Abstract: This paper explores the development of a sustainable production system for tilapia and the research implications involved with ensuring commercial viability of such a system for UK farmers. The tilapia is a warm water fish with firm texture, white flesh and mild taste quite similar to a cod or haddock. Whilst tropical in origin it is thought to be highly suitable for low cost aquaculture in temperate zones with the potential to be a more sustainable source of food with fewer environmental impacts than other substitutes. Drawing on a literature review and findings from technical trials the paper will review and compare two production systems - novel Activated Suspension Technology (AST) and conventional Recirculating Aquaculture Systems (RAS) - considering their feasibility in terms of potential and financial viability for scaling up to commercial production of tilapia and their environmental and sustainability benefits. The review concludes that AST based only on microbial floc is currently uncompetitive with RAS in a UK context although the approach has benefits that might be incorporated in a new generation of mixed systems. Refinement of such systems needs to occur with potential adopters and could be part of diversification of mixed farms. Such development might further enhance the ethical values of fish produced in small-scale, modular RAS.
Table 1 Water and land productivity of tilapia RAS compared to selected intensive open-water production systems (after Phillips et al., 1991 reported in Timmons et al 2002).

<table>
<thead>
<tr>
<th>System</th>
<th>Species</th>
<th>Water productivity $\text{kg m}^{-3}$</th>
<th>Production intensity $\text{mt ha}^{-1} \text{yr}^{-1}$</th>
<th>Ratio of land and water use to RAS use</th>
</tr>
</thead>
<tbody>
<tr>
<td>RAS</td>
<td>Nile tilapia</td>
<td>10</td>
<td>1,340</td>
<td>1</td>
</tr>
<tr>
<td>Intensive ponds</td>
<td>Nile tilapia</td>
<td>0.05</td>
<td>17.4</td>
<td>200</td>
</tr>
<tr>
<td></td>
<td>Paneid shrimp</td>
<td>0.05 - 0.09</td>
<td>4.2-11</td>
<td>110-200, 120-320</td>
</tr>
<tr>
<td></td>
<td>Channel catfish</td>
<td>0.2-0.3</td>
<td>3</td>
<td>400-500, 446</td>
</tr>
<tr>
<td>Raceways</td>
<td>Rainbow trout</td>
<td>0.005</td>
<td>150</td>
<td>2,100, 9</td>
</tr>
</tbody>
</table>
Table 2. Comparison of production parameters in experimental RAS and AST grow-out systems for *O. niloticus* (source: Murray et al. 2007)

<table>
<thead>
<tr>
<th></th>
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<th></th>
<th></th>
</tr>
</thead>
<tbody>
<tr>
<td>Production system</td>
<td>RAS</td>
<td>AST</td>
<td>AST</td>
<td>AST</td>
</tr>
<tr>
<td>Indoor/ outdoor</td>
<td>indoor</td>
<td>indoor</td>
<td>outdoor</td>
<td>outdoor</td>
</tr>
<tr>
<td>Dietary crude protein %</td>
<td>30.4</td>
<td>32</td>
<td>32</td>
<td>20</td>
</tr>
<tr>
<td>Secondary carbohydrate source</td>
<td>na</td>
<td>none</td>
<td>none</td>
<td>cellulose</td>
</tr>
<tr>
<td>Solids management (TSS mg l⁻¹)</td>
<td>&lt; 2</td>
<td>250-1000</td>
<td>898 (100-1960)</td>
<td>no data</td>
</tr>
<tr>
<td>Culture unit</td>
<td>2.8 m³ tanks</td>
<td>1.5 m³ tanks</td>
<td>200 m³ tank</td>
<td>50m² earthen pond</td>
</tr>
<tr>
<td>Culture days</td>
<td>107</td>
<td>no data</td>
<td>201</td>
<td>30</td>
</tr>
<tr>
<td>Mean temperature (°C)</td>
<td>29.4</td>
<td>no data</td>
<td>28.5</td>
<td>-</td>
</tr>
<tr>
<td>Mean start weight (g)</td>
<td>19.1</td>
<td>41</td>
<td>73.6</td>
<td>112</td>
</tr>
<tr>
<td>Mean end weight (g)</td>
<td>405</td>
<td>134</td>
<td>678</td>
<td>218</td>
</tr>
<tr>
<td>Cumulative SGR (%)</td>
<td>2.8</td>
<td>1.27</td>
<td>1.11</td>
<td>1.31</td>
</tr>
<tr>
<td>Final biomass kg m⁻³⁻¹</td>
<td>25.6</td>
<td>9.8</td>
<td>13.7</td>
<td>16.5</td>
</tr>
<tr>
<td>Cumulative FCR</td>
<td>1.1</td>
<td>1.83</td>
<td>1.9</td>
<td>2.17</td>
</tr>
<tr>
<td>Cumulative PCR</td>
<td>3.1</td>
<td>no data</td>
<td>-</td>
<td>2.18</td>
</tr>
<tr>
<td>Survival</td>
<td>98.2</td>
<td>no data</td>
<td>81</td>
<td>94.8</td>
</tr>
<tr>
<td>Water productivity kg m⁻³⁻¹</td>
<td>8.1</td>
<td>no data</td>
<td>9.7</td>
<td>no data</td>
</tr>
</tbody>
</table>

*Extrapolated from start and end weights; Specific growth rate; Food Conversion ratio; Protein conversion ratio.
REVIEW: Options for producing a warm-water fish in the UK: Limits to “Green Growth”?

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This paper explores the development of a sustainable production system for tilapia and the research implications involved with ensuring commercial viability of such a system for UK farmers. The tilapia is a warm water fish with firm texture, white flesh and mild taste quite similar to a cod or haddock. Whilst tropical in origin it is thought to be highly suitable for low cost aquaculture in temperate zones with the potential to be a more sustainable source of food with fewer environmental impacts than other substitutes. Drawing on a literature review and findings from technical trials the paper will review and compare two production systems - novel Activated Suspension Technology (AST) and conventional Recirculating Aquaculture Systems (RAS) - considering their feasibility in terms of potential and financial viability for scaling up to commercial production of tilapia and their environmental and sustainability benefits.

The review concludes that AST based only on microbial floc is currently uncompetitive with RAS in a UK context although the approach has benefits that might be incorporated in a new generation of mixed systems. Refinement of such systems needs to occur with potential adopters and could be part of diversification of mixed farms. Such development might further enhance the ethical values of fish produced in small-scale, modular RAS.

Introduction

Seafood consumption in the UK is on the rise (Seafish, 2006), although in comparison with other countries the amounts consumed are relatively small. In recent years a high media profile has affected UK consumers’ interest in and perceptions of seafood. Some of the many issues making headlines range from the health benefits of including fish in the diet (Britton, 2006), to concerns with the safety of consuming both wild and farmed fish (Foran, Carpenter, Hamilton, Knuth & Schwager, 2005). Declining wild fish stocks (Worm et al., 2006) and the quality of the marine environment (Royal Commission on Environmental Pollution, 2004) are also frequently brought to the public eye, creating a complex picture for the public. The diversity of contradictory messages received by the public instigates confusion (Young, Grady, Little, Watterson & Murray, 2006).

Most of the fish used for human consumption currently comes from wild capture fisheries; however seafood from aquaculture is growing rapidly and is set to account for 50% of the worlds’ food fish in the near future (FAO, 2007a). The rapid growth in aquaculture production is attributed to declining wild stocks even as these continue to be exploited for use in feed for farmed carnivorous aquatic species as well as for other forms of intensive livestock production. This is a major cause of controversy (White, O’Neill & Tzankova, 2004)
The largely negative view of aquaculture in the UK, as a highly intensive, specialised and vertically integrated business model contrasts with traditional practice elsewhere around the world. In Asia where global aquaculture remains concentrated, low trophic species such as carps and tilapias still dominate farmed production and much of this is based on pond-based semi-intensive or extensive systems. This type of aquatic farming is characterised by the high proportion of feed being produced through natural food webs in situ (Azim & Little, 2006). Traditionally aquaculture was one component of mixed farming systems and geared to meet subsistence and local market needs (Beveridge & Little, 2002). But soaring demand and limitations of these systems has fuelled a major scale-up in the world wide production of farmed ‘seafood’ over the last two decades both to meet local and, increasingly, international markets. The shrimp boom in the mid-1980s-90s based on a limited number of species (mainly Penaeus spp.) and more latterly Nile tilapia (Oreochromis niloticus) and Asian river catfish (Pangasius hypothalmus) have both spread and intensified, particularly in developing countries where land, water and labour are abundant and cheap.

Tilapias have been heralded as a seafood commodity with major potential (Josupeit, 2005). In contrast to shrimp production the rapid scale-up in tilapia production has attracted little criticism from environmental groups and instead been portrayed as a white fish alternative to species higher up the food chain (Marine Conservation Society, 2006). They are being produced in a wide range of production systems and countries in the Tropics and Sub-tropics unlike the Asian river catfish where significant production is concentrated in one area—the Mekong Delta of Vietnam. It might be argued that these factors increase the relative opportunity for sustained growth of tilapia production, especially as despite the levels of growth, prices have remained relatively firm (Josupeit, 2005).

Global production of tilapias has soared over the last decade (FAO, 2007b) with particularly significant growth in South America for export markets and China for both internal and export markets (Josupeit, 2007a, b). This tropical species that originated in Africa is now the 6th most popular seafood choice in the USA (National Fisheries Institute, 2005) and major aquaculture producers turn to tilapia as a new species to invest in (Josupeit, 2007b). Although major centres of tilapia production are in Asia, South and Central America and Africa, culture has also become established in North America and Europe in the last few years. Tilapia production in the UK has been mainly characterised by high profile failures to date (Bunting & Little, 2005). This review assesses the technological options for tilapia production within insulated agricultural buildings proposed as a potential option for rural diversification (Little, 2006).

Towards greener aquaculture

In light of the contradictory messages conveyed in the media, consumer understanding of the ethical and human health issues surrounding aquaculture is understandably confused; however there is still a strong desire for fresh, traceable fish amongst UK consumers (Young et al, 2006). The natural shoaling behaviour of many fish species make the farming of fish at high density both practical and ethical (>100
kg m\(^{-3}\)) provided that nutritionally balanced diets can be cost effectively delivered and the quality of the water can be maintained (Ebeling, Timmons & Bisogni, 2006).

Most of the fish species raised intensively are top carnivores most dependent on high quality feeds conventionally based on fishmeal and oils derived from wild fisheries. These feed ingredients are subject to contamination with persistent organic compounds and their amplification through the food chain (Worm et al, 2006). The relative risk of consumption of such farmed fish compared to fish of wild origin and other food stuffs for different groups is the focus of increasing consumer and scientific interest (Foran, Good, Carpenter, Hamilton, Knuth & Schwager, 2005; Ellingsen & Aanonsen, 2006).

The environmental costs of feed and water supply to aquaculture are also becoming a major cause of criticism (Naylor et al., 2000); particularly for carnivorous species but intensification of low tropic species such as tilapias and carps is also utilising increasing amounts of such feeds. So called ‘flow through’ or ‘open’ intensive systems, in which there is little or no water re-use, can be highly polluting on receiving waters partly because the cost effective removal of dilute soluble nutrients is problematic. Open systems includes raceways and cages that produce most of the tilapias traded internationally (Coward & Little, 2001).

Rapid global growth in farming fish and shrimp has occurred in tandem with strong commercial and environmental incentives to reduce the costs of feed and the impact of effluents respectively. Any review of the short history of aquaculture will illustrate that intensification is based on increasing the density of stocked animals stimulating increased use of both water exchange (to maintain water quality) and higher protein feeds. High water exchange aggravates nutrient loss and restricts opportunities for recycling these expensive inputs; it has also been linked to poor biosecurity and spread of disease in shrimp culture. This has caused a paradigm shift in recent years towards use of lower protein feeds in low exchange, green water systems, initially in shrimp production (McIntosh, 2000), but increasingly for other species. This approach appears to be particularly attractive for systems based on low trophic species such as the tilapias.

Recirculating aquaculture systems (RAS) are increasingly common land based systems in temperate countries in which water is reused after removal of waste nutrients and heat may be cost effectively retained in the system. The mechanisms for removal of suspended and dissolved wastes to reduce solids and nitrogenous compounds hazardous to fish are key parameters of RAS. There is no requirement for continual discharge of effluents into the environment, as is the case with the majority of conventional flow-through aquaculture systems. The ‘price’ to pay is the external energy cost of moving water through an appropriate water treatment system, temperature control, provision of adequate dissolved oxygen and need for complete balanced nutrition. The commercial culture of tilapia in RAS is now established in North America and parts of Europe as specialised enterprises targeting high value markets. Such operations have been either based on integration with waste heat or as stand-alone enterprises (Melard & Philippart, 1981; Bunting & Little, 2005). This factor together with the fact that they can produce food locally with few effluents suggests they meet some of the criteria of ‘green’ food production systems. The recent history of limited RAS development in the UK suggests that production technologies
and markets are undeveloped; it is however established practice for value-added aquaculture such as accelerated production of juveniles for on-growing in open systems or ornamental production. Elsewhere in Europe they have become more established for catfish and eel production supplying diverse ethnic and cultural markets (Eding & Kamstra, 2002). More limited access to water for UK and European aquaculture and growing regulation on effluents is likely to increase the attraction of RAS. Rapid production cycles for warmwater fish are also attractive. Tilapias can reach marketable size in as little as six months while 18-24 months is the norm for UK farmed rainbow trout (*Oncorhynchus mykiss*) or Atlantic salmon (*Salmo salar*). The extent of potential water and land productivity gains for tilapia cultured in RAS compared with other intensive production systems for warm-water and temperate species are highlighted in Table 1.

**Why tilapia, why intensive??**

Tilapias have many characteristics amenable to farming, such as its fast growth under a range of conditions, resilience against disease and a flavour and texture comparable with valuable marine fish (Beveridge & McAndrew, 2000). An ability to feed low in the food chain in principle means that production costs can be low but also, importantly, the fish can be marketed to appeal to increasingly informed consumers on environmental and broader ethical grounds. The trends towards more intensive practices by most commercial tilapia producers threatens these potential core advantages but is a response to current commercial realities.

An ability to feed low in the food chain is matched by a high responsiveness to intensification such that tilapias perform well in intensive systems based on complete, but relatively low protein diets. Their tolerance of high densities (lower densities in fact often trigger aggressive territorial behaviour) has meant a rapid uptake of more intensive operations including more use of higher quality supplementary feeds in semi-intensive ponds (Edwards, Yakupitiyage & Lin, 2000) or complete, formulated feeds in intensive systems. Typically only 20-25% of fed protein is retained in the fish raised in intensive systems (Avnimelech, 2006) the balance becoming pollutants that must be removed. In principle if these waste nutrients could be retained in the system they become substrate for protein–rich bacteria that are re-ingested and utilised by the tilapia. Such nutrient recovery *in situ* occurs in conventional ponds but can be operated at a higher level of intensity through use of aeration to maintain microbial floc in suspension. These activated suspension ponds or technology (ASP, AST) have been advocated for both tilapias and shrimp (Avnimelech, Kochva & Diab, 1994; McIntosh, 2000). The nutritional value of such microbial floc to aquatic animals is dependent on several factors: food preference, ability to both ingest and digest it but also the density of the suspended particles (Hargreaves, 2006). Tilapias being both capable of filter feeding and detritivory are ideal candidates for such systems (Dempster, Baird & Beveridge, 1995; Azim, Verdegem, Mantingh, van Dam & Beveridge, 2003).

Potentially the relative operational simplicity of AST can be combined with a production intensity that is economically viable in the context of a diversification option for mixed farms in the UK. Moreover the ‘green’ characteristics of the approach could be favourable especially as the market for premium ethical food of all
types has developed rapidly but is under-supplied by local producers. The theoretical basis for the AST is now considered before its application in a UK context is assessed.

From feeding to floc

Intensification of aquaculture systems imposes two major technical challenges—the maintenance of dissolved oxygen and the removal of inorganic nitrogenous products. The latter is critical within intensive aquaculture systems as even low levels of unionized ammonia in water are toxic to most cultured species (Timmons, Ebeling, Wheaton, Summerfelt & Vinci, 2002). Oxygen levels typically become the limiting production factor in optimised culture-systems with adequate ammonia/nitrite treatment capacity.

There are three principle nitrogen pathways to remove hazardous N species in aquaculture (1) photo-autotrophic removal by algae, (2) immobilisation by heterotrophic bacteria as proteinacious microbial biomass and (3) chemo-autotrophic oxidation to nitrate by ‘nitrifying’ bacteria (Ebeling et al., 2006). The relative importance of each varies with system type and production intensity. Hargreaves (2006) distinguishes between ‘photosynthetic growth’ (PSG) and ‘mixed suspended growth’ systems (MSG) based on the degree to which water quality is maintained by photosynthetic and bacterial processes. Suspended particulates formed by heterotrophic bacteria also provide efficient substrates for nitrifying bacteria in bio-floc systems. These can be visualised as ‘green’ and ‘brown’ water systems.

Suspended-growth systems are further differentiated from ‘attached growth’ systems as the waste assimilation, recycling and food production occur within the culture unit as opposed to external bio-filters. Most aquaculture occurs in earthen ponds; which can be considered as PSG systems. Conversely, most RAS rely primarily on chemo-autotrophic bacteria attached as aerobic bio-films on filter media. These examples reflect opposing management goals; in attached systems the aim is to remove nitrogen from the system. In suspended systems the aim is to conserve and recycle nitrogen as useful microbial biomass. Suspended growth systems have also been referred to by a range of terms based on biological or containment characteristics: activated suspension ponds (ASP) activated suspension technology (AST), bio-floc technology (BFT), organic detrital algae soup (ODAS) etc.

Intensification of any suspended growth system requires oxygenation and good water mixing to increase the rate of ammonia immobilisation, both of which can be achieved simultaneously through vigorous aeration. Phytoplankton-rich systems will also benefit from in-situ oxygen generation, but with intensification they will ultimately become light limited through self shading. Thus sustained aeration and mixing are essential requirements for intensification of both green and brown water systems.

Although few cross-references exist in the literature these processes are also the basis of the ‘activated-sludge’ sewage treatment process (Ganczarczyk, 1983; Thiel, 2002). The main difference is that bio-floc accumulations in sewage treatment systems are periodically settled and voided in a continuous or semi-continuous process. In closed-AST the goal is to conserve bio-floc as a food source through internal nutrient
recycling. This mode of operation has two further beneficial features. Theoretically, water exchange rates can be reduced compared to conventional RAS, which are themselves conservative consumers of water (Table 1). Secondly, accumulation of waste inorganic nitrogen compounds; unionised ammonia and nitrite (NH$_3$ and NO$_2$) will result in growth inhibition or mortality of fish. *In-situ* heterotrophic ammonia and nitrite assimilation therefore also conserves water quality in this vital respect.

These attributes provided the impetus behind two major trends in the development and application of microbial bio-floc systems in aquaculture. The first has its origins in attempts to optimise natural feed production in semi-intensive ponds through various types of bio-manipulation. The second has its basis in the ‘zero-water exchange’ and water quality remediation possibilities of AST in contexts where water conservation is paramount. This driver had two threads. Researchers in Israel assessed AST as a potential means of simultaneously intensifying yields and water productivity in arid environments. Elsewhere, the same AST features, offered a means of addressing bio-security and environmental concerns associated with shrimp production. The development of intensive ‘zero exchange’ shrimp systems provided a highly effective means of disease and effluent management (Burford, Thompson, McIntosh, Bauman & Pearson, 2004, Hari, Kurup, Varghese, Schrama, & Verdegem, 2005; Lemonnier & Faninoz, 2006; Samocha et al., 2007) with feed optimisation as a secondary benefit (Burford, Thompson, McIntosh, Bauman & Pearson, 2003; Wasielosky, Atwood, Stokes & Browdy, 2006). The concurrent evolution of these two drivers is considered below.

The limits of natural productivity in ponds were initially explored using input: output work (e.g. Schroeder, 1978) based on the premise that light-limited primary productivity of conventional shallow ponds in the Tropics of 30kg ha$^{-1}$ d$^{-1}$ could be further enhanced by optimising heterotrophic productivity through addition of carbon rich substrates. Initially, this approach assumed that photo-autotrophic and heterotrophic feed pathways were partitioned and emphasised the role of heterotrophic pathways in achieving further yield gains. However, the interdependence of these pathways and the mechanisms by which fish such as tilapia could filter feed or harvest micro-organisms from the water column soon became apparent (Colman & Edwards, 1987; Avnimelech, Mokady & Schroeder, 1989).

Concurrent work carried out in Israel on more intensive systems suggested that sorghum and other energy-rich grains could be used cost effectively as supplements to natural food—especially micro-algae rather than more protein-rich feeds (Hepher, 1988). Yields in these intensive water-limited systems, were constrained by water quality limits stimulating further work aimed at enhancing AST function. Theoretically, optimising ratios of C:N will enhance conversion of toxic inorganic-nitrogen compounds to microbial biomass available as food for fish or shrimp while further improving water quality. Goldman et al. (1987) elucidated the fundamental nutrient balance principles underlying growth efficiency of marine bacteria. They found C:N ratios >10 :1 were optimal for optimising bio-floc production while minimising ammonia regeneration. Many investigators (Avnimelech et al, 1989, 1994, 1999, Hari et al., 2004, Burford et al, 2004) then applied this principle as an approach to optimising nutrient inputs and recycling within intensive bio-floc aquaculture systems. The use of a carbohydrate source in addition to conventional feeds or use of
feeds with lower protein content was advocated on this basis for systems in which bio-floc was aerated and retained in the system (Avnimelech, 1999).

The AST concept of further intensification of natural food production and use in situ has developed from this practice and theory for species such as tilapias and shrimp that are capable of utilising microbial floc as a major element of the diet and tolerating water high in suspended solids. Higher intensification rates also involve a move from earthen pond systems to lined-pond systems (shrimp) and tanks (tilapia). Most published accounts of AST however, relate to systems which maintain algal-rich water i.e. green water / PSG systems.

Generally green water systems are known to suffer inconsistent water quality, partly related to algal succession that is difficult to control or influence. Bacteria dominated systems tend to be more consistent (Hargreaves, 2006) but the nature and impacts of succession and change within systems with minimal phytoplankton are unknown. Our understanding of low-plankton systems is informed by the experience of managing partitioned aquaculture systems that alternate between autotrophic and heterotrophic status depending on ambient climate (Hargreaves, 2006). The principle of using compartments of algal rich water to remove ammonia is complicated by mixed success in controlling algal biomass.

The adaptation of these principles to a brown water / MSG system in light-limited conditions in which natural feed was mainly bacterial rather than derived from phytoplankton was the major objective of our research. The relative stability of heterotrophic microbial populations and their independence of light conditions on water quality were considered as positive factors (Avnimelech, 2006). For the sunlight-limited seasonal conditions in the UK the concept of well insulated smaller intensive tank-based systems located inside buildings was developed based on such a ‘brown-water’ approach. We now consider some of fundamental issues that differentiate AST as researched and promoted to date with their potential for use within the farming sector in the UK.

**Tilapia as a farm diversification strategy in the UK**

Intensive fish culture in the UK has been the preserve of an entrepreneurial business sector and the attraction of this type of diversification for risk-adverse farmers must be considered (Rosa, Kodithuwakku, Young & Little, 2007). Diversification into a novel product (i.e. tilapia) based on a new technical approach (AST) is likely to further increase risk. The potential benefits of using AST for tilapia production rather than RAS scaled down to meet the investment profiles and potential local market niches available to them need to be established.

The costs and risks of maintaining optimal temperatures for warm-water fish are an initial concern to most potential adopters. The optimal temperature range for tilapia production is 28-32°C, however, energy costs (heating and pumping) are proportionately low (15% total direct costs; Timmons, 2005). In the past, RAS have often been linked to waste heat utilisation from distilleries, power stations, factories etc. Whilst an apparently green and cost-effective approach, over-reliance on third-
party waste energy has also contributed to failures. A source of low value heat on-farm may be a motivation for diversification into warm-water fish culture. Another incentive is the utilisation of disused or underutilised agricultural buildings although low cost-purpose built structures such as insulated polytunnels also have potential.

Reducing the capital requirement and design complexity is an important advantage for any production system. In principle AST are simpler to design and manage than RAS; solids (feed and floc) are kept in suspension and dissolved oxygen levels maintained through aeration. As the culture unit also acts to treat wastes, there is no requirement for external biofilter, piping or pumps which results in lower capital costs and theoretically, more straightforward management. The capital outlay of these components for an RAS can range typically from 10- 35% of initial fixed costs. Low cost, simple AST could also be temporary or moveable structures allowing farmers to take advantage of seasonal availability of space, resources and marketing opportunities.

A potential incentive for producing tilapia using a microbial floc-based system rather than conventional RAS is the possibility that local feeds can be used. The overall reduction in feedstock quality required to raise tilapia in AST is potentially a substantial saving on production costs over RAS in which feed cost typically make up from 30-40% of total operating costs depending on the scale of the operation and other factors (Timmons et al., 2002, Timmons, 2005). Using a feed of lower overall quality feed i.e. 20% crude protein feed rather than typical formulations (28-32%CP) could reduce reliance on feed ingredients such as fish meal and soybean meals. Potentially it could open opportunities for growing or using feed ingredients locally or on-farm in a similar manner to that practiced for intensive dairy production thus reducing risk and enhancing familiarity that were important priorities for potential adopters (Rosa et al., 2007).

Over-ambitious production schedules, steep technical learning curves and lack of prior aquaculture experience have been inter-related causes of recurrent failure in RAS. Contract farming packages which emphasise potential gains while underestimating risk has contributed to spectacular failures in other novel farm diversification start-ups (e.g. ostrich, and Alpaca farming). Research indicates a similar threat in the UK tilapia sector. Small-scale modular approaches hold potential for limiting risks carried by new adopters with no previous aquaculture experience. These adopters then have the option of scaling up to more economically efficient units required to supply higher volume/ low margin commodity chains (food processors and supermarkets), or continuing to produce smaller volumes of fresh product for higher value niche markets. In the US, innovative tilapia production initially targeted value added markets but relatively high labour costs undermined their capacity to compete with imports leading them to target specialist live sales, often to ethnic minorities (Serfling, 2000). Significant scale-up in production of tilapia and other species such as Pangasius spp. in tropical countries threatens competitiveness of producers in the commodity sector in the UK.

A key research question is; can such a production approach be maintained at production levels that would be cost effective and attractive to farmers in the UK? The use of aquatic microbial floc as the basis for tilapia production has been advocated, but research on intensive indoor/ brown-water production systems is still
required to justify promotion of the AST approach to farmers in temperate climates such as the UK.

Comparing performance of AST and RAS

There is a recent history of research on the operation and efficiency of AST systems, most of which is based on intensively fed, green water systems in ponds or tanks (reviewed by Hargreaves, 2006). Most of the commercial application appears to relate to the relatively much lower-density shrimp production with relatively little published information regarding higher density fish production systems (Avnimelech, 2007). Unfortunately there is a dearth of data for replicated large-scale research systems and most conclusions have been drawn based on either short term small-scale experiments and/or observation of commercial or semi-commercial systems based on variable sized fish (Table 2). Only two trials (Rakocy, Bailey, Thoman & Shultz, 2002, Murray et al., 2007) report on-growing to the minimum harvest size of 400g feasible for markets in the UK. Most reports have emphasised the potential for improved feeding efficiency based on nutrient recycling in AST systems compared to RAS or conventional pellet-fed ponds (Avnimelech et al, 1989; Milstein, Avnimelech, Zoran & Joseph., 2001). Avnimelech (2006) for example cites feed: cost ratios in C:N manipulated pond-AST as being almost double control systems with higher crude dietary protein inclusion. However, meaningful evaluation of the commercial potential of AST compared to RAS also requires knowledge of fish growth rates and system carrying capacities. Unfortunately key parameters that would allow interpretation of growth are often lacking (e.g. water temperature) or inadequately presented. In particular crude daily weight gain rather than specific growth rates are routinely used for comparisons using fish of highly variable stocking and harvest weights.

Even allowing for these limitations the magnitude of difference between growth rates is evident. Under controlled temperatures, stocking and feed conditions with C:N manipulation and solids removal, Murray et al (2007) found growth rates in AST were only 68% of those achievable in RAS (achieving an SGR of 2.8 % for fish grown from 19g-405g; both systems fed on 30% CP diets). SGRs fell to 36% of RAS levels in AST fed on 18% CP diets. When one accounts for the slower growth rate of larger-fish, grow-out time to 400g is almost doubled in the fastest growing AST compared to the RAS control (Murray et al., 2007).

Low carrying capacities make the commercial case for intensive AST appear still more marginal. Stocking densities exceeding 100 kg fish m$^{-3}$ are routinely achievable in RAS with oxygenation and densities up to 70-80 kg fish m$^{-3}$ with aeration (Timmons et al., 2002). This compares to reported levels of only 10 - 16.5 kg fish m$^{-3}$ in AST (Table 2). Murray et al. (2007) achieved levels of 28 kg fish m$^{-3}$, but only using complete feeds and solids removal. Clearly the benefits of feed and water use efficiencies reported for AST need to be viewed in the context of growth inhibition and reduced carrying capacity in intensive systems. Both factors have consequences for overall production costs when capital and variable costs for building size, floor area, insulation, labour and heating etc. are considered. The same constraints also eliminate gains in water efficiency; Murray et al. (2007) and Rakocy et al. (2002) measured broadly comparable optimal rates of 7.2 and 9.7 kg m$^{-3}$ achievable in AST compared to typical RAS rates of 8-10 kg m$^{-3}$ (Tables 1 and 2).
The potential for further intensification appears to be fundamentally limited by biological factors which correlate with bio-floc concentration in closed systems. Hargreaves (2007) observed process instability at feeding levels above 200 g m\(^{-3}\) equivalent to a stocking density of 10 kg m\(^{-3}\) (at a feed rate of 2\% bw day\(^{-1}\)). Rakocy et al. (2002) and Murray et al. (2007) observed severe growth inhibition and increased mortality at TSS levels above 850 mg l\(^{-1}\). In practice therefore, there is a requirement for solids removal in AST to maintain a level of suspended solids which will not significantly retard food intake and growth or constrain economically viable stocking densities. This requires some form of external clarifier (Murray et al., 2007; Hargreaves, 2006; Rakocky et al., 2002). However, the variable quality (size, consistency and specific gravity) of microbial floc that occurs over time complicates the design and operational management of such clarifiers in indoor AST systems (Murray et al., 2007).

Operation of AST incorporating solids removal also represents a partial step-back towards RAS-type compartmentalisation with semi-continuous or continuous water re-circulation. Solids settled in external clarifiers could be removed or managed entirely independently for controlled release to the grow-out compartment. Investigators found floc composition varied in closed culture-systems with implications for chronic and acute event-mortalities (Murray et al., 2007; Azim & North, 2007; Rakoky et al, 2002). Compartmentalisation could also provide a means of floc-stabilisation potentially incorporating activated-sludge techniques borrowed from the water and sanitation sector where steady-state operation is a critical feature. One commercial producer in the United States has already moved along this route in a hybrid system; maintaining TSS within 70-130 mg l\(^{-1}\) and achieving net yields of 60 kg m\(^{-3}\) year\(^{-1}\) (Serfling, 2000); in other words, resorting to use of bio-floc primarily as a low-cost in-situ water treatment process with low water exchange requirements.

The fundamental theoretical benefit of AST; improved feed efficiency can also be challenged. Analysis of feed and crude protein conversion and retention indicate that the amounts of microbial floc in a brown water system utilised as feed over a range of commercial stocking densities in fish offered feeds of a range of quality and presentational form were minimal (Murray et al., 2007). This contrasts markedly with the values published by Avnimelech (1999) based on observations of light-driven AST systems but could reflect differences in interpretation of data. Attempts to manage microbial floc production by manipulating C:N ratio, or floc levels through solid removal are also highly variable in the systems described. Azim et al (2007) also reported increased feed conversion efficiency in AST compared to RAS systems in which fish were maintained at low densities and fed similar amounts of feed confirming the utilisation of microbial floc by fish.

There are other important characteristics of tilapia culture in AST that deserve mention. The specific conditions of AST appear to favour beneficial bacteria and reduce disease incidence compared to alternative systems. The absence of disease in AST systems has been related to the probiotic nature of microbial floc (Serfling, 2000; Murray et al 2007; Avnimelech and Bejarano, 2007).
The natural habitat of tilapias are turbid water lakes of Africa but the high levels of suspended solids that characterise AST has raised issues regarding the welfare and taste of the fish produced in such systems. The impacts of high levels of microbial floc in AST systems on the taste and welfare of tilapias has been recently assessed. Off-flavours are related to the absorption and accumulation of natural chemicals or compounds (such as geosmin and MBT) through the gills, skin or gastrointestinal tract of fish (Boyd & Tucker, 1998; Gautier et al., 2002). Contrary to common perception culture systems rich in natural food are not necessarily more likely to produce fish with off-flavour (Serfling, 2000; Eves, Turner, Yakupitiyage, Tongdee, & Ponza, 1995). Bue (2005) conducted organoleptic taste trials on fish raised in both RAS and AST and found no perceived differences among fish direct from tanks or after standard depuration techniques in fresh or saline water (Rungreungwudhikrai, 1995).

Generally high levels of suspended solids are related to poor fish welfare as indicated by poor growth, fusion of gill lamellae (Mettam, 2005) and susceptibility to bacterial or parasite infections (Noble & Summerfelt, 1996). Lower feed intakes and performance withstanding, Vincent (2006) found no indication of gill damage on fish raised over extended periods within AST or RAS systems nor differences in tail erosion, scale loss etc characteristic of poor welfare.

**Future research needs**

Assessment of the development process towards an intensive system for UK farmers to produce and market an exotic food fish species has identified a number of interesting issues. Prototype RAS systems now require testing with potential producers and this will require an iterative action learning approach whereby insights of the adopters are incorporated. Studies on the nature of entrepreneurship give some insight as to the characteristics of potential adopters and whether diversification was driven by need or opportunism (Rosa et al., 2007).

Clearly there are trade-offs in terms of environmental and broader ethical values of fish produced by RAS and AST. Both systems have very limited effluents which through virtue of their nutrient concentration are useful fertilisers (Watten, & Busch, 1984; McMurtry et al., 1997). Further research to quantify the potential synergisms between water and nutrient use in tank based systems and associated high value horticulture is required. Integration with hydroponics has particular market potential as demand for closed cycle, pesticide-free fruit and vegetables increases.

Although fish produced in AST systems had few overt signs of poor welfare, the lower feed intake, slower individual growth and chronic mortalities observed suggest that RAS provided more consistent and optimal conditions. Further development of mixed systems has been advocated in which culture units are partitioned with algae, microbial floc and/or periphyton (e.g. Avnimelech, 2006; 2007; Azim and Little, 2006; Serfling, 2000). Optimisation of floc levels for commercial applications is a research priority.

The value of microbial floc in terms of preventing fish disease problems warrant scientific investigation. Probiotic approaches are now widespread in the market but
the relative control possible in AST and observations of the high health of fish produced makes further investigation worthwhile. Designs in which the natural feed component can be optimised with respect to nutritional quality and energy efficient ingestion, digestion and assimilation should be prioritised. Development with producers in an action research mode is most likely to result in models which are management efficient and adoptable.

Intensive tilapia production is land efficient (Table 1) and may be located in periurban, rather than rural locations. Benefits include the improved access to a range of consumers, potentially reducing marketing costs. Controlled environments leading to improved predictability of production and expected genetic and feeding gains as has occurred in the broiler industry over the longer term are expected to further improve competitiveness compared with other fish species and substitutes (Timmons, 2005)

The market context for tilapia sales in the UK is dynamic. Consumers are increasingly willing to try new preparations and species of fish (Seafish, 2006b), whether it be for health reasons, indulgence or environmental grounds. Potential for tilapia therefore exists, not only in ethnic markets as a fresh or live alternative to frozen imports, but as a locally available ‘green’ fish product possibly with eco-credentials (Young, et al, 2006). Tilapia also has potential in the food service sector, where novel, exciting fish products are of interest, particularly if they have amenable aesthetic and preparation qualities (Seafish, 2006a). A locally available, small-scale and high quality tilapia supply would therefore meet industry wide interest in fresh, traceable fish supplies, however, a comparative analysis of the relative competitiveness of tilapias produced locally against imports and substitutes is required.

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