

This is the peer reviewed version of the following article: Chiti, T., Mihindou, V., Jeffery, K. J., Malhi, Y., De Oliveira, F. L., White, L. J. T. and Valentini, R. (2017), Impact of woody encroachment on soil organic carbon storage in the Lopé National Park, Gabon. *Biotropica*, 49: 9–12, which has been published in final form at <https://doi.org/10.1111/btp.12369>. This article may be used for non-commercial purposes in accordance With Wiley Terms and Conditions for self-archiving.

Insight

LRH: Chiti et al.

RRH: Natural forest succession in the tropics

Impact Of Tropical Forest Succession On Soil Organic Carbon Storage In The Lopé National Park, Gabon

Chiti T^{1,2*}, Mihindou V^{3,4}, Jeffery KJ^{4,5,6}, Malhi Y⁷, Oliveira F⁸, White LJ^{4,5,6}, Valentini R^{1,2}

¹*Dept. for Innovation in Biological, Agro-food and Forest system (DIBAF), University of Tuscia, via San C. De Lellis s.n.c., 01100 Viterbo, Italy*

²*Euromediterranean Center on Climate Change (CMCC), Via Augusto Imperatore 16, 73100 Lecce, Italy*

³*Agence Nationale des Parcs Nationaux, Libreville, BP 20379 Gabon;*

⁴*Ministère de la Forêt, de l'environnement et de la Protection des Ressources Naturelles*

⁵*School of Natural Sciences, University of Stirling, FK9 4LA, Scotland, UK;*

⁶*Institut de Recherche en Écologie Tropicale, CENAREST, Libreville, Gabon;*

⁷*Environmental Change Institute, School of Geography and the Environment, University of Oxford, Oxford, UK;*

⁸*Instituto Nacional de Pesquisas da Amazônia – INPA*

*Corresponding author: Phone +39 0761 357394, Fax: +39 0761 357389

E-mail: tommaso.chiti@unitus.it

Received _____; revision accepted _____

1 **ABSTRACT**

2 This study quantifies the soil organic carbon (SOC) stock changes occurring during different phases
3 of forest colonization on savannahs. SOC stock changes occur below 30 cm depth, indicating the
4 subsoil as the principal compartment contributing to SOC sequestration, and suggesting the need to
5 consider the entire profile (0-100 cm) to thoroughly assess the effect of woody encroachment on
6 SOC stock.

7
8 *Key words:* Marantaceae forests; Natural succession; Savannahs; Soil carbon stock; Tropical forests

9
10 WOODY PLANT ENCROACHMENT INTO SAVANNAHS AND GRASSLANDS REPRESENTS A
11 SIGNIFICANT GLOBAL CHANGE PHENOMENON (Archer *et al.* 2001), and is commonplace on the
12 African continent, with a considerable impact on carbon (C) dynamics at an ecosystem level
13 (Mitchard *et al.* 2013, Buitenwerf *et al.* 2012, Angassa *et al.* 2010, Mitchard *et al.* 2009, Spichiger
14 & Pamard 1973). Trees and shrubs influence the spatial distribution and cycling of nutrients by
15 altering soil respiration rates (Raich & Schlesinger 1992), hydrology (Wilcox 2002), microclimate
16 (Hoffman & Jackson 2000) and by concentrating soil organic carbon (SOC) beneath their canopies
17 (Binkley & Giardina 1998). Understanding SOC changes in relation to woody encroachment is
18 becoming increasingly important, given the worldwide occurrence of this phenomenon and the
19 growing interest in the C balance of the tropics (Aragão *et al.* 2014, Grace *et al.* 2014, Valentini *et*
20 *al.* 2014).

21 In this study, we aimed to quantify SOC levels in different soil layers down to 1 m depth
22 across the different phases of a natural succession from savannah to forest. We hypothesized that
23 SOC levels increase during the transition from savannah to forest due to both the increase in C
24 inputs to soil via litter deposition and the presence of herbaceous vegetation during the different

1 phases of natural succession. Evaluating SOC changes at different depths, down to 1 m, is of
2 outmost importance since often SOC changes are reported only for the upper soil layers, those
3 where the changes should be easily detectable (e.g. 0-30 cm depth), while below 30 cm depth
4 changes are often considered to be negligible.

5 The Lopé National Park in central Gabon (00° 12' 04" S; 11° 36' 05" E), offers a unique
6 opportunity for investigating SOC variations as a result of the natural expansion of forests into
7 savannahs. The park is part of the Congo Basin lowland forests and it is principally covered by
8 closed canopy rainforests with savannah systems in the north and east that are interspersed with
9 natural forest fragments, marshes and gallery forests (Fig. S1). The main forest formations resulting
10 from woody encroachment are the Marantaceae forests, characterized by a thick layer of herbaceous
11 plants of the Marantaceae and Zingiberaceae families that dominates the understory. The landscape
12 is characterized by hills with an average elevation of 500 m. The geological substrate is represented
13 by rocks of the Metamorphic series of the Ogooué river from the Proterozoic era (Schulter 2008),
14 and the typical soils of the area are comprised within the order of the Oxisols (Martin *et al.* 1981).

15 Soil samples were collected in five successional phases of forest colonization, as defined in
16 White (2001): 1) savannah (S); 2) colonising forest (CF); 3) monodominant forest (MF); 4) young
17 Marantaceae forest (YMF) and 5) mixed Marantaceae forest (MMF). Five 1-hectare plots in each of
18 the five phases were investigated. Basic characteristics of the vegetation of each plot, together with
19 site features and locations are reported in Table S1. In each of the plots five sampling points were
20 randomly selected and soil samples were collected at multiple depths. In the topsoil (0-30 cm depth)
21 the samples were collected at 0-5, 5-10, 10-20 and 20-30 cm depth using a cylinder of known
22 volume (diameter = 5 cm; height = 5 cm) to also determine the bulk density (BD), while for the
23 subsoil (30-100 cm depth) samples were collected within two intervals, 30-50 cm and 50-100 cm
24 depth, using an auger. The SOC data were reported for the topsoil, and subsoil according to IPCC
25 guidelines requirements (IPCC 2006). Five samples of the organic horizon, namely the litter layer,
26 were collected randomly in each plot using a 20 cm x 20 cm frame.

1 All samples were oven-dried (60 °C) and sieved at 2 mm, while the litter layer was ground
2 in a ball mill. All the analyses were performed on the fine soil fraction (< 2 mm). A single
3 composite sample per layer (n=5) was analysed for particle size distribution (Mikutta *et al.* 2005)
4 and pH. All samples were individually (n=5 per layer) measured for total C and N by dry
5 combustion (ThermoFinnigan Flash EA112 CHN). Since the auger did not allow for a precise
6 measure of the BD for the 30-50 and 50-100 cm depths, BD was determined using the clod method
7 (Blake & Hartge 1986) on three soil peds from each subsoil layer at each site. The SOC stocks were
8 calculated considering the C concentration, the BD, the rock fragment content and the depth of each
9 layer (Boone *et al.* 1999). Differences in C concentrations and stocks were determined among all
10 the phases of the succession and within each phase using analysis of variance (ANOVA) with depth
11 treated as a repeated measure using R software. When significant interaction effects were observed
12 a multiple comparisons test (Tukey HSD) was completed. Statistical significance for ANOVA were
13 tested at $p < 0.001$.

14 Bulk density did not differ among the phases of the natural succession in the topsoil and
15 subsoil, clustering around 1.3 Mg/m³, and 1.4-1.5 Mg/m³, respectively (Table S2). The similarity of
16 the particle size distribution and pH between the phases supports sites comparability indicating
17 similar soil conditions in the different plots (Table S2), with the soil always classified as Inceptic
18 Hapludox (Soil Survey Staff 2014). Within each soil depth there was very little SOC variation
19 among plots within successional phases, but significant differences were observed between
20 successional phases (Table 1; Table S3). Particularly, in the 0-5 cm depth the SOC concentration
21 increases progressively from 12.5±0.7 g C kg⁻¹ in the S to 24.8±0.2 g C kg⁻¹ in the MMF (Table 1).
22 In the other layers, a SOC decline is evident below 30 cm depth in the S sites, 2.8±0.2 g C kg⁻¹,
23 compared to all the other phases, ~3.8-4.8 g C kg⁻¹ (Table 1).

24 The SOC stock variations can be entirely attributable to variations in SOC concentration,
25 due to the non-significant variations in BD between the successional phases, and the absence of
26 rock fragments in these soils (Table S2). In the topsoil, the SOC stocks do not show a significant

1 increase in the different phases of the natural succession, with values around 30 Mg/ha, and only in
2 the last phase the SOC stock is significantly higher than all the other phases (MMF = 38.6 ± 1.7
3 Mg/ha; Fig. 1). In the subsoil SOC stock increases significant in the first phase (CF = 39.5 ± 2.2
4 Mg/ha), then it remains stable to increases again in the last phase (Fig. 1). The C stock of the litter
5 layer, absent in the savannah plots, increases across the succession: 3.6 ± 1.5 Mg/ha in the CF,
6 4.8 ± 1.9 Mg/ha in the MF, 4.8 ± 1.7 Mg/ha in the YMF and $6.8.1 \pm 3.4$ Mg/ha in the MMF.

7 The rate or direction of the SOC changes occurring after woody plant establishment are not
8 consistent in the scientific literature. Some studies indicate SOC increases as a consequence of the
9 increase in the soil nutrient pool (Scholes & Archer 1997, Archer 1995), while other studies
10 indicate a significant SOC decline with increasing in precipitation (Jackson *et al.* 2002, Scott *et al.*
11 1999). After woody plant invasion on herbaceous vegetation a negative relationship between
12 precipitation and SOC changes was shown by Jackson *et al.* (2002) and Guo & Gifford (2002), with
13 the SOC levels decreasing in areas of high precipitation (>1200 mm), and increasing in areas with
14 low precipitation. Accordingly, in the Lopè National park, with an average annual precipitation of
15 about 1500 mm, woody encroachment should cause a decline in SOC levels. Nevertheless, in the
16 topsoil of all the phases SOC levels are stable and increase only in the last phase, while in the
17 subsoil the changes in SOC levels occur already in the first phase.

18 The main driver of the SOC increases in the topsoil of the last phase is possibly the increase
19 in litter inputs, leading to a progressive C accumulation in the 0-5 cm layer of the different phases.
20 In forests litter production is a significant fraction of the net primary productivity, rather than
21 stemwood biomass production, hence representing an important C input to soil (Brown & Lugo
22 1990). However, despite an increases in litter C should cause an increases in topsoil C (Guo &
23 Gifford 2002), the increasing litter inputs along the succession do not affect consistently the change
24 in topsoil C, apart in the last phase where the SOC concentration and stock increase of about 1
25 percent and 8.5 Mg/ha compared to savannahs. This fact suggests that in the long term there is a
26 positive impact of litter inputs on SOC sequestration and can help to explain the increase in SOC

1 levels only in the last phase of the succession. The effect of the litter in increasing the SOC
2 concentration is the result of a higher recalcitrance of the C inputs from woody species rather than
3 herbaceous vegetation (Marín-Spiotta *et al.* 2008). Woody plants produce lipids, such as waxes,
4 suberin, cutin, and terpenoids, that are resistant to oxidation and consumption, as protection against
5 herbivory and parasitism (Gleixner *et al.* 2001). The production of these and other plant secondary
6 compounds is thought to increase during tropical forest succession (Coley & Barone 1996).

7 Contrarily, compared to savannah the SOC levels in the subsoil increase significantly in the
8 first phase of succession (+ 16.4 Mg/ha) and again in the last phase (+ 29.9 Mg/ha) leading to
9 hypothesize a significant role of the root apparatus in the translocation of the SOC in deeper layers.
10 In fact, while herbaceous vegetation has the root apparatus concentrated in the topsoil, roots from
11 woody vegetation can go much deeper. The SOC increases in the subsoil could be also related to the
12 increases in soil cover due to the increased number of trees that protect the soil from heavy rains
13 and reduce the loss of SOC as dissolved organic carbon.

14 Considering the whole soil profile, the SOC stock in the last phase of the succession
15 increases by 43 percent compared to the savannah, and increases to 52 percent if we also account
16 for the C stored in the litter layer. In Amazonia San Jose *et al.* (1988) observed a similar trend
17 during woody encroachment on savannah, with significant SOC increases only when the forest
18 became established, while in the intermediate phases the SOC was rather stable. On the other hand,
19 other studies report a decline in SOC following encroachment (Don *et al.* 2011). These contrasting
20 results support the evidence that SOC can react differently to the events depending on the
21 geographic locations of the investigated area, suggesting that no general rules can be derived for
22 extrapolating general data that describe the effect of this natural land use change (Powers *et al.*
23 2011).The impact of the secondary forest vegetation in increasing the SOC levels is observable
24 immediately only in the subsoil, while in the topsoil the changes are evident only on the long term.
25 Different mechanisms are possibly responsible for this different behaviour between the phases, with
26 increasing litter inputs influencing the SOC accumulation in the topsoil and a different roots

1 distribution being responsible for the SOC accumulation in the subsoil. Finally, a crucial point is the
2 time needed for the natural transition to occur. The process of natural succession can be relatively
3 fast in the initial stages (e.g. years or few decades) but can take several centuries to reach forest
4 maturity, as suggested by Saldarriaga *et al.* (1988) in a study across a range of soil types.

5 In conclusion, this study points out the importance of measuring the SOC along the whole
6 profile (0-100 cm depth) to avoid overlooking the great amount of C that can accumulate in the
7 subsoil during natural succession to forest.

8

9

10 **ACKNOWLEDGEMENTS**

11 This paper is dedicated to the memory of Fabiane Lima de Oliveira. This work was funded by the
12 Gordon and Betty Moore Foundation, the David and Lucille Packard Foundation and by the ERC
13 grant GHG Africa (no. 247349). We also thank the Direction Générale de l'Environnement, the
14 Agence Nationale des Parcs Nationaux, the Centre Internationale de Recherche Medicales de
15 Franceville, and the staff of the Station d'Etudes des Gorilles et Chimpanzés, particularly Dibakou
16 J.

17

18

19

20

21

22

23

24

25

26

1 **SUPPORTING INFORMATION**

2 Additional Supporting Information may be found with online material:

3 **FIGURE S1.** a) Map of the Lopé National Park. The sampling area where the different plots are
4 located is represented by the square; b) the position of the Park within Gabon and; c) Gabon in the
5 African continent.

6 **TABLE S1.** *Location, altitude and dominant vegetation of each of the investigated plots.*

7 **TABLE S2.** *Bulk density, particle size distribution and pH for the topsoil (0-30 cm) and subsoil
8 (30-100 cm) of the twenty-five plots describing the natural succession. * S = Savannah; CF =
9 colonizing forest; MF = Monodominant forest; YMF = Young Marantaceae forest; MMF = Mixed
10 Marantaceae forest.*

11 **TABLE S3.** *Soil organic carbon concentration along the soil profile of all plots from each phase of
12 the chronosequence. Within each column, different letters indicate significant differences ($p < 0.001$)
13 only for the same phase of the natural succession, while no letters indicate no any significant
14 difference within the same layer of phase.*

15

16 **LITERATURE CITED**

17 ANGASSA, A., G. OBA. 2010. Effects of grazing pressure, age of enclosures and seasonality on bush
18 cover dynamics and vegetation composition in southern Ethiopia. *J. Arid. Environ.* 74: 111–
19 120.

20 ARAGAO L. E. O. C., B. POULTER, J. BARLOW, L. ANDERSON, Y. MAHLI, S. SAATCHI, O. L.
21 PHILIPS, E. GLOOR. 2014. Environmental change and the carbon balance of Amazonian
22 forests. *Biol. Rev.* 89: 913–931.

23 ARCHER, S., T. W. BOUTTON, K. A. HIBBARD. 2001. Trees in grasslands: Biogeochemical
24 consequences of woody plant expansion. In E. D. Schulze, S. P. Harrison, M. Heimann, E.
25 A. Holland, J. Lloyd, I. C. Prentice, D. Schimel (eds). *Global biogeochemical cycles in the
26 climate system*, pp 115–137. Academic Press, San Diego.

- 1 ARCHER, S. 1995. Tree–grass dynamics in a subtropical savannah: reconstructing the past and
2 predicting the future. *Ecoscience* 2: 83–99.
- 3 BINKLEY, D., C. GIARDINA. 1998. Who do tree species affect soils? The warp and woof of tree–soil
4 interactions. *Biogeochemistry* 1-2: 89–106.
- 5 BLAKE, G. R., KH HARTGE. 1986. Bulk density. In A. Klute (ed.). *Methods of soil analysis.*
6 *Physical and mineralogical methods*, pp. 363–375. SSSA Book, American Society of
7 Agronomy, Madison.
- 8 BOONE, R. D., D. F. GRIGAL DF, P SOLLINS, R. J. AHRENS, D. E. ARMSTRING. 1999. Soil
9 sampling, preparation, archiving, and quality control. In G. P. Robertson, D. C. Coleman, C.
10 S. Bledsoe, P. Sollins (eds). *Standard soil methods for long-term ecological research*, pp 3–
11 28. Oxford University Press, New York.
- 12 BROWN, S., A. E. LUGO. 1990. Tropical secondary forests. *J. Trop. Ecol.* 6: 1–32.
- 13 BUITENWERF, R., W. J. BOND, N. STEVENS, W. S. W. TROLLOPE. 2012. Increased tree densities in
14 South African savannas: 50 years of data suggests CO₂ as a driver. *Glob. Change Biol.* 18:
15 675–684.
- 16 COLEY, P. D., J. A. BARONE. 1996. Herbivory and plant defenses in tropical forests. *Annu. Rev.*
17 *Ecol. Syst.* 27: 305–335.
- 18 DON, A., J. SCHUMACHER, A. FREIBAUER. 2011. Impact of tropical land use change on soil organic
19 carbon stocks – a meta-analysis. *Glob. Change Biol.* 17: 1658–1670.
- 20 GLEIXNER, G., C. I. CZIMCZIK, C. KRAMER, B. LÜHKER, M. W. I. SCHMIDT. 2001. Plant
21 compounds and their turnover and stabilization as soil organic matter. In E. D. Schulze, M.
22 Heimann, S. Harrison, E. Holland, J. Lloyd, I. C. Prentice, D. Schimel (eds). *Global*
23 *Biogeochemical Cycles in the Climate System*, pp 201–215. Academic Press, San Diego,
24 CA.
- 25 GRACE, J., E. MITCHARD, E. GLOOR. 2014. Perturbations in the carbon budget of the tropics. *Glob*
26 *Change Biol.* 20: 3238–3255.

- 1 GUO, L. B., R.M. GIFFORD. 2002. Soil carbon stocks and land use change: a meta analyses. *Glob*
2 *Change Biol.* 8: 345-360.
- 3 HOFFMAN, W. A., R. B. JACKSON. 2000. Vegetation-climate feedbacks in the conversion of
4 tropical savanna to grassland. *J. Climate.* 13: 1593–1602.
- 5 IPCC. 2006. Guidelines for National Greenhouse Gas Inventories, Prepared by the National
6 Greenhouse Gas Inventories Programme. Eggleston HS, Buendia L, Miwa K, Ngara T,
7 Tanabe K. (eds). Published: IGES, Japan. Volume 4 AFOLU, Chapter 2 P. 2.29.
- 8 JACKSON, R. B., J. L. BANNER, E. G. JOBBAGY, W. T. POCKMAN, H. D. WALL. 2002. Ecosystem
9 carbon loss with woody plant invasion of grassland. *Nature* 418: 623–626.
- 10 MARÍN-SPIOTTA, E., C. W. SWANSTON, M. S. Torn, W. L. SILVER, S. D. BURTON. 2008. Chemical
11 and mineral control of soil carbon turnover in abandoned tropical pastures. *Geoderma* 143:
12 49–62.
- 13 MARTIN, D., Y. CHATELIN, J. COLLINET, E. GUICHARD, G. SALA. 1981. *Le sols du Gabon*, pp 1-
14 66. Orstom (ed), Paris.
- 15 MELILLO, J. M., MCGUIRE A. D., KICKLIGHTER D. W., MOORE III. B, VOROSMARTY C. I.,
16 SCHLOSS A. L. 1993. Global climate change and terrestrial primary production. *Nature* **362**:
17 234–240.
- 18 MIKUTTA, R., M. KLEBER, K. KAISER, R. JAHN. 2005. Review: Organic Matter Removal from
19 Soils using Hydrogen Peroxide, Sodium Hypochlorite, and Disodium Peroxodisulfate. *Soil*
20 *Sci. Soc. Am. J.* 6: 120–135.
- 21 MITCHARD, E. T. A., C. M. FLINTROP (2013) Woody encroachment and forest degradation in sub-
22 Saharan Africa's woodlands and savannas 1982–2006. *Philos. T. Roy. Soc. B.* 368:
23 20120406.
- 24 MITCHARD, E. T. A., S. S. SAATCHI, F. F. GERARD, S. L. LEWIS, P. MEIR. 2009. Measuring woody
25 encroachment along a forest–savanna boundary in Central Africa. *Earth Interaction* 13: 1–
26 29.

- 1 POWERS, J. S., M. D. CORRE, T. E. TWINE, E. VELDKAMP. 2011. Geographic bias of field
2 observations of soil carbon stocks with tropical land use change preclude spatial
3 extrapolation. *P. Natl. Acad. Sci. USA*. 108: 6318–6322.
- 4 RAICH, J.W., W. H. SCHLESINGER. 1992. The global carbon dioxide flux in soil respiration and its
5 relationship to vegetation and climate. *Tellus B* 44: 81–89.
- 6 SALDARRIAGA, J. C., D. C. WEST, M. L. THARP, C. UHL. 1988. Longterm chronosequence of
7 forest succession in the upper Rio Negro of Colombia and Venezuela. *J. Ecol.* 76: 939–958.
- 8 SAN JOSÈ, J. J., R. A. MONTES, M. R. FARIÑAS. 1998. Carbon stocks and fluxes in a temporal
9 scaling from a savanna to a semi-deciduous forest. *Forest Ecol. Manag.* 105: 251–262.
- 10 SCHOLES, R.J., S. R. ARCHER. 1997. Tree-grass interactions in savannas. *Annu. Rev. Ecol. Evol. S.*
11 28: 517-544.
- 12 SCHLÜTER, T. 2008. *Geological Atlas of Africa*. Springer-Verlag Berlin, pp110–113.
- 13 SCOTT, N. A., K. R. TATE, J. FORD-ROBERTSON, D. J. GILTRAP, C. T. SMITH. 1999. Soil carbon
14 storage in plantation forests and pastures: land-use change implications. *Tellus B* 51: 326–
15 335.
- 16 SOIL SURVEY STAFF. 2014. *Keys to soil taxonomy*, 12th edn. United States Department of
17 Agriculture, Natural Resources Conservation Service, Washington,, p. 360
- 18 SPICHIGER, R., C. PAMARD. 1973. Recherches sur le contact forêt-savane en Côte-d’Ivoire: Etude
19 du recrû forestier sur des parcelles cultivées en lisière d’un îlot forestier dans le sud du pays
20 baoulé. *Candollea* 28: 21–37.
- 21 WHITE, L. J. T. 2001. Forest-savanna dynamics and the origins of Marantaceae Forest in the Lopé
22 Reserve, Gabon. In B. Weber, L. J. T. White, A. Vedder (eds.). *African Rain Forest Ecology*
23 *and Conservation*, pp. 165–192. Yale University Press.
- 24 WILCOX, B. P. 2002. Shrub control and streamflow on rangelands: a process-based viewpoint. *J*
25 *Range Manag.* 55: 318–326.

1 VALENTINI, R., A. ARNETH, A. BOMBELLI, et 24 al. 2014. A full greenhouse gases budget of
2 Africa: synthesis, uncertainties, and vulnerabilities. *Biogeosciences*. 11: 381–407.

3

4 **TABLE 1.** *SOC concentrations in the mineral soil for the different phases of the natural*
5 *succession. Within each column, means followed by the same letters are not significantly different*
6 *(ANOVA; $p < 0.001$; $n = 25$ per layer).*

7

8 * S = Savannah; CF = colonizing forest; MF = Monodominant forest; YMF = Young Marantaceae forest; MMF =
9 Mixed Marantaceae forest.

Vegetation type *	0-5 cm	5-10 cm	10-20 cm	20-30 cm	30-50 cm	50-100 cm
	g/kg	g/kg	g/kg	g/kg	g/kg	g/kg
S	12.5±0.7a	9.6±1.6ab	7.5±1.4a	5.1±0.9a	2.8±0.2a	2.8±0.2a
CF	13.2±0.6ab	8.8±0.3a	7.9±0.6a	6.4±0.4a	4.6±0.3b	3.8±0.2b
MF	14.7±0.5b	8.5±0.5a	6.9±0.6a	5.7±0.3a	5.4±0.9b	4.2±0.4b
YMF	18.5±1.3c	10.3±0.4b	7.2±0.9a	5.4±0.4a	4.7±0.6b	4.5±0.4bc
MMF	24.8±0.2d	10.7±0.1b	7.5±0.1a	6.3±0.1a	5.6±0.1b	4.8±0.1c

10

11

12

13

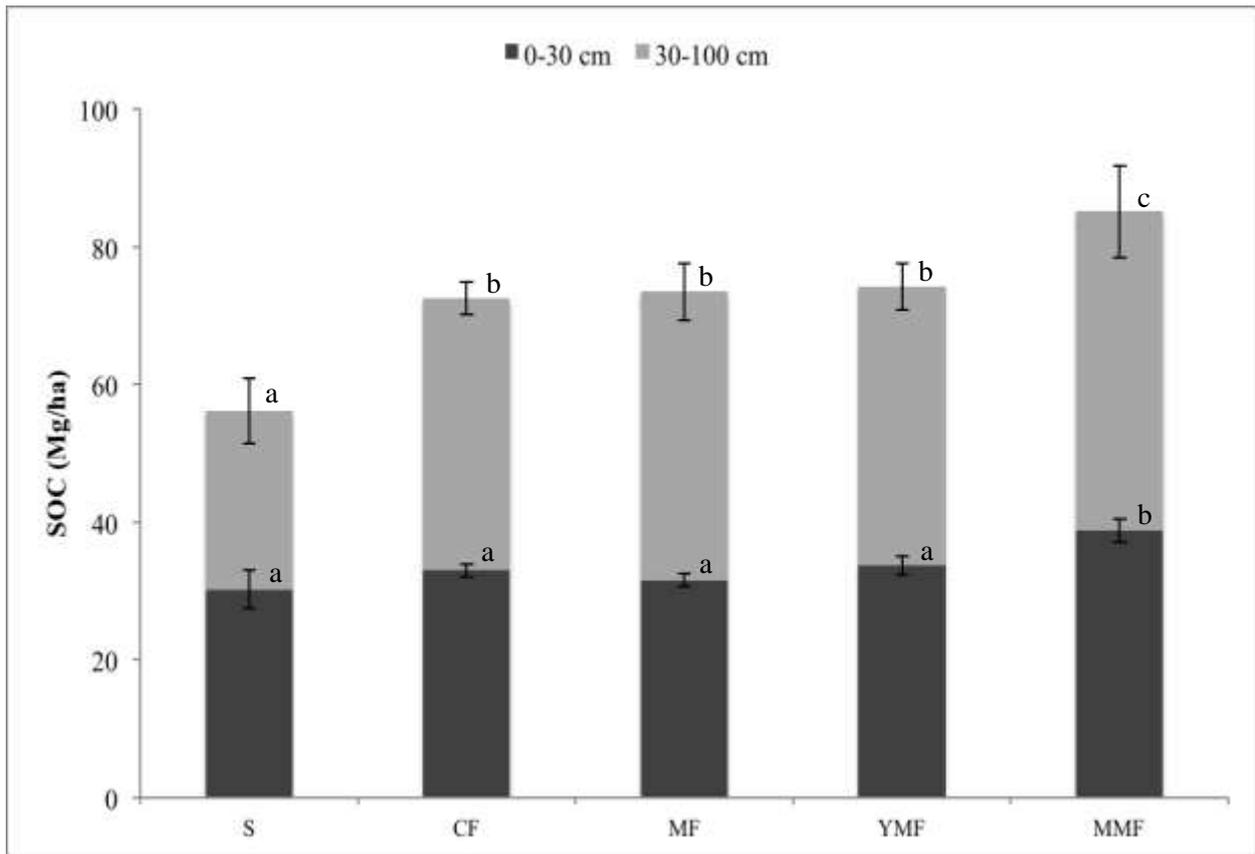
14

15

16

17

18



1

2 **FIGURE 1.** Mean SOC stock for the 0-30 cm and 30-100 cm depth of mineral soil under the
 3 different vegetation types across the natural succession. Different letters indicate significant
 4 differences (ANOVA; $p < 0.001$; $n = 25$ per vegetation type both in the topsoil and subsoil) only
 5 within the same compartment. S = Savannah; CF = colonizing forest; MF = Monodominant forest;
 6 YMF = Young Marantaceae forest; MMF = Mixed Marantaceae forest.

7

8

9

10