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1 Effectiveness of Interventions to Reduce Household Air Pollution and/or Improve
2 Health in Homes using Solid Fuel in Low-and-Middle Income Countries: a
3 Systematic Review and Meta-analysis
4

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1 **Abstract**

2 **Background:** Cookstove intervention programs have been increasing over the past two (2) decades in Low
3 and Middle Income Countries (LMICs) across the globe. However, there remains uncertainty regarding the
4 effects of these interventions on household air pollution concentrations, personal exposure concentrations
5 and health outcomes.

6 **Objectives:** The primary objective was to determine if household air pollution (HAP) interventions were
7 associated with improved indoor air quality (IAQ) in households in LMICs. Given the potential impact of
8 HAP interventions on health, a secondary objective was to evaluate the effectiveness of HAP interventions to
9 improve health in populations receiving these interventions.

10 **Data sources:** OVID Medline, Ovid Embase, SCOPUS and PubMed were searched from their inception
11 until December 2015 with no restrictions on study design. The WHO Global database of household air
12 pollution measurements and Members' archives were also reviewed together with the reference lists of
13 identified reviews and relevant articles.

14 **Study eligibility criteria, participants and intervention:**

15 We considered randomized controlled trials, or non-randomized control trials, or before-and-after studies;
16 original studies; studies conducted in a LMIC (based on the United Nations Human Development Report
17 released in March 2013 (World Bank, 2013); interventions that were explicitly aimed at improving IAQ
18 and/or health from solid fuel use; studies published in a peer-reviewed journal or student theses or reports;
19 studies that reported on outcomes which was indicative of IAQ or/and health. There was no restriction on the
20 type of comparator (e.g. household receiving *plancha* vs. household using traditional cookstove) used in the
21 intervention study.

22 **Study appraisal and synthesis methods:** Five review authors independently used pre-designed data
23 collection forms to extract information from the original studies and assessed risk of bias using the Effective
24 Public Health Practice Project (EPHPP). We computed standardized weighted mean difference (SMD) using
25 random-effects models. Heterogeneity was computed using the Q and I²-statistics. We examined the
26 influence of various characteristics on the study-specific effect estimates by stratifying the analysis by
27 population type, study design, intervention type, and duration of exposure monitoring. The trim and fill
28 method was used to assess the potential impact of missing studies.

1 **Results:** Fifty-five studies met our *a priori* inclusion criteria and were included in the systematic review.
2 Fifteen studies provided 43 effect estimates for our meta-analysis. The largest improvement in HAP was
3 observed for average particulate matter (PM) (SMD=1.57) concentrations in household kitchens (1.03),
4 followed by daily personal average concentrations of PM (1.18), and carbon monoxide (CO) concentrations
5 in kitchens. With respect to personal PM, significant improvement was observed in studies of children (1.26)
6 and studies monitoring PM for ≥ 24 hrs (1.32). This observation was also noted in terms of studies of kitchen
7 concentrations of CO. A significant improvement was also observed for kitchen levels of PM in both adult
8 populations (1.56) and in RCT/cohort designs (1.59) involving replacing cookstoves without chimneys. Our
9 findings on health outcomes were inconclusive.

10 **Limitations, conclusions and implications of key findings:** We observed high statistical between
11 study variability in the study-specific estimate. Thus, care should be taken in concluding that HAP
12 interventions - as currently designed and implemented - support reductions in the average kitchen
13 and personal levels of PM and CO. Further, there is limited evidence that current stand-alone HAP
14 interventions yield any health benefits. Post-intervention levels of pollutants were generally still
15 greatly in excess of the relevant WHO guideline and thus a need to promote cleaner fuels in LMICs
16 to reduce HAP levels below the WHO guidelines.

17 **Systematic review registration number:** The review has been registered with PROSPERO
18 (registration number CRD42014009768)

19

20 **Keywords** Developing country, HAP, health improvement, intervention, meta-analysis, systematic
21 review

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1 **1. Introduction**

2 Nearly one-third of the world’s population use solid fuels such as wood, animal dung,
3 and crop residues as their primary source of domestic energy use (e.g. Bonjour et al., 2013;
4 Balakrishnan et al., 2013; Chafe et al., 2014). Cooking and heating with solid fuel on open fires or
5 traditional stoves emits a complex mixture of organic compounds and gases, which include carbon
6 monoxide (CO), oxides of nitrogen (NO_x) and sulphur (SO_x), aldehydes, polycyclic aromatic
7 hydrocarbons (PAHs), volatile organic compounds (VOCs), chlorinated dioxins, fine particulate
8 matter (PM), and free radicals (Albalak, 2001; Mishra, 2003). Health problems associated with
9 household air pollution (HAP) from solid fuel/biomass fuel use includes, but is not limited to,
10 respiratory tract infection (Mishra, 2003) exacerbation of inflammatory lung conditions (Gordon et
11 al., 2015), cardiac events (Bruce et al., 2015), asthma (Bruce et al., 2015; Gordon et al., 2015),
12 chronic obstructive pulmonary disease (COPD) (Assad et al., 2015; Bruce et al., 2015; Gordon et
13 al., 2015), low birth weight (Amegah et al., 2014) and tuberculosis (Kurmi et al., 2014).

14 Cookstove intervention programs have been implemented and studied extensively in
15 Low and Middle Income Countries (LMICs). However, there remains significant uncertainty
16 regarding the effectiveness of these interventions. Rehfuess et al. (2014) and Thomas et al. (2015)
17 recently reported reviews on this subject. Rehfuess et al. (2014) conducted a systematic review and
18 meta-analysis covering the period between 1998 and July 2012. The authors identified 38 studies
19 published in LMICs and ICs and noted reduction in average daily concentrations of the two most
20 commonly measured pollutants: PM and CO. Thomas et al. (2015) conducted a systematic review
21 of studies published up to April 2014. These authors captured almost the same studies previously
22 reviewed by Rehfuess et al. (2014) but the findings from these two reviews were contradictory. The
23 household air pollution field is changing rapidly and new evidence has accumulated since the last
24 review (Thomas et al. 2015). In such a rapidly evolving field there is the need to confirm or refute
25 previous findings. Also timely evaluation of methods and results of studies can help inform public

1 health policy and future studies. It is from these perspectives that we are conducting this systematic
2 review to determine if HAP interventions are associated with improved IAQ in households in
3 LMICs. A secondary objective is to evaluate the effectiveness of HAP interventions to improve
4 health in populations living in LMICs.

5

6 **2. Methods**

7 *2.1 Search strategy*

8 This systematic review was carried out according to established methods (NICE,
9 2012; IRIS, 2012) and reported according to recommendations from the Preferred Reporting Items
10 for Systematic Reviews and Meta-Analyses statement (PRISMA) (2014). A review protocol is
11 reported elsewhere (Quansah et al., 2015).

12 We performed a systematic literature search of OVID Medline, Ovid Embase,
13 SCOPUS and PubMed databases (Supplementary, search strategy) from their inception until
14 October 2013 and updated our search in December 2015 (Fig.1). Furthermore, the reference lists of
15 identified articles and that of a recent review (Thomas et al., 2015) were searched. Three authors
16 (RQ, SS, and CO) carried out the initial screening of titles and abstracts from the searches. Full
17 papers of potentially relevant publications were located and independently appraised by five
18 reviewers (RQ, CO, SS, FA and IL) to select those satisfying the inclusion criteria.

19

20 *Study selection*

21 ***Population of interest***

22 Studies had to be carried out in populations located within LMICs based on the United
23 Nations Human Development Report released in March 2013) (World Bank, 2013). There were no
24 other restrictions on population type.

25

1 ***Type of Intervention***

2 We considered any type of household intervention that was explicitly aimed at
3 improving indoor air quality and/or health by changing or reducing emissions from solid fuel use
4 within the home. Such interventions include for example, changes in stove or heating apparatus,
5 changes in ventilation arrangements and changes in behavior geared towards reducing emission and
6 exposure to cooking smoke. This allowed us to examine the influence of intervention type on study-
7 specific estimates. Interventions targeting for example, deforestation, fire wood use, particle size
8 distribution and cooking time were excluded because they did not address our research questions.

9

10 ***Type of comparisons***

11 Different types of comparators have been used in intervention studies. We aimed to
12 provide a comprehensive evaluation of the evidence and did not impose any restrictions on the type
13 of comparator used in the intervention studies (for example, convenience comparison group,
14 randomized control group, no intervention control, and usual practice control). Studies with and
15 without comparators were included in the review.

16

17 ***Types of outcomes***

18 We assessed both health and exposure outcomes. The primary outcomes were
19 measures of indoor air quality (IAQ), e.g. airborne concentrations of carbon monoxide or fine
20 particulate matter or soot or smoke. We also included biomarkers of exposure to air pollutants in the
21 form of metabolites of poly-aromatic hydrocarbons. Secondary outcome measures were common
22 health indicators and included (but again not be limited to) acute lower respiratory infection,
23 sensory irritation (for example, itchy/watery/sore eyes), cough and high blood pressure (Quansah et
24 al., 2015). Studies that only reported on fuel use, cooking time, climate, and non-IAQ/health related
25 outcomes were excluded.

1 ***Type of study***

2 We included randomized controlled trials (RCT or quasi-RCT), or non-randomized
3 control trials (i.e. cohort, case-controlled and cross-sectional studies), or before-and-after studies.
4 We excluded all controlled experimental studies (i.e. both laboratory and field) because they did not
5 qualify as interventions. We further excluded studies conducted in developed countries because
6 they did not answer our research questions.

7
8 ***Types of publications***

9 In order to provide a comprehensive review of the literature we considered both
10 articles in peer-reviewed journals and student theses. Our search included publications in the
11 following languages: English, Spanish, and Chinese.

12
13 ***2.2. Data extraction and risk of bias assessment***

14 Relevant characteristics of eligible studies were extracted and recorded independently
15 by five authors (i.e. RQ, CO, SS, FA and IL). Discrepancies were resolved through discussion.
16 TABLE 1 displays the main characteristics of the eligible studies. Risk of bias was assessed with
17 the Effective Public Health Practice Project Quality Assessment Tool (EPHPP) (1998). The results
18 of the risk of bias assessment tool for each individual study (**Supplemental Table 1**) and the data
19 extraction forms (**Supplemental Table 2**) are available in the supplementary materials.

20
21 ***2.3. Statistical analysis***

22 We anticipated substantial between study variability and we computed standardized
23 weighted mean difference (SMD) for each of our outcomes: personal levels of particulate matter (P-
24 PM) and carbon monoxide (PCO) and kitchen levels of particulate matter (MPM) and carbon
25 monoxide (MCO) using a random effects model. The studies presented mean differences and
26 corresponding exact p-values, or mean difference and their 95% confidence interval, or mean and

1 standard deviation. The mean difference and their corresponding standard error were computed in
2 excel using standard methods (Borenstein et al., 2009; Follmann et al., 1992; Higgins et al., 2010).
3 Heterogeneity was computed using the Q ($p < 0.1$ considered significant), and I^2 -statistics (I^2 -
4 statistic $> 50\%$ indicates high, 25–50% moderate, and $< 25\%$ low heterogeneity). We examined the
5 influence of various characteristics on the study-specific effect estimates by stratifying the analysis
6 by: a) population type (children vs. female adult vs. both children and female adult); b) study design
7 (cross-sectional vs. pre-post design vs. RCT); c) intervention type (*plancha* vs. *justa* vs. *patsari* vs.
8 other); d) duration of exposure monitoring (≤ 24 hours vs. > 24 hours). Publication bias was assessed
9 using the Egger test of asymmetry (Egger et al., 1998). The trim and fill method was used to assess
10 the potential impact of missing studies. Statistical analysis was performed using STATA software
11 version 11 (StataCorp, College Station, TX, USA).

12

13 **3. Results**

14 *3.1. Literature Search*

15 Our systematic search of the literature is shown in Figure 1. Fifty-five studies met our
16 *a priori* inclusion criteria and were included in the systematic review. Fifteen studies provided 43
17 effect estimates for our meta-analysis (Supplementary Table 3). Of the 55 studies, 46 were
18 identified from the searched databases, three were identified from reference lists of relevant studies,
19 with a further six retrieved from a recent review (Thomas et al., 2015). Sixty-eight studies were
20 excluded for reasons given in Supplementary Table 4.

21

22 *3.2. Study Characteristics*

23 Characteristics of the 55 eligible studies by study design are presented in Tables 1a
24 and 1b. The interventions identified were carried out in three continents: South America (Guatemala
25 (n=12); Honduras (n=2); Nicaragua (n=1); Peru (n=6); and Mexico (n=7)); Asia (Bangladesh (n=1);

1 China (n=8); India (n=4); Nepal (n=1); Pakistan (n=1); and Africa (Senegal (n=4); Ghana (n=1);
2 Nigeria (n=1); Kenya (n=4); South Africa (n=1); Malawi (n=1); and Rwanda (n=1)). Sixteen of the
3 intervention studies were cross-sectional designs, nineteen were before-and-after designs, eleven
4 were randomized control trials (RCTs), and eight cohort designs. In two studies, two
5 complementary study designs were applied. That is, between-group comparisons based on
6 randomized stove assignment, and before-and-after comparisons within control subjects who used
7 open fires during the trial and received chimney stoves after the trial. The interventions were carried
8 out among female adults population or among children or both population groups. Most of the
9 studies applied a single intervention and these included improved cookstoves, mostly wood burning
10 stoves such as *patsari* cookstove, *plancha* cookstove, improved *justa* stove, OPTIMA cookstove,
11 eco-cookstove, *sukhad* cookstove, ONIL stove, *gyapa* cookstove and “smoke free stove”. There was
12 one study carried out on the use of biogas digesters, three on solar ovens and one on a switch to
13 ethanol fuel. Other interventions assessed in the studies were: installation of chimneys (n=1); health
14 education campaign (n=1); and behavioural interventions such as education and counselling on
15 cooking outdoors, opening windows/doors and reducing the amount of time the child spends in the
16 kitchen (n=1). The funding model for the intervention was often poorly described. Some studies
17 indicated that the improved cookstoves were offered for free but most studies failed to report the
18 delivery mechanism and financing of the intervention.

19 Several outcomes were reported in the intervention studies and we classified them into
20 personal exposure outcomes and micro-environment exposure outcomes (Table 1a); and health
21 outcomes (Table 1b). Personal exposure outcomes were outcomes measured at the individual level
22 using for example, personal monitors attached to individuals clothing, measurement of metabolites
23 in urine and so on. Micro-environment exposure outcomes refer to particulate matter (PM) and/or
24 carbon monoxide (CO) measured from a fixed point in the home, most commonly in the kitchen.
25 Health outcomes reported in the identified studies were generally sparse and heterogeneous.

1 3.3. Personal Exposure Outcomes

2 Twenty-eight studies reported on personal indicators of indoor air quality (IAQ) and
3 are described below (Table 1a).

4 3.3.1 Studies on Personal Particulate Matter

5 Of the 11 studies that reported on daily average personal particulate matter (P-PM),
6 five did not provide sufficient data for our quantitative analysis and were therefore analyzed
7 qualitatively. Cynthia et al. (2008) studied the impact of improved wood burning stove (Patsari) in
8 reducing personal exposure to PM_{2.5} and CO in 60 homes in rural Michoacan in Mexico. The daily
9 average personal 24-hr PM_{2.5} was 0.29 mg/m³ for women using a traditional open fire. Installation
10 of the *Patsari* cookstove resulted in a 35% reduction in the median 24-hr personal PM_{2.5}. The
11 corresponding reduction in 48-hr personal CO exposure was 77%. Li et al. (2011) investigated
12 whether replacement of open pit stoves by improved stoves equipped with a chimney reduced
13 exposure to PAHs, PM_{2.5} and CO exposure in rural Peru. Two stove types were evaluated (A, n=30;
14 B, n=27). Installation of improved cookstoves reduced personal exposures by 47-74% and urinary
15 hydroxylate PAH metabolites (OH-PAHs) by 19-52%. Mukhopadhyay et al. (2012) assessed two
16 brands of commercial advanced cookstoves (i.e. Philips and Oorja) with small blowers to improve
17 combustion. These advanced cookstoves produced reductions in personal PM_{2.5} and CO.

18 In total, seven effect estimates were provided by five studies for our quantitative
19 analysis of P-PM (Supplementary Table 1). Of this, Naeher et al. (2000) provided data on both the
20 populations of children and their mothers. Fitzgerald et al. (2012) used two different improved
21 cookstoves, referred to in the forest plot as ICS1 and ICS2. The overall summary SMD was 1.18
22 (95% CI: 1.05, 1.32) (Fig 1). High between-study variability was observed (I²-index=89.7%,
23 P=0.000, Q-statistics (n=7) =58.17). In the stratified analysis, moderate to large improvements in
24 average daily personal PM was observed (Fig. 2 and Table 1). With the exception of studies on
25 *plancha* cookstoves and those studies monitoring personal PM levels for less than 24-hr,

1 heterogeneity persisted (Table 2). The Egger test of small study effects showed no evidence of
2 publication bias ($p=0.123$). However, adjustment for publication bias by the trim and fill method
3 imputed three studies and the overall summary SMD was reduced marginally (1.08; 95% CI: 0.95,
4 1.22) and heterogeneity persisted (91.66 (10), $p=0.00$, 99.2%) (Table 2).

5

6 3.3.2. *Studies on Personal Carbon monoxide*

7 Twenty-one intervention studies reported daily average personal carbon monoxide
8 levels (PCO). Clark et al. (2013) examined the impact of a cleaner-burning cookstove intervention
9 among non-smoking female cooks and measured indoor $PM_{2.5}$, CO and PCO concentrations. Large
10 mean reduction concentrations were observed for all exposure metrics following installation of a
11 subsidized eco-stove. Diaz et al. (2007) observed a median reduction in exhaled breath CO in the
12 intervention group compared to the control group following the installation of improved *plancha*
13 stoves. Rollin et al. (2004) conducted a feasibility study to assess the impact of reduction of IAQ on
14 acute lower respiratory infection in infants. When mean concentrations of CO were compared
15 between electrified and un-electrified dwellings, there was strong evidence ($p=0.0004$) that the
16 mean concentrations of log (CO) in the kitchen was higher in the electrified areas (1.25 vs. 0.69)
17 and strong evidence ($p<0.0001$) that the mean concentration of log (CO) on the child was higher in
18 electrified areas (0.83 s. 0.34). Beltramo et al. (2013) did not observe any evidence that solar ovens
19 reduced exposure to carbon monoxide.

20 Altogether seven studies provided 10 effect estimates for our quantitative analysis of
21 PCO (Supplemental Material, Table S1). The overall summary SMD was 0.81 (95% CI: 0.63, 1.05)
22 and substantial heterogeneity was noted ($I^2=99.8\%$, $p=0.00$, Q-statistics (11) =5089.79). Slight to
23 moderate improvement in IAQ related to PCO was observed across study level characteristics (Fig.
24 3 and Table 2) and substantial heterogeneity persisted. Egger small study effect ($p=0.415$) and the
25 trim and fill did not show any evidence of publication bias (Table 2).

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3.2. *Micro-environment Exposure Outcomes*

A total of 26 studies reported micro-environment indicators of HAP from solid fuel use and are described below (Table 1a).

3.2.1. *Studies on Micro-environment Particulate Matter*

With respect to particulate matter (PM), 10 studies provided 13 effect estimates for our quantitative analysis. Masera et al. (2007) evaluated the impact of improved *patsari* cookstoves in the Purepecha region of Michoacin state in Mexico. Average concentrations of CO and PM_{2.5} were measured both before and after the introduction of the improved cookstove at 1- minute intervals for 48 hrs. PM_{2.5} and CO were reduced by 67% and 66% respectively. In the study by Chowdury et al. (2013), 24-hr PM_{2.5} and CO concentrations in the kitchen ranged between 0.15-0.71 ppm for PM_{2.5} and 3.0-11 ppm for CO when using the traditional cookstove; and between 0.08-0.18 ppm for PM_{2.5} and 0.7-5.5 ppm for CO when using improved cookstoves. In three Andean communities within the Santiago de Chuco province of Peru, two different models of improved cookstoves (i.e. stove 1 and stove 2) were installed in 64 homes. In the community receiving stove 1, baseline 48-hr personal exposure and kitchen concentrations of PM_{2.5} were 116.4 and 207.3 µg/m³, respectively; and 48-hr hour personal and kitchen CO levels were 1.2 and 3.6 ppm respectively. After introducing the new stove to this community, personal exposure and kitchen concentrations of PM_{2.5} reduced to 68.4 and 84.7 µg/m³ respectively; and that of personal and kitchen CO levels to 0.4 and 0.8 ppm respectively, representing reductions of 41.3%, 59.2%, 69.6% and 77.7%. In the two communities receiving stove 2, corresponding levels were 126.3 µg/m³, 173.4 µg/m³, 0.9 ppm , and 2.6 ppm before the installation of the stoves, and they reduced to 58.3, 51.1 µg/m³ and 0.6, 1.0 ppm. Overall, homes receiving stove 2 saw reductions of 53.8, 70.5, 25.8 and 63.6%.

1 In the meta-analysis of 13 effect estimates from 10 studies (Supplementary Table 3),
2 the overall summary SMD was 1.57 (1.22, 2.01). High statistical heterogeneity was observed
3 ($I^2=98.2\%$, $p=0.00$, Q-statistics (13) =661.63). A slight to large improvement in average kitchen
4 levels of PM (MPM) was noted across study level characteristics and with the exception of studies
5 looking at the exposure of children substantial between-study variability persisted (Table 2). A test
6 of publication bias showed no evidence of small study effects ($p=0.184$). Again this was not
7 confirmed by the trim and fill method which showed a decline in the summary SMD (1.16; 0.85,
8 1.53) (Table 2).

9

10 3.2.2. *Micro-environment Carbon Monoxide*

11 In the study by Chengappaa et al. (2007) in the Bundelkund region of India, CO was
12 measured for a 48-hr period in 60 rural kitchens before and after installation of a *sukhad* improved
13 cookstove. One year after the intervention, CO concentrations were reduced by 70% ($p<0.001$) and
14 44% ($p<0.01$) respectively. Khushk et al. (2005) reported levels of 15.4 ppm for smoke-free stoves
15 compared to 28.5 ppm for traditional cookstoves. A 3-stage risk reduction program was applied by
16 Torres-Dorsal et al (2007). These steps included (i) removal of indoor soot adhered to roofs and
17 internal walls (ii) paving dirty floors and (iii) introduction of a new wood stove with a metal
18 chimney. Blood caroxyhaemoglobin (% COHb) and urinary 1-OHP levels were measured before
19 and after the intervention. In the 20 participants the levels of COHb reduced by an average of 2.5%
20 one month after the intervention. A similar observation was noted for 1-OHP levels.

21 In all, eleven studies provided 13 effect estimates for our quantitative analysis of
22 kitchen levels of CO. The overall summary SMD was 1.21 (0.89, 1.66; $I^2=99.5\%$, $p=0.00$, Q-
23 statistics (13) =2578.71) (Fig 4). A slight to large improvement in indoor air quality (IAQ) related
24 to average kitchen levels of CO (MCO) was noted across study level characteristics and, except for
25 studies scoring weak on the Effective Public Health Practice Project Quality Assessment Tool

1 (EPHPP), substantial heterogeneity persisted. There was no evidence of the small study effect
2 ($p=0.154$) and this was confirmed by the trim and fill method (1.03; 0.76, 1.41) (Table 2).

3

4 3.3. Health Outcomes

5 A total of twenty-nine studies reported health outcomes. Of these, 10 studies reported
6 on respiratory health problems alone, 10 studies on non-respiratory health problems, and 8 studies
7 on both respiratory and non-respiratory health problems (Table 1b). Due to the sparse nature of
8 individual health outcomes, it was not possible to conduct a meta-analysis. These studies are
9 discussed below.

10

11 3.3.1. Respiratory Health Problems

12 Different definitions of asthma were applied in the studies and this included asthma based on
13 lung function measurements and self-reported symptoms of asthma such as wheezing, cough,
14 phlegm production, difficulty breathing, runny or stuffy nose and chest tightness. Beltramo and
15 Levine (2013) compared respiratory symptoms in 465 women who purchased solar ovens and 325
16 control women. The authors did not observe any evidence that the use of solar ovens reduced the
17 incidence of cough and/or sore throat. They concluded that their study was a policy success because
18 it halted a nationwide proposal to roll-out solar ovens. Women who reported using *patsari*
19 cookstoves most of the time compared to those using open fire experienced significantly lower
20 levels of cough and wheezing (Romieu et al., 2009). Significant reductions in several respiratory
21 symptoms such as dry cough, chest tightness, difficulty breathing and runny nose were observed in
22 mothers and children in homes that used improved cookstoves compared to those that used
23 traditional cookstoves in Guatemala (Albalak et al., 2001). These findings were confirmed by
24 Ludwinski et al. (2011) who reported a 48.6% and a 63.3% reduction in respiratory symptoms in
25 mothers and children respectively, and in Romieu et al. (2009) and Dohoo et al. (2012). However,

1 no significant improvement in measures of lung function was observed in users of *plancha* (Smith-
2 Sivertsen et al., 2009), *justa* (Clark et al., 2009) and *Gram vikas* (Hanna et al., 2012) cookstoves. In
3 a parallel randomized controlled trial in Guatemala, Smith et al. (2011) investigated whether an
4 intervention to lower indoor wood smoke emissions would reduce pneumonia in children.
5 Pneumonia was defined as physician-diagnosed pneumonia, without use of a chest radiograph or
6 fieldworker-assessed pneumonia (all and severe) and seven other conditions of physician-diagnosed
7 pneumonia. Significant reductions in the intervention group for three severe outcomes: fieldworker-
8 assessed, physician-diagnosed, and RSV-negative pneumonia were noted. In the exposure-response
9 analysis, a 50% exposure reduction was significantly associated with a reduction in physician-
10 diagnosed pneumonia (RR 0.82; 0.70, 0.98). Hosgood et al. (2002) evaluated lung cancer mortality
11 reduction after changing from a traditional smoky stove to an improved cookstove in China. A
12 significant reduction in lung cancer mortality was observed in women and men who changed to
13 improved cookstoves compared to those who did not change. Reductions in lung cancer incidence
14 (Lan et al. 2002) and pneumonia mortality (Shen et al. 2009) were also observed in similar
15 populations following a switch to an improved cookstove .

16

17 3.3.2. Non-respiratory Health Problems

18 Household air pollution (HAP) intervention studies on non-respiratory symptoms have
19 been inconclusive. Whereas Jary et al. (2014) did not observe any significant improvement in
20 headache and back pain following the use of wood burning clay improved cookstove for 7 days,
21 Burwen and Levine (2012), Diaz et al. (2008) and Alam et al. (2006) noted marginal to significant
22 improvements in these indicators. A small number of studies have observed improvements in blood
23 pressure (Hanna et al., 2002; McCracken et al., 2007) and low birth weight (Thompson et al., 2011)
24 following the use of improved cookstoves.

25

1 **4. Discussion**

2 This systematic review and meta-analysis of HAP interventions conducted in Low and
3 Middle Income Countries (LMICs) aims to address the question, whether HAP interventions to
4 improve indoor air quality (IAQ) and/or health in homes using solid fuel for cooking and heating
5 are effective. Fifteen of the 55 studies were eligible for quantitative analysis. The largest reduction
6 in HAP was particulate matter (PM) levels in the kitchen, followed by daily personal average levels
7 of PM, levels of carbon monoxide (CO) in the kitchen and daily personal average levels of CO.
8 Slight to large improvement related to study level characteristics in average kitchen levels of PM
9 and CO as well as average daily levels of PM and CO. The findings from the qualitative analysis
10 corroborate that of the quantitative analysis. Findings on health outcomes were inconclusive. We
11 also observed high statistical between study variability in the study-specific estimates and this
12 persisted in most cases in the stratified analysis. Thus, caution is warranted in concluding that HAP
13 interventions - as currently designed and implemented – support reductions in the average kitchen
14 and personal levels of PM and CO.

15

16 *4.1. Validity issues*

17 Our study has a number of strengths. We searched several databases and used
18 secondary references that were cited in the original articles and a recent review. Five reviewers
19 independently assessed the articles based on *a priori* eligibility criteria. We also followed the
20 methods of the National Institute of Healthcare Excellence (NICE) and the National Academy of
21 Science review of the EPA Integrated Risk Information System process; and reported the findings
22 according to recommendations by the Preferred Reporting Items for Systematic Reviews and Meta-
23 analysis statement. We also evaluated the possibility of publication bias using Egger’s test of
24 asymmetry and the trim and fill method.

1 We acknowledge a number of limitations in our study. We applied the Effective
2 Public Health Practice Project Quality Assessment Tool (EPHPP) to assess the risk of bias, but due
3 to a lack of data we were unable to explore the influence of each of the six domains on our overall
4 summary SMD. Nevertheless, we applied a single global rating score to assess the influence of risk
5 of bias on overall summary SMD and the majority of the studies were rated as weak on this scale.
6 How well a study fairs on the scale is dependent on the amount of information available in the
7 article for evaluation. Thus, a well conducted study may score poorly on the scale because the
8 author(s) failed to provide adequate information when writing up their manuscript. As a result,
9 interpretation on how well a study does on the EPHPP should be carried out with caution. We also
10 observed high statistical between study variability in the study-specific estimates and this persisted
11 in most cases in the stratified analysis. Thus, caution may be warranted in concluding that HAP
12 interventions as currently designed and implemented support reductions in the average kitchen and
13 personal levels of PM and CO.

14 Furthermore, of the studies meeting our *a priori* inclusion criteria, the majority came
15 from cross-sectional studies in which it was not possible to judge the temporal relation between
16 HAP intervention and our outcomes of interest. However, the generally consistent results from the
17 different study designs and study level characteristics support the hypothesis that HAP interventions
18 reduce average kitchen levels of PM/CO and daily average personal levels of PM/CO. A variety of
19 exposure outcomes and interventions were reported in the original studies. With respect to the
20 outcomes we carefully segregated them into average daily PM/CO personal levels and levels of
21 PM/CO in the kitchen and this allowed us to study the impact of HAP intervention on each
22 outcome. Due to practical, ethical and budgetary constraints the studies have typically been small in
23 size with a very small proportion of them being RCT design. The monitoring period tended to be
24 short and the duration of measurement was inconsistent across the studies making comparison
25 difficult. We also carefully categorized the monitoring period into 2 levels (i.e. <24hrs vs ≥24hrs) to

1 understand how daily exposure variability impacts on the overall SMD. Most of the studies also had
2 qualitative components and these did not contribute to the evaluation of the impact of HAP
3 intervention on our outcome of interest. The studies generally differed on baseline indoor PM/CO
4 and in most cases post-intervention measurements were above the WHO guideline.

5

6 *4.2. Comparison with previous studies*

7 Two previous reviews: a WHO report (Rehfuess et al., 2014) and a systematic review
8 by Thomas et al. (2015) were available on this subject. In the first study, Rehfuess et al. (2014)
9 conducted a systematic review and meta-analysis to answer the question whether improved
10 cookstoves in everyday use, compared to traditional cookstoves are effective in reducing average
11 concentrations of, or exposure to, particulate matter and carbon monoxide among households in.
12 The authors identified 38 studies and reported average kitchen levels of PM and CO reducing by 38
13 to 82%. Reductions of average personal levels of PM and CO ranging between 47 to 76% were also
14 observed. In the second study, Thomas et al. (2015) conducted a qualitative review of
15 epidemiologic evidence that improved cookstove interventions reduced household air pollution and
16 improved health. They searched 10 databases and identified almost the same studies included in
17 Rehfuess et al. (2014) but failed to reach any conclusion. .

18 We identified **55** studies, and across these household air pollution (HAP) interventions
19 resulted in improvement in daily average personal and average kitchen levels of particulate matter
20 by 24% and 18% respectively. A much smaller (3%) improvement in average kitchen levels of
21 carbon monoxide was observed.

22 This observation was also observed across different study designs, population types
23 and monitoring duration. Our findings are consistent with the WHO report (Rehfuess et al., 2014),
24 although ours is more comprehensive as we include 17 additional studies. We observed that the
25 performance of the *Plancha* and the *Justa* on average daily personal levels of PM was broadly

1 similar. However, both the *Justa* and the *Patsari* outperformed the *Plancha* in reducing average
2 kitchen levels of PM/CO. The reason for this later observation is not clear, but the former
3 observation corroborates the finding of a WHO report (Ruhfuess et al., 2014) which suggests a
4 substantial improvement of personal PM levels with stoves with chimneys. In spite of the observed
5 improvement in average PM and CO levels, post-intervention concentrations of PM/CO are much
6 higher than the WHO guidelines for these air pollutants.

7 Our findings have important public health implications for populations living in Low
8 and middle-income countries (LMICs) given the tremendous health burden of household air
9 pollution (HAP) (Amegah and Jaakkola, 2015) and the fact that the total number of users of
10 solid/biomass fuels in these countries have not declined (Bonjour et al., 2013). Our findings further
11 suggest that current stand-alone HAP interventions yield little if any health benefit. Thus, there is
12 the need to re-examine the ways in which interventions are designed and implemented in homes in
13 LMICs. Multi-faceted HAP interventions could offer an opportunity to reduce exposures to HAP
14 that could have marked public health impact in LMICs. Cleaner fuels such as liquid petroleum gas,
15 ethanol, solar and electrification may reduce indoor emission levels substantially. However, to date
16 only five qualitative studies (1 on LPG, 2 on ethanol, 1 on electrification and 1 on solar) have
17 evaluated the impact of cleaner fuel on HAP or health. An ongoing HAP randomized trial in Ghana
18 may shed light on the impact of cooking with cleaner fuels on health/IAQ (Jack et al., 2015).

19 Our findings on health outcomes were inconclusive, supporting the findings of a
20 previous systematic review (Thomas et al., 2015) and recent large intervention studies (Bensch and
21 Peters, 2012; Hanna et al., 2012). The exposure-response (E-R) curve for most health effects is
22 poorly understood but data for cardio-pulmonary and cardiovascular effects from outdoor air
23 pollution, second-hand tobacco smoke and smoking studies tend to suggest that the curve, at least
24 for PM, is steep and may plateau at moderate PM doses (Pope, 2009). Such an E-R curve would
25 then suggest that minimal or no health improvements may accrue from the moderate HAP

1 improvements seen by some of the studies reported in this review especially when the
2 improvements achieved are within the upper, plateau part of the curve. Behavioral change
3 interventions have the potential to reduce average concentrations of particulate matter and carbon
4 monoxide by 20-98% in laboratory settings and by 31-94% in field setting (Barnes, 2014).
5 Behavioral strategies may include cooking outdoors, reducing time spent in the cooking area,
6 keeping the kitchen door/windows open while cooking, avoiding leaning over the fire while cooking,
7 avoiding carrying children while cooking and keeping children away from the cooking area.
8 Opportunities to educate communities on reducing household air pollution exposure include
9 durbars, festival celebrations, religious meetings and child welfare outreach clinics. Community
10 health workers are the fulcrum of the health system in many developing countries and represent
11 important change agents (Leon et al., 2015). Adoption and sustainability are big issues for most
12 HAP interventions. There is a need for qualitative research on the barriers and facilitators for
13 adoption and continued use. There is also a clear need for a standard method to evaluate HAP
14 interventions. A Standard Operating Procedure for stove evaluation programs that included sections
15 on HAP assessment, health assessment and intervention adoption/use assessment is required for the
16 HAP research community.

17

18 **5. Conclusions**

19 Our study suggests improvement in average PM/CO concentrations at the personal
20 and micro-environment level occur following HAP interventions. Despite this, post-intervention
21 levels are generally still far above the WHO guidelines for PM and CO, and perhaps because of
22 these continued high exposures, the study finds little evidence of improvements in health outcomes.
23 There is a need to develop effective interventions in LMICs that are capable of reducing HAP levels
24 below the WHO guidelines, particularly in many communities in the developing world where
25 adoption of improved cookstoves continues to prove challenging. There is also a need for a

1 Standard Operating Procedure for stove evaluation programs that include sections on HAP
2 assessment, health assessment and intervention adoption/use assessment for the HAP research
3 community.

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