Improvements in bathing water  
488 quality in the EU, legislated under the BWD, have provided much safer environments  
489 for the public. However, there remains a disconnect between the regulator who is  
490 enforcing policy and the beach-user as a stakeholder. Environmental regulators  
491 channel significant time and resources into conforming to the BWD, but there now  
492 needs to be greater focus on the methods for disseminating bathing water quality

493 information to beach users, with further development needed in order to attract  
494 greater public engagement. Social media may play a more significant future role for  
495 publicising information to beach-users, although it is well recognised that social  
496 media is not uniformly used across age groups, and tourists to the area may not  
497 speak the local language. There is a willingness to pay for improvements in bathing  
498 water quality among beach-users (Hynes et al., 2013); therefore, investing in  
499 innovative and novel methods for clearly disseminating the significance of  
500 compliance data under the BWD would increase public engagement, and allow  
501 beach-users to make more informed decisions about using bathing waters and  
502 deciding whether to allow their children to play in the water.

503

504

505 **References**

506

507 Bedri, Z., Corkery, A., O'Sullivan, J.J., Deering, L.A., Demeter, K., Meijer, W.G.,  
508 O'Hare, G. and Masterson, B., 2016. Evaluating a microbial water quality prediction  
509 model for beach management under the revised EU bathing water directive. *Journal*  
510 *of Environmental Management*, 167, pp.49-58.

511

512 Bradford, S.A., Morales, V.L., Zhang, W., Harvey, R.W., Packman, A.I., Mohanram,  
513 A. and Welty, C., 2013. Transport and fate of microbial pathogens in agricultural  
514 settings. *Critical Reviews in Environmental Science and Technology*, 43(8), pp.775-  
515 893.

516

517 Canter, L.W., Nelson, D.I., and Everett, J.W. (1994) Public perception of water  
518 quality risks – influencing factors and enhancement opportunities. *Journal of*  
519 *Environmental Systems*. **22(2)**: 163-167.

520

521 DeFlorio-Barker, S., Arnold, B.F., Sams, E.A., Dufour, A.P., Colford Jr, J.M.,  
522 Weisberg, S.B., Schiff, K.C. and Wade, T.J., (2018a). Child environmental exposures  
523 to water and sand at the beach: findings from studies of over 68,000 subjects at 12  
524 beaches. *Journal of Exposure Science & Environmental Epidemiology*, 28(2), p.93.

525

526 DeFlorio-Barker, S., Wing, C., Jones, R. M., & Dorevitch, S. (2018b). Estimate of  
527 incidence and cost of recreational waterborne illness on United States surface  
528 waters. *Environmental Health*, 17(1), 3.

529

530 Dodds, R. and Holmes, M.R., (2018). Education and certification for beach  
531 management: is there a difference between residents versus visitors?. *Ocean &*  
532 *Coastal Management*. 160, 124-132.

533

534 Dorevitch, S., Shrestha, A., DeFlorio-Barker, S., Breitenbach, C., & Heimler, I.  
535 (2017). Monitoring urban beaches with qPCR vs. culture measures of fecal indicator  
536 bacteria: Implications for public notification. *Environmental Health*, 16(1), 45.

537

538 European Commission. (2016) Bathing water quality. Signs and symbols adopted by  
539 the commission. [ONLINE] [http://ec.europa.eu/environment/water/water-](http://ec.europa.eu/environment/water/water-bathing/signs.htm)  
540 [bathing/signs.htm](http://ec.europa.eu/environment/water/water-bathing/signs.htm)

541

542 Field, K. G., & Samadpour, M. (2007). Fecal source tracking, the indicator paradigm,  
543 and managing water quality. *Water Research*, 41(16), 3517-3538.

544

545 Fleisher, J. M., & Kay, D. (2006). Risk perception bias, self-reporting of illness, and  
546 the validity of reported results in an epidemiologic study of recreational water  
547 associated illnesses. *Marine Pollution Bulletin*, 52(3), 264-268.

548

549 Figueras, M.J., de Torres, M., Silvera, C., Corrales, M.J., 2015. Monitoring  
550 Programmes for Bathing Waters Within the Frame of the EU Bathing Water  
551 Directive: The Experience of Catalonia. In: Munné, A., Ginebreda, A., Prat, N. (Eds.),  
552 Experiences from Ground, Coastal and Transitional Water Quality Monitoring. The  
553 EU Water Framework Directive Implementation in the Catalan River Basin District 2.  
554 Springer, Switzerland, pp. 301–333

555

556 Georgiou, S. and Bateman, I.J. (2005). Revision of the EU Bathing Water Directive:  
557 economic costs and benefits. *Marine Pollution Bulletin*, 50,.430-438.

558

559 House, M.A. (1996) Public perception and water quality management. *Water*  
560 *Science and Technology*. **34(12)**: 25-32.

561

562 Hynes, S., Tinch, D., & Hanley, N. (2013). Valuing improvements to coastal waters  
563 using choice experiments: An application to revisions of the EU Bathing Waters  
564 Directive. *Marine Policy*, 40, 137-144.

565

566 Jennings, W. C., Chern, E. C., O'Donohue, D., Kellogg, M. G., & Boehm, A. B.  
567 (2018). Frequent detection of a human fecal indicator in the urban ocean:  
568 environmental drivers and covariation with enterococci. *Environmental Science:*  
569 *Processes & Impacts*, 20, 480-492.

570

571 Jones, J., Aslan, A., Trivedi, R., Olivas, M. and Hoffmann, M., (2018). Water quality  
572 and the perception of risk: A study of Georgia, USA, beachgoers. *Ocean & Coastal*  
573 *Management*, 158,.116-119.

574

575 Kay, D., Bartram, J., Pruss, A., Ashbolt, N., Wyer, M.D., Fleisher, J.M., Fewtrella, L.,  
576 Rogers, A., and Rees, G. (2004) Derivation of numerical values for the World Health  
577 Organization guidelines for recreational waters. *Water Research*. **38**: 1296-1304.

578

579 Kay, D., Edwards, A.C., Ferrier, R.C., Francis, C., Kay, C., Rushby, L., Watkins, J.,  
580 McDonald, A.T., Wyer, M., Crowther, J. Wilkinson, J. (2007). Catchment microbial  
581 dynamics: the emergence of a research agenda. *Progress in Physical*  
582 *Geography*, 31, 59-76

583

584 Kay, D., Crowther, J., Stapleton, C.M., Wyer, M.D., Fewtrell, L., Edwards, A.,  
585 Francis, C.A., McDonald, A.T., Watkins, J. and Wilkinson, J., (2008). Faecal indicator  
586 organism concentrations in sewage and treated effluents. *Water Research*, 42, 442-  
587 454.

588

589 Kay, D., Crowther, J., Stapleton, C.M. and Wyer, M.D. (2018). Faecal indicator  
590 organism inputs to watercourses from streamside pastures grazed by cattle: Before  
591 and after implementation of streambank fencing. *Water Research*, 143, 229-239.

592

593 Kelly, E.A., Feng, Z., Gidley, M.L., Sinigalliano, C.D., Kumar, N., Donahue, A.G.,  
594 Reniers, A.J. and Solo-Gabriele, H.M., 2018. Effect of beach management policies  
595 on recreational water quality. *Journal of Environmental Management*, 212, 266-277.

596

597 Kim, J. H., & Grant, S. B. (2004). Public mis-notification of coastal water quality: A  
598 probabilistic evaluation of posting errors at Huntington Beach, California. *Environ.*  
599 *Sci. Technol.* 38, 2497–2504

600

601 King S , Exley J, Winpenny E, Alves L, Henham M, Larkin J. (2014). The Health  
602 Risks of Bathing in Recreational Waters: A Rapid Evidence Assessment of Water  
603 Quality and Gastrointestinal Illness. Cambridge, UK: RAND Europe.



604

605 Klein, L., & Dodds, R. (2017). Perceived effectiveness of Blue Flag certification as an  
606 environmental management tool along Ontario's Great Lakes beaches. *Ocean &*  
607 *Coastal Management*, 141, 107-117.

608

609 Langford, I.H., Georgiou, S., Bateman, I.J., Day, R.J. and Turner, R.K., (2000).  
610 Public perceptions of health risks from polluted coastal bathing waters: a mixed  
611 methodological analysis using cultural theory. *Risk Analysis*, 20, 691-704.

612

613 Leonard, A. F., Singer, A., Ukoumunne, O. C., Gaze, W. H., & Garside, R. (2018). Is  
614 it safe to go back into the water? A systematic review and meta-analysis of the risk  
615 of acquiring infections from recreational exposure to seawater. *International Journal*  
616 *of Epidemiology*, 47, 572-586.

617

618 Lušić, D. V., Lušić, D., Pešut, D., Mićović, V., Glad, M., Bilajac, L., & Peršić, V.  
619 (2013). Evaluation of equivalence between different methods for enumeration of  
620 fecal indicator bacteria before and after adoption of the new Bathing Water Directive  
621 and risk assessment of pollution. *Marine Pollution Bulletin*, 73, 252-257.

622

623 Mansilha, C. R., Coelho, C. A., Heitor, A. M., Amado, J., Martins, J. P., & Gameiro,  
624 P. (2009). Bathing waters: new directive, new standards, new quality  
625 approach. *Marine Pollution Bulletin*, 58, 1562-1565

626

627 McPhail, C.D. and Stidson, R.T. (2009) Bathing water signage and predictive water  
628 quality models in Scotland. *Aquatic Ecosystem Health & Management*. 12: 183-186.

629

630 Oliver, D.M., van Niekerk, M., Kay, D., Heathwaite, A.L., Porter, J., Fleming, L.E.,  
631 Kinzelman, J.L., Connolly, E., Cummins, A., McPhail, C., Rahman, A., Thairs, T., de  
632 Roda Husman, A.M., Hanley, N.D., Dunhill, I., Globevnik, L., Harwood, V.J.,  
633 Hodgson, C.J., Lees, D.N., Nichols, G.L., Nocker, A., Schets, C., and Quilliam, R.S.  
634 (2014) Opportunities and limitations of molecular methods for quantifying microbial  
635 compliance parameters in EU bathing waters. *Environment International*. 64: 124-  
636 128.

637

638 Oliver, D.M., Hanley, N.D., van Niekerk, M., Kay, D., Heathwaite, A.L., Rabinovici,  
639 S.J.M., Kinzelman, J.L., Fleming, L.E., Porter, J., Shaikh, S., Fish, R., Chilton, S.,  
640 Hewitt, J., Connolly, E., Cummins, A., Glenk, K., McPhail, C., McRory, E., McVittie,  
641 A., Giles, A., Roberts, S., Simpson, K., Tinch, D., Thairs, T., Avery, L.M., Vinten,  
642 A.J.A., Watts, B. & Quilliam, R.S. (2016) Molecular tools for bathing water  
643 assessment in Europe: balancing social science research with a rapidly developing  
644 environmental science evidence-base. *AMBIO*. 45: 52-62.

645

646 Pendleton, L., Martin, N., and Webster, D.G. (2001) Public Perceptions of  
647 Environmental Quality: A Survey Study of Beach Use and Perceptions in Los  
648 Angeles County. *Marine Pollution Bulletin*. 42: 1155-1160.

649

650 Pouso, S., Uyarra, M. C., & Borja, Á. (2018). The recovery of estuarine quality and  
651 the perceived increase of cultural ecosystem services by beach users: A case study  
652 from northern Spain. *Journal of Environmental Management*, 212, 450-461.

653

654 Quilliam RS, Clements K, Duce C, Cottrill SB, Malham SK, Jones DL. (2011). Spatial  
655 variation of waterborne *Escherichia coli* – implications for routine sampling of water  
656 quality. *Journal of Water & Health* 9, 734-737

657

658 Quilliam RS, Kinzelman J, Brunner J, Oliver DM (2015). Resolving conflicts in public  
659 health protection and ecosystem service provision at designated bathing  
660 waters. *Journal of Environmental Management* 161, 237–242

661

662 Russell, T.L., Sassoubre, L.M., Zhou, C., French-Owen, D., Hassaballah, A., and  
663 Boehm, A.B. (2014) Impacts of Beach Wrack Removal via Grooming on Surf Zone  
664 Water Quality. *Environmental Science and Technology*. **48(4)**: 2203-2211.

665

666 Saayman, M., & Saayman, A. (2017). How important are Blue Flag awards in beach  
667 choice? *Journal of Coastal Research* 33, 1436 – 1447

668

669 Schernewski, G., Baltranaitė, E., Kataržytė, M., Balčiūnas, A., Čerkasova, N., &  
670 Mėžinė, J. (2018). Establishing new bathing sites at the Curonian Lagoon coast: an  
671 ecological-social-economic assessment. *Journal of Coastal Conservation*, In press.

672

673 Searcy, R. T., Taggart, M., Gold, M., & Boehm, A. B. (2018). Implementation of an  
674 automated beach water quality nowcast system at ten California oceanic  
675 beaches. *Journal of Environmental Management*, 223, 633-643.

676

677 Smith, D.G., Cragg, A.M., and Croker, G.F. (1991) Water clarity criteria for bathing  
678 waters based on user perception. *Journal of Environmental Management*. **33(3)**:  
679 285-299.

680

681 Thoe, W., Gold, M., Griesbach, A., Grimmer, M., Taggart, M. L., & Boehm, A. B.  
682 (2014). Sunny with a chance of gastroenteritis: predicting swimmer risk at California  
683 beaches. *Environmental Science & Technology*, *49*(1), 423-431.

684

685 Wu J, Long SC, Das D, Dorner SM. (2011). Are microbial indicators and pathogens  
686 correlated? A statistical analysis of 40 years of research. *Journal of Water & Health*.  
687 *9*, 265–78.

688